A new porous $Ag_3PO_4/(Cs, Rb)_xWO_3/g-C_3N_4/CoAl-LDH$ composite towards efficient photocatalytic degradation of phenol and its derivatives

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ABSTRACT

In this work, a new porous $Ag_3PO_4/(Cs, Rb)_xWO_3/g-C_3N_4/CoAl-LDH$ (CoAl-LDH – CoAl-layered double hydroxide) composite was synthesized through a facile method and utilized as a photocatalyst towards efficient photocatalytic degradation of phenol and its derivatives, 2-chlorophenol (2-CP) and 2-nitrophenol (2-NP). A series of characterization and experiments results showed the resultant composite had excellent structure and outstanding performance, 2-NP was degraded most completely, followed by 2-CP and finally phenol. Furthermore, the composite still remained favorable stability after five cycles and the possible photocatalytic degradation mechanism was proposed. This paper provided a new idea for the preparation of novel photocatalysts.

Keywords: Porous composite; Layered double hydroxides; Photocatalytic degradation; Phenol; 2-chlorophenol; 2-nitrophenol

1. Introduction

Currently, largish interest has been invested to dispose of phenol and its derivatives, because of their high virulence and carcinogenicity in many industrial activities, such as cosmetics, disinfectants, herbicides, pesticides and artificial resins [1–3]. In 1976 eleven phenolic compounds were detailed in the United States Environmental Protection Agency prior pollutant list [4], among which the derivatives of phenol like chlorophenols and nitrophenols are deemed as a potential danger for their toxic nature. The massive discharge of these pollutants into the environment will lead to severe contamination to the aquatic system [5]. Furthermore, anthropogenic activities and unconscionable sewage disposal result in the more consecutive discharge of phenol and its derivatives to aquatic environments. Therefore, an easy and effective methodology that can simultaneously remove both chlorophenols and nitrophenols is important for water remediation. Currently, the photocatalytic degradation technology has been considered the most feasible route for decomposing phenol and its derivatives via utilizing versatile photocatalysts [6,7].

Among the reported hierarchical semiconductor photocatalyst, CoAl-layered double hydroxide (CoAl-LDH) is regarded as a promising photocatalyst, because it has a convenient cost, avirulence, appropriate oxidation-reduction electric potential and outstanding photochemical absorption performance [8–10]. However, the practical application of CoAl-LDH is hampered by its inherent defects, such as weak conductivity, low energy conversion efficiency and so forth [11]. In order to handle the above issues, multiple methodologies have been exploited to design CoAl-LDHbased structurally modulated heterostructures to enhance photocatalytic activities [12,13]. Recent literature provided a

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novel and facile strategy to boost the photocatalytic properties of LDH by coupling it with several hierarchical semiconductor materials, which drastically render high-efficiency physical contact and a strong electronic coupling [14,15], and g-C₃N₄ is a preferential candidate because of its similar laminar structure to LDH. Besides, the unique delocalized conjugated π structures of g-C₃N₄ can notably improve the charge separation efficiency, improving the stability and photocatalytic performance of LDH [16,17]. Regrettably, the heterostructure formed by LDH and g-C₃N₄ is still a problem because there is often the toilless formation of irreversible agglomerates, and the inferior stability of powdered photocatalysts seriously limit their application [18].

To dispose these questions and further obtain exceptionally effective charge transport, fabricating a heterojunction through collaborating with Ag-based photocatalysts is substantiated to be one feasible approach. Ag₃PO₄ is the optimal candidate for it admittedly possesses the highest quantum efficiency and presents distinctive photocatalytic activities toward hazardous organic pollutant degradation [19–21]. Cs_wWO₂ (a tungsten bronze type nanocrystal) is another preferential candidate for photocatalysis by right of its excellent optical absorption performance in a wide range including the UV, Vis and NIR region and distinctive conductivity. Beyond these, the existent mixed chemical valence of W6+ and W5+, oxygen vacancies and freeelectron also accelerate the immigration of photogenerated charges by forming localized states underneath the conduction band [22-25]. Recently, several researches manifested that the co-doped structure by introducing double cations into the M point of tungsten bronze can increase the free electrons in the conduction band, improve the lattice distortion degree and reduce the release of doped ions, thus significantly improving the selective light transmittance and structural stability of tungsten bronze [26-29]. Introducing Cs and Rb simultaneously to form co-doped tungsten bronze has not been studied. Except for the above measures, the creating of porous structures is a convenient and effective approach based on energy absorption to improve photocatalytic performance through speeding up the mass transfer and enhancing exciton dissociation efficiency, which will also significantly slow the toilless formation of irreversible agglomerates, increase the photoactive sites and the inaccessibility of internal surfaces of the synthetic samples for facile charge transfer, hence seriously improving the spring-board for fabricating and exploring new-fashioned and multifunctional photocatalyst, which shows an enticing prospect in photochemical reactions and energy conversion.

In this work, a novel porous $Ag_3PO_4/(Cs, Rb)_xWO_3/g-C_3N_4/CoAl-LDH$ composite was successfully synthesized. Ag_3PO_4, (Cs, Rb)_xWO_3 and g-C_3N_4 were introduced to improve the photocatalytic activity of CoAl-LDH. Phenol and its derivatives, 2-chlorophenol (2-CP) and 2-nitrophenol (2-NP) were regarded as model pollutants to evaluate the photocatalytic performance of prepared photocatalysts, and Ag_3PO_4/(Cs, Rb)_xWO_3/g-C_3N_4/CoAl-LDH exhibited favorable degradation efficiency. This work provides a novel strategy to enhance the catalytic activity of photocatalysts, which also provides a useful concept for the design of economical and efficient catalysts.

2. Experimental section

2.1. Materials

Cobalt nitrate hexahydrate (>98.5%, Sinopharm Chemical Reagent Co., Ltd., China), aluminum nitrate nonahydrate (>99%, Guangdong Jinhuada Chemical Reagent Co., Ltd., China), silver nitrate (>99.8%, Sinopharm Chemical Reagent Co., Ltd., China), rubidium chloride (AR degrade, Tianjin Guangfu Fine Chemical Research Institute, China), cesium hydroxide monohydrate (>99%, Adamas Reagent Co., Ltd., China), tungsten hexachloride (AR degrade, Alfa Aesar, China), tungsten hexachloride (AR degrade, Alfa Aesar, China), diammonium hydrogen phosphate (>99%, Sinopharm Chemical Reagent Co., Ltd., China), dicyandiamide (>99%, Sinopharm Chemical Reagent Co., Ltd., China), phenol (Alfa Aesar, China), 2-chlorophenol (Saen Technology Co., Ltd., China), 2-nitrophenol (Alfa Aesar, China).

2.2. Synthesis of Ag₃PO₄/(Cs, Rb)_xWO₃/g-C₃N₄/ CoAl-LDH (AWCL)

The CoAl-LDH, (Cs, Rb), WO₃/CoAl-LDH and Ag₃PO₄/ (Cs, Rb) WO₂/CoAl-LDH were synthesized in advance (supporting information). For AWCL composite (scheme 1), 0.2 g the resultant Ag₂PO₄/(Cs, Rb)₂WO₂/CoAl-LDH particles were ultrasonically dispersed in 50 mL saturated dicyandiamide solution. Pieces of soft cubic PU sponges (0.5 cm × 0.5 cm × 0.5 cm) were adequately impregnated into the above suspension with moderately stirring. After approximately 2 h adsorption, the polyurethane sponges (PU sponges) were placed at 80°C to evaporated the liquid completely. Repeatedly operating these process several times to make the PU sponges enrich the suspension adequately. The ultima $Ag_2PO_4/(Cs, Rb)_{v}$ WO₃/g-C₃N₄/CoAl-LDH was obtained through calcining the obtained PU sponges in a tube furnace in 550°C for 1 h (5°C min⁻¹). The hierarchically porous $g-C_3N_4$ was synthesized only using PU sponges and saturated dicyandiamide solution, and the bulk g-C₂N₄ was also synthesized through direct heating dicyandiamide at 550°C for 4 h as a contrast.

2.3. Characterization

The crystal structures of the samples were measured through X-ray powder diffraction (XRD, Bruker D8). The morphologies were examined via scanning electron microscopy (SEM, TESCAN MIRA3 LMU) equipped with the energy-dispersive X-ray spectroscopy analysis system. The chemical constitutions and valence band spectrum were investigated using X-ray photoelectron spectroscopy (XPS, ESCALAB 250Xi). The Fourier-transform infrared spectra were recorded on an AVATAR 360 spectrometer. The steadystate UV-vis absorption and UV-vis diffused reflectance spectra (UV-vis DRS) were obtained by a UV-vis spectrophotometer (Shimadzu 2401 model). The pore size distributions and Brunauer–Emmett–Teller (BET) were evaluated through nitrogen adsorption isotherms.

2.4. Photocatalytic activity tests

For typical photocatalytic runs, the photocatalytic activities of the obtained samples were investigated by photocatalytic removing varied contaminants under visible



Scheme 1. Schematic illustration of the synthesis process of AWCL composite.

light. In this experiment, phenol, 2-CP, 2-NP and a mixture of the above three pollutants were used as model pollutants, and a 500 W Xe lamp equipped with a full spectrum cut-off glass filter was employed as the simulated solar light source. Before light irradiation, the suspensions were moderately stirred in dark for 30 min to reach adsorption– desorption equilibrium ulteriorly. Thereafter turning on the 500 W Xe lamp, about 3 mL mixed solution was sampled and detected using a UV-vis spectrophotometer at a fixed interval. Simultaneously, a blank experiment under dark condition or no catalyst was carried out to explore the effects of absorption between photocatalysis and pollutant molecules causing by strong electrostatic attraction in accordance with the Langmuir adsorption.

3. Results and discussion

3.1. Characterization of the as-prepared materials

The phase identification of all the as-synthesized materials was analyzed by XRD measurement and presented in Fig. 1. The sharp characteristic peaks of g-C₂N₄ and Ag₂PO₄ rendered other peaks not prominent. The XRD patterns of (Cs, Rb), WO, exhibited the characteristic peaks of both Cs₂WO₂ and Rb₂WO₂ without any other impurities, indicating the facile hydrothermal process was feasible to prepare the (Cs, Rb), WO₃ without significant structural and composition changes [30,31]. The CoAl-LDH XRD patterns presented various diffraction peaks at 20 values of 11.3°, 23.4°, 34.7°, 39.2°, 46.6° and 60.1°, corresponding to (003), (006), (012), (015), (018) and (110) lattice planes, which indicated the CoAl-LDH belonged to the space group r3·m with a three-layer 3R polytypic rhombohedral symmetry and CO_3^{2-} was the intercalating anion [32,33]. After calculating according to the Bragg law, $2d \sin(\theta) = n\lambda$ [34], the interlamellar distance (d) in stacked CoAl-LDH was 0.76 nm, and the value of lattice parameter (a) was estimated to be about 0.32 nm. The $g-C_3N_4$ and $g-C_3N_3$ sponges exhibited similar XRD patterns with two typical



Fig. 1. XRD patterns of Ag_3PO_4 , (Cs, Rb)_xWO₃, CoAl-LDH, g-C₃N₄, g-C₃N₄ sponges and AWCL composite.

diffraction peaks at 13° and 27° corresponding to (002) and (100) lattice planes respectively [35], which turned out the retention of the π -conjugated system [36,37]. Nevertheless, the g-C₃N₄ sponges had relatively broadened peaks, which was ascribed to the crystallographic changes induced by process-temperature in original g-C₃N₄ [38–40]. For Ag₃PO₄, the presented diffraction peaks accorded with a bodycentered cubic phase (JCPDS NO.06–0505). Furthermore, the sharp and high-intensity diffraction peaks suggested the Ag₃PO₄ possessed high crystallinity. The AWCL composite presented a coexistence of Ag₃PO₄, (Cs, Rb)_xWO₃ and g-C₃N₄ sponges phases with no other impurities. The values of "a" and "c" on behalf of the unit cell directions of CoAl-LDH in AWCL were estimated to be 0.30 and 46 nm. The changes in "c" exposed Ag₃PO₄ and (Cs, Rb) $_x$ WO₃ successfully grew and distributed over the stacked exfoliated CoAl-LDH sheets with positively charged lamellar self-assembly structure and negatively charged g-C₃N₄ sponges.

The morphologies of the as-synthesized samples were investigated through SEM and the characterization results are shown in Fig. 2. The pristine CoAl-LDH displayed a prominent layer-by-layer structure, which was the typical characteristic of LDHs. The original g-C₃N₄ exhibited bulk morphologies without any obvious pores. Inversely the g-C3N4 disposed by PU sponges possessed distinct pores structure because the soft PU template would decompose and vaporize in the process of heating, the generated gases contributed to the formation of the pores. Simultaneously the shrinking of void stemming from the PU skeleton fibers' burning also promoted the formation of pore structures. The introduction of pores facilitated the penetration and absorption of visible light, accelerated the quality conversion in the pores of $g-C_3N_{4'}$ therefore enhanced the photocatalytic activities. The new AWCL composite was synthesized by introducing the above advantages. In Fig. 2d, the AWCL composite displayed excellent morphologies with favorable dispersibility, the Ag_3PO_4 and $(Cs, Rb)_vWO_3$ could be dispersed over the surface of CoAl-LDH irregularly and the introducing of $g-C_3N_4$ sponges further increased the dispersion as well as absorption of visible light. Besides, Fig. S1 also proved the formation of the AWCL composite.

The nitrogen adsorption–desorption technique was employed to calculate the specific surface area and pore size distribution. As shown in Fig. 3a, the $g-C_3N_4$, $g-C_3N_4$ sponges and AWCL composite all accorded with type IV isotherm with H3 hysteresis loops, and the pore-size distribution curve displayed there were abundant mesopores intermingling with few micropores. Apparently, the g-C₃N₄ sponges possessed larger pore size in contrast to the pristine g-C₃N₄, which meant affluent large channels have been introduced into the surface of g-C₃N₄, therefore accelerated visible light infiltration and absorption. Beyond the mesoporous channels similar to g-C₃N₄ sponges, the AWCL composite had a lager BET surface area (65.2 m² g⁻¹) and a larger pore volume (0.724 0.57 cm³ g⁻¹) than g-C₃N₄ and g-C₃N₄ sponges, which conformed with the SEM results (Table 1).

Regarding the enhancement mechanism of visible light absorption, the UV-visible diffuse reflection spectrum (UV-vis DRS) results further attested the AWCL composite achieved outstanding photoabsorption behaviors than other photocatalysts. As shown in Fig. 4, the (Cs, Rb), WO₃ had the best visible light absorption ability by virtue of its unique performance, which also assisted the AWCL composite to expend light regions. The g-C₂N₄ sponges obtained a wider absorption region of about 500 nm than $g-C_3N_4$, confirming the significance of pore architecture for light absorption. Besides, the micron pores and nanometer pores improved the photocatalytic performance in different aspects. The micron pores basically accelerated visible light into the nanometer pores followed into the photocatalyst, simultaneously the nano-pores mainly facilitated the light absorption on pore walls to expand the optical absorption regions [29]. Small bandgap energy (2.43 eV) render Ag₂PO₄ possesses excellent photocatalytic performance [17], thus the incorporation of Ag₃PO₄ notably



Fig. 2. SEM images of (a) CoAl-LDH, (b) $g-C_3N_4$, (c) $g-C_3N_4$ sponges and (d) AWCL composite.



Fig. 3. (a) N₂ adsorption–desorption isotherms and (b) pore size distribution of $g-C_3N_4$, $g-C_3N_4$ sponges and AWCL composite.

Table 1 Texture properties of $g\text{-}C_{_3}N_{_{4'}}$ $g\text{-}C_{_3}N_{_4}$ sponges and AWCL composite

Materials	BET area (m²/g)	Centred pore size distribution (nm)	Pore volume (cm ⁻³ g ⁻¹)
$g-C_3N_4$	10	5	0.247
g-C ₃ N ₄ sponges	60.4	10	0.676
AWCL composite	65.2	15	0.724

influenced the optical properties of AWCL composite. Furthermore, Fig. 4b shows the bandgap of the AWCL composite was smaller than $g-C_3N_4$ and $g-C_3N_4$ sponges, more firmly certifying the AWCL composite owned favorable photoabsorption activity.

The XPS technique was regarded as a rigorous tool to detect the prepared samples' elementary composition and

electronic environment. The XPS survey scan showed there were only Ag, Cs, Rb, Co, Al, P, O, W, C and N, elements existing in AWCL composite without other elements, conforming the AWCL composite was formed resoundingly (Supporting information Fig. S2). For Co 2p spectra, the two peaks centered at 783.3 and 796.7 eV were attributed to Co $2p_{\scriptscriptstyle 3/2}$ and Co $2p_{\scriptscriptstyle 1/2}.$ Correspondingly, two symmetrical peaks at 61.3 and 73.8 eV corresponded to the Al 2p spectrum [41]. The C 1s peaks revealed three evident peaks at 284.2, 286.2 and 288 eV for C-C containing defect in graphitic carbon, C-NH, and N=C-N respectively, and N 1s spectra presented three prominent peaks, which could be ascribed into C-N=C (397.5 eV), N-(C)₃ (398.2 eV) and -NH₂ (401 eV) [29,42]. As can be seen from Fig. 5c, the Ag 3d spectrum was divided into Ag⁰ along with Ag⁺ peaks. The former contained two main peaks at 368.2 and 374.2 eV, and the latter also possessed two peaks located at 367.8 and 373.8 eV [43,44]. For O 1s spectra, the peak at 530.6 eV was related to the lattice oxygen-O₂ in Ag_3PO_4 and another peak at 532.5 eV was connected with the absorbed water-O₄



Fig. 4. (a) UV-Vis diffuse reflectance spectra and (b) energy band gaps of $g-C_3N_4$, $g-C_3N_4$ sponges and AWCL composite.



Fig. 5. XPS spectra of AWCL composite for (a) Co 2p, (b) Al 2p, (c) Ag 3d, (d) C 1s, (e) N 1s, (f) O 1s and (g) P 2p.

presented on the surface of AWCL composite. The P 2p only had one symmetrical peak centered at 134.6 eV, which was put down to the P⁺⁵ came from PO_4^{3-} .

3.2. Photocatalytic degradation of phenol and its derivatives

3.2.1. Photocatalytic degradation of phenol

The photocatalytic degradation capability of phenol with AWCL composite was evaluated and compared with CoAl-LDH, (Cs, Rb)_xWO₃, g-C₃N₄, g-C₃N₄ sponges and Ag₃PO₄. In Fig. 6a, only 17.6%, 12.5%, 11.0%, 14.5%, 17% and 13.3% removing efficiency were obtained in dark, manifesting the adsorption of catalysts played a negligible role in oxidizing phenol. After exposure with a light source, an evident decrease was achieved for pristine Ag₃PO₄ (86.4%), (Cs, Rb)_xWO₃ (43.8%), g-C₃N₄ (58.8%), g-C₃N₄ sponges (71.3%) and CoAl-LDH (53.7%), and the AWCL composite presented the notably outstanding photocatalytic effect (94.6%), convincingly elucidating the mutual effect between Ag₃PO₄ (Cs, Rb)_x

 WO_3 , $g-C_3N_4$, $g-C_3N_4$, sponges, and CoAl-LDH occupied a significant place in photocatalytic degradation of phenol.

The photocatalytic degradation efficiencies of phenol in diverse pHs were also studied because the solution pH played a significant role in the existence of active substances. The solution pH was changed from 3 to 12, and the experiment results are shown in Fig. 6c. Obviously, the sequence of degradation efficiency was pH 7 > pH 3 > pH 5 > pH 9 > pH 12, the AWCL composite obtained the most prominent degradation efficiency at pH 7 and presented better removing effect in acidic solution than the basic solution. A strong base solution, the oxygen functional groups generated by AWCL composite were more liable to dissolve to make the structure of catalyst more unstable, thus tending to lose the photocatalytic activity and hamper the degradation efficiency.

To deeply research the reaction kinetic of degradation of phenol, the kinetics study of different catalysts was carried out and analyzed using a UV-visible spectrophotometer. The experimental results followed pseudo-first-order kinetic model, which can be expressed as below:



Fig. 6. Adsorption of phenol with different catalysts in the dark (a), photocatalytic degradation of phenol with different catalysts in visible light (b), the influence of pH on degradation of phenol with AWCL composite (c) and pseudo-first-order degradation kinetics for phenol with different catalysts (d).

$$\frac{C}{C_0} = e^{-kt} \tag{1}$$

where *k* is the reaction rate constant in min⁻¹, C_0 stands the original concentration and *C* represents the ultima concentration at instantaneous time *t*. The change curve of $\ln(C/C_0)$ over time can be synthesized into a straight line by linearly fitting the regression curve, and the corresponding rate constants *k*, coefficient of determinations (R^2), and standard errors are tabulated in Table 2. After being calculated, the AWCL composite had the maximum *k* of 0.02921 min⁻¹, which was 1.44, 3.04, 2.63, 1.87 and 4.05 times as higher as that of Ag₃PO₄ (0.02024 min⁻¹), (Cs, Rb)_xWO₃ (0.00962 min⁻¹), g-C₃N₄ (0.01107 min⁻¹), g-C₃N₄ sponges (0.01559 min⁻¹) and CoAl-LDH (0.00722 min⁻¹), respectively.

3.2.2. Photocatalytic degradation of 2-CP and 2-NP

The photocatalytic degradation capability towards 2-CP and 2-NP with $Ag_3PO_{4'}$ (Cs, Rb)_xWO_{3'}, g-C₃N_{4'} g-C₃N₄ sponges, CoAl-LDH and AWCL composite were also

Table 2 Fitted results of phenol degradation by different photocatalysts

Catalyst	K_{app} (min ⁻¹)	R^2	Standard error
CoAl-LDH	0.00722	0.95886	0.02069
$(Cs, Rb)_x WO_3$	0.00962	0.98562	0.02593
$g-C_3N_4$	0.01107	0.99645	0.15232
g-C ₃ N ₄ sponges	0.01559	0.98655	0.00362
Ag ₃ PO ₄	0.02024	0.98229	0.00532
AWCL composite	0.02921	0.99439	0.00152

investigated. As shown in Fig. 7, in the absence of irradiation (Supporting information Fig. S3.), all the reaction mediums with diverse catalysts presented a negligible amount of degradation towards 2-CP and 2-NP, which presented the sensitivity of catalysts to light and the stability of 2-CP and 2-NP. When exposed to light, the photocatalytic degradation performance dramatically enhanced, the removing capacity for 2-CP of disparate catalysts was put in the following order: the AWCL composite (97.5%) > Ag₃PO₄ (94.1%) > g-C₃N₄ sponges



Fig. 7. Photocatalytic degradation of 2-CP (a) and 2-NP (b) with different catalysts, the influence of pH on degradation of 2-CP and 2-NP with AWCL composite (c), and photocatalytic degradation of the mixed solution with AWCL composite (d).

 $(93.3\%) > g-C_3N_4$ (87.7%) > CoAl-LDH (79.8%) > (Cs, Rb)_x WO₃ (69.8%) > no catalyst (10.2%), and the degradation towards 2-NP was in a similar order: the AWCL composite (98.5%) > Ag₃PO₄ (97.5%) > g-C₃N₄ sponges (95.1%) > g-C₃N₄ (90.2%) > (Cs, Rb)_xWO₃ (84.7%) > CoAl-LDH (75.4%) > no catalyst (11.3%).

The influence of disparate pHs on degrading 2-CP and 2-NP was also researched and is presented in Fig. 7c. For 2-CP, there was a similar degradation efficiency trend to phenol, and the order was pH 7 > pH 5 > pH 3 > pH 9 > pH 12. The maximum degradation rate around 97% was reached at a neutral solution, and the tendency was gradually to fall at alkaline pH and to rise at acidic pH. Dissimilarly, as to 2-NP, the excellent photocatalytic efficiency were observed at basic solution (pH = 9) and neutral solution (pH = 7), and the order can be summarized as follows: pH 9 > pH 7 > pH 5 > pH 3 > pH 12.

To assess the kinetics of degrading 2-CP and 2-NP, the first-order model was employed to study the experimental and fitting results (Supporting information Fig. S4). The reaction rate constants (k), coefficient of determinations (R^2),

and standard errors obtained through linear fitting and then entered in the supporting information (Tables S1 and S2). In photocatalytic eliminating 2-CP, the *k* of the AWCL composite (0.03818 L/mg min) was 1.32-, 1.47-, 1.69-, 2.21- and 3.19fold to that of Ag₃PO₄ (0.02882 L/mg min), g-C₃N₄ sponges (0.02581 L/mg min), g-C₃N₄ (0.02246 L/mg min), CoAl-LDH (0.01731 L/mg min) and (Cs, Rb)_xWO₃ (0.01196 L/mg min). In photocatalytic exorcising 2-NP, the *k* of the AWCL composite (0.02595 L/mg min) was 1.31-, 1.57-, 2.21-, 2.61-, and 3.25fold to that of Ag₃PO₄ (0.01975 L/mg min), g-C₃N₄ sponges (0.01649 L/mg min), g-C₃N₄ (0.01172 L/mg min), (Cs, Rb)_x WO₃ (0.00998 L/mg min) and CoAl-LDH (0.00797 L/mg min). The above results presented an enhanced photocatalytic degradation capacity over the AWCL composite.

3.2.3. Comparative rates of photocatalysis by AWCL composite towards phenol, 2-CP and 2-NP

On the basis of the above experimental results, the most outstanding degradation efficiencies by AWCL composite over phenol, 2-CP and 2-NP were 94.6%, 97.5% and 98.5% respectively. In the mixed solution, the elapsed time eliminating mixed contaminants was longer relative to degrade individual phenol, 2-CP and 2-NP, and 2-NP was oxidized most thoroughly, according to the situation discussed above (Fig. 7d). The comparative rates of AWCL composite over phenol, 2-CP and 2-NP were concerned with their own structures.

Notoriously, phenols in solutions exist in the form of hydrogen ions and phenoxide anions, and the released hydrogen ions can delocalize the redundant negative charges generated on the aromatic ring to stabilize the structure of phenols. The -NO₂ groups in 2-NP and -Cl groups in 2-CP as electron-withdrawing groups can produce (-I) inductive effect to decrease the electron density on the benzene ring, thus reducing the reactivity of phenolic compounds. However, the two substituents are in the ortho position and can be used as electron donors to participate in π -electron conjugation on the benzene ring to increase electron cloud density, therefore reducing the stability of phenolic compounds to oxidants. For phenol, there are no substituents on the aromatic ring to produce an inductive effect and conjugate effect, hence the degradation efficiency over phenol is lower compared with 2-CP and 2-NP.

3.2.4. Photocatalytic mechanism of AWCL composite

For studying the primary reactive species participating in photocatalytic degradation of phenol 2-CP and 2-NP, the effect of multifarious scavengers was explored in irradiation conditions. During the process, ethylenediaminetetraacetic acid (EDTA), isopropyl alcohol (IPA), *p*-benzoquinone (BQ) and dimethyl sulfoxide (DMSO) were used as photogenerated holes (h⁺), hydroxyl radical (•OH), superoxide radical (•O₂⁻), and electrons (e⁻) scavengers, respectively. As shown in Fig. 8, the degradation efficiency of suspensions adding IPA, EDTA, P-BQ, and DMSO were 27.4%, 45.5%, 83.4% and 89.2%, affirm the primary active species was •OH and •O₂⁻ was the secondary active species. The same experiments were carried out with 2-NP and 2-CP, and similar results were obtained (Supporting information Fig. S5).



Fig. 8. Scavengers experiment for detecting active species using AWCL composite for phenol degradation.

According to the experiment data and relevant references, the possible degradation mechanism can be proposed as follows [45]. After visible light exposure, the AWCL composite would generate many e^-/h^+ pairs, then the e^- and h^+ would react with H_2O_2 , H_2O , OH^- and O_2 to form 'OH and ' O_2^- . Simultaneously, the ' O_2^- could be transformed into H_2O_2 , HO_2^+ , sequentially the H_2O_2 would collaborate with e^- to generate 'OH. Finally, the phenolic composites will be oxidized into the corresponding degraded products by ' O_2^- and 'OH.

$$AWCL + hv \rightarrow h_{vB}^{+} + e_{CB}^{-}$$
⁽²⁾

$$AWCL (h_{vB}^{+}) + H_2O \rightarrow {}^{\bullet}OH + H^{+}$$
(3)

$$AWCL(h_{\nu B}^{+}) + OH^{-} \rightarrow OH$$
(4)

$$AWCL\left(e_{CB}^{-}\right) + O_{2} \rightarrow O_{2}^{-}$$

$$\tag{5}$$

$$2^{\bullet}O_{2}^{-} + 2H^{+} \to H_{2}O_{2} + O_{2}$$
(6)

$$^{\prime}O_{2}^{-} + H^{+} \rightarrow HO_{2}^{\bullet}$$

$$\tag{7}$$

$$\mathrm{HO}_{2}^{\bullet} + 2\mathrm{H}^{+} \to \mathrm{H}_{2}\mathrm{O}_{2} + \mathrm{O}_{2} \tag{8}$$

$$AWCL (e_{CB}^{-}) + H_2O_2 \rightarrow OH + H_2O$$
(9)

 $O_2^- + OH + Phenolic composites \rightarrow Degraded products (10)$

3.2.5. Recyclability study of the AWCL composite

In general, the degradation efficiency of a catalyst will decline after interacting with specific contaminants for several cycles, which determined its recyclability must be tested before practical application. As found in Fig. 9, the AWCL composite still can maintain a degradation rate of over 80% after 5 cycles, indicating the prepared photocatalyst possessed



Fig. 9. Reusability study of AWCL composite for phenol, 2-CP and 2-NP degradation.

excellent stability. The lower degradation efficiency was possibly ascribed to that partial photocatalysts were covered by pollutants, reducing the surface active sites and specific surface area and therefore decreasing the activity of catalysts.

4. Conclusion

In summary, the novel porous $Ag_3PO_4/(Cs, Rb)_xWO_3/g-C_3N_4/CoAl-LDH$ composite was prepared successfully. The introduction of Ag_3PO_4 and $(Cs, Rb)_xWO_3$ effectively broadened the absorption region of CoAl-LDH, coupling with $g-C_3N_4$ rendered a high-efficiency physical contact and a strong electronic coupling. These measures endowed $Ag_3PO_4/(Cs, Rb)_xWO_3/g-C_3N_4/CoAl-LDH$ with superior photocatalytic performance over phenol, 2-CP and 2-NP, the degradation efficiency was 94.6%, 97.5% and 98.5% respectively. After five circulation runs, the AWCL composite still can obtain beyond 80% photodegradation rate, indicating the composite possessed favorable reusability and stability. This work opens a new horizon for the preparation of novel photocatalyst for eliminating multiple organic pollutants.

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Supplementary information

S1. Experimental section

S1.1. Synthesis of CoAl-LDH

In a typical procedure, CoAl-LDH was prepared by a simple hydrothermal method: 0.6 mol Co(NO₂), 6H₂O and 0.2 mol Al(NO₃)₃·H₂O were dissolved into 50 mL deionized water, another solution containing 0.2 mol Na₂CO₂ and 0.1 mol urea were drop-wisely added and subsequently, the mixed solution was translated into a Teflon-lined stainless autoclave to start a hydrothermal reaction at 120°C for 12 h. The final precipitates were obtained through centrifugation, wash and drying at 60°C under static condition overnight.

S1.2. Synthesis of (Cs, Rb), WO₃/CoAl-LDH

In a typical procedure, 0.2 g resultant CoAl-LDH and 0.1 mol WCl₂ were dissolved in 50 mL of anhydrous alcohol with continuous stirring for 20 min at normal temperature. Subsequently, 0.05 mol CsOH·H₂O, 0.05 mol RbCl and 20 mL acetic acid were added in the sequence listed with stirring and then the mixed solution was put into a Teflon-lined autoclave and afterward by heat-treatment at 200°C for 24 h. After cooling down, the resultant precipitates were collected by centrifuging, washing and drying at 60°C overnight. The pristine (Cs, Rb), WO, was synthesized following a similar method without CoAl-LDH.

S1.3. Synthesis of Ag₃PO₄/(Cs, Rb), WO₃/CoAl-LDH

In the general procedure, 0.2 g (Cs, Rb), WO₂/CoAl-LDH was dispersed in 100 mL ethanol to ultrasonic dispersion for 1 h. Afterward, a given mass of AgNO₂ was added into the dispersion with vigorous stirring for 1 h at dark and subsequently 0.1 mol (NH4), HPO4 was dropwise put into for another 2 h. Finally, the whole precipitates were collected by washing, extraction filtration and drying. The pristine Ag₃PO was prepared in the same way without adding (Cs, Rb), WO₃/ CoAl-LDH.

S2. Results and discussion



Fig. S1. Energy-dispersive X-ray spectroscopy image of AWCL composite.



Fig. S2. X-ray photoelectron spectroscopy pattern of AWCL composite.



Fig. S3. Photocatalytic degradation of 2-chlorophenol (2-CP) (a) and 2-nitrophenol (2-NP) and (b) with different catalysts in dark.



Fig. S4. Pseudo-first-order degradation kinetics for 2-CP (a) and 2-NP (b) with different catalysts.



Fig. S5. Scavengers experiment for detecting active species using AWCL composite for 2-CP (a) and 2-NP (b) degradation.

Table S1 Fitted results of 2-CP degradation by different photocatalysts

Table S2 Fitted results of 2-NP degradation by different photocatalysts

Catalyst	K_{app} (min ⁻¹)	R^2	Standard error	Catalyst	$K_{\rm app}$ (min ⁻¹)	R ²	Standard error
$(Cs, Rb)_x WO_3$	0.01196	0.98638	0.00324	CoAl-LDH	0.00797	0.89082	0.00422
CoAl-LDH	0.01731	0.99569	0.00398	$(Cs, Rb)_r WO_3$	0.00998	0.94927	0.00434
$g-C_3N_4$	0.02246	0.99292	0.00527	g-C ₃ N ₄	0.01172	0.94902	0.00446
g-C ₃ N ₄ sponges	0.02581	0.99380	0.00585	$g-C_3N_4$ sponges	0.01649	0.96804	0.00114
Ag ₃ PO ₄	0.02882	0.99362	0.00559	Ag ₃ PO ₄	0.01975	0.95785	0.00165
AWCL composite	0.03818	0.99591	0.00721	AWCL composite	0.02595	0.94123	0.00395