Application of ferric-carbon inner electrolysis coupling Fenton process pretreatment of coal tar distillation wastewater with a focus on optimization by the response surface method

Wenjing Lv^{a,b}, Jinying Li^{a,*}, Chunwei Yang^{a,b,*}, Qian Tang^{a,b}, Yiqun Zhang^c

^aKey Laboratory of Environmental Materials and Pollution Control, the Education Department of Jilin Province, Jilin Normal University, Siping, China Tel. +13694001402; email: lily_ying2003@163.com (J. Li); Tel. +86 04343291050; email: 739489998@qq.com (W. Lv); Tel. +86 13694001403; email: chunwei_yang@jlnu.edu.cn (C. Yang); Tel. +86 13844406327; email: tangqian327@163.com (Q. Tang) ^bCollege of Environmental Science and Engineering, Jilin Normal University, Siping, Jilin, China ^cSchool of Environmental and Safety Engineering, Qingdao University of Science and Technology, Qingdao, China, Tel. +86 15610015692; email: 12733167@qq.com (Y. Zhang)

Received 25 October 2019; Accepted 3 April 2020

ABSTRACT

Coal resources are abundant in China. The coal deep processing industry produces a large amount of coal tar. In China, the coal tar distillation process is important for obtaining chemical products such as naphthalene, anthracene, and phenanthrene. However, the wastewater generated from the coal tar distillation process is a typical degradation-resistant effluent, and the conventional sewage treatment process is not sufficient to treat it well. Alternative technologies must be developed. In this study, coal tar distillation wastewater was degraded utilizing the ferric-carbon inner electrolysis (FCIE) process following the Fenton process. The single-factor method and response surface method (RSM) were used to obtain the optimal conditions of the FCIE and Fenton processe, respectively. Under the optimal conditions, the total organic carbon (TOC) removal rates of the real coal tar distillation wastewater can reach 24.3% for the FCIE process, 65.1% for the Fenton process, and 73.6% in total. An analysis of variance was implemented, and the Box–Behnken RSM cubic model was obtained. This model shows less than 5% error between the calculated results and the experimental results. The TOC removal rate by the Fenton process followed the pseudo-zero-order kinetic model. This study could provide technical support for coal tar distillation wastewater pretreatment.

Keywords: Fenton; Wastewater; Coal tar distillation; Pretreatment

1. Introduction

Coal tar is the key chemical material in China. Chemicals such as naphthalene, anthracene, and phenanthrene can be obtained by treating the coal tar with the distillation process [1,2]. However, wastewater is also produced in the separation and purification unit of the coal tar treatment plant [3,4]. Coal tar distillation wastewater is a typical refractory organic wastewater that has the characteristics of high organic matter concentration, complex composition, high toxicity, and mutagenic and carcinogenic features [5,6]. It is challenging for engineers to treat coal tar distillation wastewater cheaply, effectively, and safely. Because traditional sewage treatment processes are not qualified for this type of refractory polluted water, an alternative treatment technology must be developed. The ferric-carbon inner electrolysis (FCIE) and Fenton combination process could be a feasible choice to

^{*} Corresponding authors.

^{1944-3994/1944-3986 © 2020} Desalination Publications. All rights reserved.

accomplish this treatment task. FCIE is a typical high concentration wastewater pretreatment process that is mostly used in the treatment of industrial wastewater [7,8]. After the FCIE process, the chemical oxygen demand (COD) of wastewater can be reduced by approximately 50%. Moreover, biodegradability can be increased to different levels [9]. However, the coal tar distillation wastewater concentration is too high to meet the effluent standard after undergoing the solo FCIE process treatment [10]. The Fenton process can also be used to pretreatment industrial wastewater [11,12]. The Fenton process degrades organic pollutant reactions, as illustrated by the hydroxyl radical theory [13,14]. The FCIE-Fenton combination process is a feasible choice for coal tar distillation wastewater pretreatment. In this study, real coal tar distillation wastewater was selected as the research object. FCIE factors, such as the mass ratio of iron and active carbon (Fe:AC), pH value, and the Fe concentration $[Fe]_{\scriptscriptstyle 0'}$ were studied, and the optimal conditions of FCIE were obtained. Furthermore, the response surface method (RSM) is utilized to analyze and optimize the Fenton process. The results of this study could supplement technical support for the coal tar distillation wastewater pretreatment.

2. Materials and methods

2.1. Chemicals

Coal tar distillation wastewater selected for this study was from a chemical plant located in Northeast of China. The pH value and total organic carbon (TOC) content of this wastewater were in the range of 7.5–8.8 and 14,000– 17,000 mg/L, respectively. Iron powder (analytical reagent grade, 99%) and active carbon (analytical reagent grade, 99%) were utilized in the FCIE process. H_2SO_4 (analytical reagent grade, 98%), and NaOH (analytical reagent grade, 99%) were utilized to adjust the pH value. FeSO₄·7H₂O (analytical reagent grade, 99%) and H_2O_2 (analytical reagent grade, 30%) were prepared to carry out the Fenton process. All chemicals were purchased from Sinopharm Chemical Reagent Co., China. All sample solutions were prepared with deionized water from an ion exchange system.

2.2. Degradation experiments

The FCIE process was studied in a 100 mL volume beaker. The coal tar distillation wastewater was adjusted to preset pH value, and then a certain amount of iron powder and active carbon was added, which followed a certain Fe:AC mass ratio. The reaction was maintained for 4 h with magnetic stirring (150 rpm). The samples were collected after treatment, and the TOC content was tested by a Vario TOC analyzer (Elementar Analysensysteme GmbH, Germany). The TOC removal rate, which was calculated following function (1), was a key factor to evaluate and optimize the FCIE process. The optimal FCIE conditions, such as Fe:AC, pH, and [Fe]₀, were obtained by the single factor studied method.

$$\eta = \frac{C_0 - C_t}{C_0} \times 100\%$$
(1)

where η is the TOC removal rate (%) at reaction time *t*, and *t* is the reaction time (min). *C*₀ is the initial TOC of wastewater (mg/L), and *C*, is the TOC at reaction time *t* (mg/L).

After the FCIE treatment process, the coal tar distillation wastewater was filtered and then moved to another beaker. The pH value was adjusted, and the preset amount of FeSO₄·7H₂O was added. The preset amount of H₂O₂ was injected five times over 2 h. Then, after 1 h of sedimentation, the TOC content was tested. In the kinetics study, the samples were collected at the Fenton reaction time of 10, 20, 30, 60, 90, and 120 min. These study data are the arithmetic means of two solo test results under the same conditions.

2.3. RSM factors and level design

The RSM is an effective tool to evaluate the reaction factors and obtain the optimal conditions by limiting the experimental results [15]. The TOC removal law can also be indicated by the fitting calculation with the Box–Behnken RSM model. However, the Box–Behnken RSM requires suitable factors and levels, which could have a large influence on the reaction. According to previous studies [16,17], the initial pH value, initial FeSO₄ concentration ([FeSO₄]₀), and total H_2O_2 concentration ([H₂O₂]) were selected because these factors can determine the Fenton reaction efficiency. The factors and levels of this study are shown in Table 1. The pH levels do not reach alkaline conditions because most reports indicated that the Fenton process is efficient under acidic conditions. The experimental list and the setting conditions are shown in Table 2.

3. Results and discussion

3.1. FCIE process

To illustrate the optimal conditions of the FCIE process, single factor experiment studies were implemented. The Fe:AC mass ratio, initial pH value, and $[Fe]_0$ concentration were considered, and the results are shown in Fig. 1. The Fe:AC mass ratio was the first factor to be obtained under the condition of 20 g/L $[Fe]_0$ and pH of 2.0. The 1:1 mass ratio of Fe:AC has better performance on the TOC removal rate at 22.5%. When the Fe:AC mass ratio changed to 1:2 and 2:1, the TOC removal rate only achieved 19.7% and 20.3%, respectively. Therefore, the optimal Fe:AC mass ratio was considered to be 1:1, which was consistent with previous research [10]. Next, the influence of pH value was tested under the condition of 20 g/L $[Fe]_0$ and a 1:1 mass ratio of Fe:AC. TOC removal rates reached 22.5%, 14.5%,

Table 1

Experimental factors and levels of Box–Behnken response surface method

Factors		Levels		
	Low	Medium	High	
A: pH	3.0	5.0	7.0	
<i>B</i> : $[FeSO_4]_0$ (g/L)	5.0	12.5	20.0	
C: [H ₂ O ₂] (30%, mL/100 mL)	2.0	3.5	5.0	



Fig. 1. Single factor study results of FCIE pre-treatment coal tar distillation wastewater.

16.5%, 22.2%, 8.1%, and 7.7% when the initial pH value was adjusted to 2.0, 3.0, 4.0, 5.0, 7.0, and 9.0, respectively. The results indicated that the FCIE reaction was hampered under alkaline conditions, which has also been reported previously [10,18]. An optimal initial pH of 5.0 was selected after a comprehensive consideration of the TOC removal rate and concentration for pH adjustment. Under conditions of pH 5.0 and a 1:1 mass ratio of Fe:AC, the influence of $[Fe]_0$ concentration was studied. The results indicated that an excess or a deficiency in $[Fe]_0$ concentration both had a negative impact on TOC removal. The optimal $[Fe]_0$ concentration was 40 g/L in this study, and the TOC removal rate reached 24.3%.

3.2. RSM results of the Fenton process

To investigate the interaction effects of the selected factors on the Fenton process, the Box-Behnken RSM was utilized in this study. The RSM experiment results are listed in Table 2. Based on these results, an analysis of variance (ANOVA) was carried out using Design-Expert software (V8.05b, Stat-Ease, Inc., USA), and the results are shown in Table 3. The cubic model was selected, and the ANOVA results were satisfactory. The fitted equation is shown in Eq. (2). The R^2 is 0.9960, and R^2_{adj} is 0.9872. The standard deviation (S.D.) is 1.35. The mean is 51.16, and the coefficient of variation (C.V.) is 2.63%. The F-value of 113.17 and *p*-value (Prob > F) of <0.0001 imply that this cubic model is significant and valid. There is only a 0.01% chance that this cubic model exhibits deviation due to noise. In this ANOVA, values of "Prob > F" less than 0.05 indicated that the model terms are significant. Therefore, the factors of B, C, BC, A², B², A²B, and A²C are significant and valid. Factor *B* (initial FeSO₄ concentration) has the largest "mean square" value, which indicated that B is the most important factor and has a significant influence on TOC removal in this study. The "lack of fit" Prob > F is 0.0851, and it is not significant relative to the pure error. This implies that there is only an 8.51% "lack of fit" chance occurring due to noise. The model is applicable to further discussion in this research.

$$\begin{split} \eta_{\text{TOC}} &= 232.21167 - 91.08333 \text{pH} - 12.12833 [\text{FeSO}_4]_0 + \\ &\quad 12.35667 [\text{H}_2\text{O}_2] + 6.28 \text{pH} \times [\text{FeSO}_4]_0 - 4.63333 \text{pH} \times \\ &\quad [\text{H}_2\text{O}_2] + 0.24889 [\text{FeSO}_4]_0 \times [\text{H}_2\text{O}_2] + 9.07833 \text{pH}^2 - \\ &\quad 0.070578 [\text{FeSO}_4]_0^2 - 0.36444 [\text{H}_2\text{O}_2]^2 - 0.63833 \text{ pH}^2 \times \\ &\quad [\text{FeSO}_4]_0 + 0.51667 \text{ pH}^2 \times [\text{H}_2\text{O}_2] \end{split}$$

where the η_{TOC} is the TOC removal rate, %; [FeSO₄]₀ is the initial FeSO₄ concentration, g/L; [H₂O₂] is the total H₂O₂ amount, 30% (mass concentration), mL/100 mL.

The three-dimensional response surfaces of this cubic model are shown in Fig. 2. The total H_2O_2 amount is a key factor in the Fenton process. From Figs. 2a and b, it can be found that a high H₂O₂ amount leads to a more significant efficiency in TOC removal. Under the condition of 5.0 mL/100 mL [H₂O₂], the TOC removal rate can reach approximately 70%. However, when the [H₂O₂] amount was reduced to 2.0 mL/100 mL, only approximately 55% of TOC removal rate could be obtained. It can also be observed from Figs. 2a and b that the shapes of the response surfaces are similar. When [FeSO₄]₀ maintains low concentrations from 5.00 to 14 g/L, the TOC removal rate is sensitive to the initial pH value. The acidic and neutral conditions had better TOC removal performances. However, when $[FeSO_4]_0$ is maintained at a high level (higher than 14 g/L), the influence of the initial pH value on TOC removal decreases. The $[FeSO_4]_0$ is also an important factor, as indicated by the ANOVA results. The shapes of the response surface under different [FeSO₄]₀ conditions present dramatic changes, and the results can be found in Figs. 2c and d. When $[FeSO_4]_0$ is 12.5 g/L, the response surface has a concave shape. A pH of 5.0 is a bad choice because the TOC removal rates maintain low values regardless of the $[H_2O_2]$ concentrations. However, when $[FeSO_4]_0$ is 20 g/L, the response surface changes to a convex shape. A pH of 5.0 becomes the optimal condition in this study. When the

Test	<i>A</i> : pH	B: [FeSO ₄] ₀ (g/L)	C: [H ₂ O ₂] (30%, mL/100mL)	TOC removal rate (%) calculate by model	TOC removal rate (%) experimental	Standard deviation
1	5.0	20.0	2.0	54.5	54.5	0.00
2	5.0	12.5	3.5	47.9	48.8	0.64
3	3.0	20.0	3.5	55.6	56.4	0.57
4	7.0	12.5	2.0	50.6	51.4	0.57
5	5.0	12.5	3.5	47.9	46.6	0.92
6	7.0	5.0	3.5	58.7	57.9	0.57
7	5.0	5.0	2.0	23.8	23.8	0.00
8	5.0	5.0	5.0	26.2	26.2	0.00
9	5.0	12.5	3.5	47.9	47.9	0.00
10	3.0	12.5	5.0	63.7	62.9	0.57
11	5.0	12.5	3.5	47.9	47.4	0.35
12	5.0	20.0	5.0	68.1	68.1	0.00
13	3.0	5.0	3.5	54.5	55.3	0.57
14	7.0	20.0	3.5	53.6	52.8	0.57
15	5.0	12.5	3.5	47.9	49	0.78
16	7.0	12.5	5.0	68.0	68.8	0.57
17	3.0	12.5	2.0	52.7	51.9	0.57

Table 2 Experimental design by the Box–Behnken method and the results for TOC removal by Fenton process

Table 3 ANOVA for TOC removal by response surface cubic model

Source	Sum of squares	Degree freedom	Mean square	<i>F</i> -value	<i>p</i> -value Prob > <i>F</i>	Significant
Model	2,258.73	11	205.34	113.17	< 0.0001	Significant
Α	2.42	1	2.42	1.33	0.3003	_
В	1,317.69	1	1,319.69	726.24	< 0.0001	Significant
С	64.00	1	64.00	35.27	0.0019	Significant
AB	9.61	1	9.61	5.30	0.0696	-
AC	10.24	1	10.24	5.64	0.0635	-
BC	31.36	1	31.36	17.28	0.0088	Significant
A^2	569.50	1	569.50	313.88	< 0.0001	Significant
B^2	66.36	1	66.36	36.58	0.0018	Significant
C^2	2.83	1	2.83	1.56	0.2669	-
A^2B	733.44	1	733.44	404.24	< 0.0001	Significant
A^2C	19.22	1	19.22	10.59	0.0226	Significant
Residual	9.07	5	1.81	-	-	-
Lack of fit	5.12	1	5.12	5.18	0.0851	Not significant*
Pure error	3.95	4	0.99	-	-	-
Correlation total	2,267.80	16	_	_	_	_

 $R^2 = 0.9960$, $R^2_{adi} = 0.9872$, S.D. = 1.35, Mean. = 51.16, C.V. = 2.63%.

*Lack of fit reflects the accuracy of an experiment. "Not significant" means the accuracy is high enough and does not affect the accuracy of the model.

initial pH is 7.0 and $[H_2O_2]$ is 5.0 mL/100 mL, Fig. 2e shows that the TOC removal rates are both higher than 60% under both conditions. The influence of the $[FeSO_4]_0$ condition is not obvious. Similar results can also be obtained at pH 3.0 (Fig. 2a). However, when the initial pH is 5.0, the influence of $[FeSO_4]_0$ on the TOC removal rate is remarkable (Fig.

2f). This could be because the lower $[FeSO_4]_0$ condition can cause the Fenton reaction to undergo competition with flocculation that is occurring. Under acidic conditions, the Fenton reaction occurs predominately, and under neutral conditions, flocculation becomes important. Therefore, both acidic and neutral conditions are acceptable. However,



Fig. 2. Estimated response surface of Fenton degradation coking wastewater: (a) actual factor $[H_2O_2]_0 = 5.0 \text{ mL}/100 \text{ mL}$; (b) actual factor $[H_2O_2]_0 = 2.0 \text{ mL}/100 \text{ mL}$; (c) actual factor $[FeSO_4]_0 = 12.5 \text{ g/L}$; (d) actual factor $[FeSO_4]_0 = 20.0 \text{ g/L}$; (e) actual factor pH = 7.0; (f) actual factor pH = 5.0, temperature: 20°C ± 2°C, reaction time: 120 min, magnetic stirring 150 r/min.

under the higher $[FeSO_4]_0$ condition, the $FeSO_4$ concentration is high enough to support both reactions. Furthermore, the high state hydrated ion of Fe(IV) could be generated, and the degradation process is more complicated and efficient. The optimal conditions obtained by this cubic model were an initial pH value of 5.14, 20 g/L $[FeSO_4]_0$ and 5.0 $[H_2O_2]$. The calculated TOC removal rate is 68.12%.

3.3. The degradation kinetics study

To evaluate the authenticity and reliability of the model, the degradation kinetics study was carried out under two selected conditions: initial pH of 5.0, 20 g/L [FeSO₄]_{or}

and 5.0 mL/100 mL $[H_2O_2]$ and initial pH of 3.4, 20 g/L $[FeSO_4]_{0'}$ and 5.0 mL/100 mL $[H_2O_2]$. The results are shown in Fig. 3. TOC removal rates increased steadily under both conditions. After 120 min of degradation, the TOC removal rate reached 65.1% and 61.6% for initial pH values of 5.0 and 3.4, respectively. Moreover, the calculated TOC removal rate values under these two conditions (68.1% for pH of 5.0, 64.8% for pH of 3.4) were also calculated and pointed out. The calculation and experimental result errors were maintained within 5%. The kinetics study was implemented with the Origin Pro 8 software (v8.0724), and the pseudo-zero-order, first-order, and second-order results are shown in Table 4. The pseudo-zero-order kinetic model is more reliable



Fig. 3. Actual degradation results of coking wastewater under optimal conditions were fitted by response surface method.

Table 4				
Kinetics study	results of BP	A degradation	by Fent	on process

Factors	Pseudo-zero order kinetic model equation: $C_t = C_0 - k_0 t$		Pseudo-first-order kinetic model equation: $\ln(C_0/C_t) = k_1 t$		Pseudo-second-order kinetic model equation: $1/C_t = k_2 t$	
	$k_0 (\text{mg/L/min})$	<i>R</i> ²⁺	k ₁ (1/min)	R^2	k_2 (L/mg/min)	R^2
pH = 3.4	73.134	0.99845	0.00812	0.94901	9.07×10^{-7}	0.88065
pH = 5.0	77.274	0.99534	0.00728	0.98093	7.79 × 10 ⁻⁷	0.98320

 C_t and C_0 is the TOC of BPA solution when reaction time is t and 0 min, mg/L; t is reaction time, min.

for the highest R^2 values of the model. The pseudo-zeroorder law was selected to simulate the degradation process because the organic pollutant concentration is high enough to react directly with the hydroxyl radical or Fe(IV) without degradation being disturbed. Therefore, the TOC removal process was not affected by the TOC concentration.

4. Conclusion

The FCIE-Fenton combination process was utilized to pretreat coal tar distillation wastewater in this research. Optimal conditions of the FCIE and Fenton processes were obtained by the single factor influence method and RSM. Under optimal conditions of pH of 5.0, a 1:1 mass ratio of Fe:AC, and 40 g/L [Fe], the TOC removal rate reached 24.3% under the FCIE process. The Fenton process was utilized to degrade the FCIE-treated coal tar distillation wastewater. The RSM and ANOVA results indicated that the cubic model can be used to predict the wastewater TOC removal rate. The initial FeSO₄ concentration is the key factor affecting the Fenton degradation process. The TOC removal rate can reach 65.1% under the optimal conditions of initial pH of 5.0, 20 g/L [FeSO₄]₀, and 5.0 mL/100 mL [H₂O₂], and the TOC removal rate could reach 68.1% by model calculations. The model calculations and experimental results errors were within 5% under the two different conditions.

In this study, TOC removal followed the pseudo-zero-order kinetic model. The total TOC removal rate of coal tar distillation wastewater can reach 73.6%. The results of this study could be supplemented with technical support for coal tar distillation wastewater pretreatment.

Acknowledgments

This work was supported by the National Natural Science Foundation of China, China (grant number 51708250), the Project of the Department of Science & Technology of Jilin Province, China (grant number 20180623042TC), and the Project of the Department of Education of Jilin Province, China (grant number JJKH20200424KJ and JJKH20200419KJ).

References

- H. Gai, L. Qiao, C. Zhong, X. Zhang, M. Xiao, H. Song, A solvent based separation method for phenolic compounds from lowtemperature coal tar, J. Cleaner Prod., 223 (2019) 1–11.
- [2] H. Wang, Z.-Y. Luo, M.-X. Fang, Q.-H. Wang, Controlled separation of coal tar based on different temperature, Fuel, 258 (2019) 115700–115703.
- [3] H. Gai, H. Song, M. Xiao, Y. Feng, Y. Wu, H. Zhou, B. Chen, Conceptual design of a modified phenol and ammonia recovery process for the treatment of coal gasification wastewater, Chem. Eng. J., 304 (2016) 621–628.

- [4] Y. Feng, H. Song, M. Xiao, K. Lin, K. Guo, H. Gai, Development of phenols recovery process from coal gasification wastewater with mesityl oxide as a novel extractant, J. Cleaner Prod., 166 (2017) 1314–1322.
- [5] W. Ma, Y. Han, C. Xu, H. Han, H. Zhu, K. Li, M. Zheng, Biotoxicity assessment and toxicity mechanism on coal gasification wastewater (CGW): a comparative analysis of effluent from different treatment processes, Sci. Total Environ., 637–638 (2018) 1–8.
- [6] H. Gai, X. Zhang, S. Chen, C. Wang, M. Xiao, T. Huang, J. Wang, H. Song, An improved tar-water separation process of low-rank coal conversion wastewater for increasing the tar yield and reducing the oil content in wastewater, Chem. Eng. J., 383 (2020) 123229.
- [7] X. Zheng, M. Jin, X. Zhou, W. Chen, D. Lu, Y. Zhang, X. Shao, Enhanced removal mechanism of iron carbon micro-electrolysis constructed wetland on C, N, and P in salty permitted effluent of wastewater treatment plant, Sci. Total Environ., 649 (2019) 21–30.
- [8] C. Ma, Z. Ran, Z. Yang, L. Wang, C. Wen, B. Zhao, H. Zhang, Efficient pretreatment of industrial estate wastewater for biodegradability enhancement using a micro-electrolysiscirculatory system, J. Environ. Manage., 250 (2019) 109492.
 [9] D. Wang, W. Ma, H. Han, K. Li, H. Xu, F. Fang, B. Hou,
- [9] D. Wang, W. Ma, H. Han, K. Li, H. Xu, F. Fang, B. Hou, S. Jia, Enhanced anaerobic degradation of Fischer–Tropsch wastewater by integrated UASB system with Fe–C microelectrolysis assisted, Chemosphere, 164 (2016) 14–24.
- [10] H. Lin, Y. Lin, L. Liu, Treatment of dinitrodiazophenol production wastewater by Fe/C and Fe/Cu internal electrolysis and the COD removal kinetics, J. Taiwan Inst. Chem. Eng., 58 (2016) 148–154.

- [11] Z. Wang, J. Li, W. Tan, X. Wu, H. Lin, H. Zhang, Removal of COD from landfill leachate by advanced Fenton process combined with electrolysis, Sep. Purif. Technol., 208 (2019) 3–11.
- [12] D. Gamaralalage, O. Sawai, T. Nunoura, Degradation behavior of palm oil mill effluent in Fenton oxidation, J. Hazard. Mater., 364 (2019) 791–799.
- [13] M. Zhao, M. Cheng, G. Zeng, C. Zhang, Degradation of di (2-ethylhexyl) phthalate in sediment by a surfactant-enhanced Fenton-like process, Chemosphere, 198 (2018) 327–333.
- [14] L. Zhao, Z.-R. Lin, X.-h. Ma, Y.-H. Dong, Catalytic activity of different iron oxides: insight from pollutant degradation and hydroxyl radical formation in heterogeneous Fenton-like systems, Chem. Eng. J., 352 (2018) 343–351.
- [15] C. Yang, X. Kong, L. Zhu, D. Wang, Application of electro-Fenton internal circulation batch reactor for methylene blue removal with a focus on optimization by response surface method, Desal. Water Treat., 132 (2018) 307–316.
- [16] P. Xu, H. Xu, Z. Shi, A novel bio-electro-Fenton process with FeVO₄/CF cathode on advanced treatment of coal gasification wastewater, Sep. Purif. Technol., 194 (2018) 457–461.
- [17] W. Yu, J. Yang, Y. Shi, J. Song, Y. Shi, J. Xiao, C. Li, X. Xu, S. He, S. Liang, X. Wu, J. Hu, Roles of iron species and pH optimization on sewage sludge conditioning with Fenton's reagent and lime, Water Res., 95 (2016) 124–133.
- [18] L. Zhang, Q. Yue, K. Yang, P. Zhao, B. Gao, Analysis of extracellular polymeric substances (EPS) and ciprofloxacindegrading microbial community in the combined Fe–C microelectrolysis-UBAF process for the elimination of high-level ciprofloxacin, Chemosphere, 193 (2018) 645–654.