



Experimental studies on the removal of heavy metal ion concentration using sugarcane bagasse in batch adsorption process

V. Yogeshwaran^{a,*}, A.K. Priya^b

^aDepartment of Civil Engineering, Sri Krishna College of Engineering and Technology, Coimbatore – 641008, India, email: svyogi23190@gmail.com

^bDepartment of Civil Engineering, KPR Institute of Engineering and Technology, Coimbatore – 641027, India, email: akpriya@gmail.com

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ABSTRACT

Removal of toxic heavy metal ion concentration (Cr, Pb, and Zn) from the aqueous solution has been investigated using sugarcane bagasse as an adsorbent. The various properties of sugarcane bagasse powder were analyzed and the Fourier transmission infrared spectra, scanning electron microscopy, and energy dispersive X-ray analysis of sugarcane bagasse powder, before and after adsorption of Cr(VI), Pb(II), and Zn(II) were also examined. Test results show that the maximum adsorption efficiency of 95.65% for Cr(VI), 87.26% for Pb(II), and 83.32% for Zn(II) were attained at the pH of 6.0, the temperature of 30°C, the contact time of 60 min, the adsorbent dosage of 2.5 g/L with 25 mg/L of metal ion concentrations. The Langmuir and Freundlich isotherm models were adopted for the batch adsorption experimental work and the results prove that the pseudo-second-order model was fitted well with kinetic data. Also, the thermodynamic study confirms that adsorption mechanism is endothermic in nature with the best possible correlations.

Keywords: Heavy metals; Adsorption; Fourier transmission infrared; Scanning electron microscopy; Energy dispersive X-ray; Thermodynamic studies; Sugarcane bagasse powder

1. Introduction

Water pollution is one of the serious issues that we are facing for many eras. Fresh clean water is the essence for major industrial activities, communities, and all living organisms. The supply of clean water without any pollutant is one of the critical challenges and many countries are doing lots of research works to find a viable solution to this problem by various techniques [1]. The water gets highly polluted in recent times due to extreme activities of industrial manufacturing and other pathogenic activities [2]. This makes the water unsuitable for drinking by changing its physical and chemical properties [3]. Pollution in water may be created by the presence of dyes, metal ions, suspended and dissolved solids, and other organic and inorganic pollutants at very high concentration levels.

Among various pollutants in water, heavy metal pollution is one of the serious issues due to metal ion toxicity and accumulation and poses to be very dangerous to the surrounding environment and human beings [4]. On account of the increase in heavy metal pollution day by day, the present world faces many health issues such as cancer, respiratory problems, and other health issues (Table 1). Hence, it is necessary to reduce/remove the accumulation of heavy metal ions present in wastewater before discharging them into the water bodies. Many research works have been conducted to remove the accumulation of heavy metal ions from the wastewater [5]. The batch adsorption process was developed as a promising and innovative solution for removing metal ion concentration [6]. This process has many advantages such as low capital cost, selective metal removal, desorption with no sludge generation [7]. The adsorption process, which is an accumulation of

* Corresponding author.

atoms, ions, or gaseous molecules to the adsorbate surface by batch mode or fixed-bed column type has been long used for removal/reduction of heavy metal ions [8]. Using various adsorbate materials (both natural and industrial wastes), toxic metal ions have been removed by the adsorption process. Normally, the adsorbent material is converted into the activated carbon category for increasing the efficiency of the adsorption process [9]. Many natural adsorbents such as rice husk, neem leaves, sawdust, coconut shells rice bran, etc., were used for removing different metal ion concentrations from the aqueous medium [10].

Sugarcane bagasse powder is a type of organic material that has been used for removing heavy metal ion contamination from aqueous solutions in this study. These are the fractions of biomass which is produced after the extraction of sugarcane juice. Sugarcane bagasse is the heterogeneous fiber residue and around 30–34 tons can be produced from 100 tons of sugarcane in the factory process and these fibers are renewable and are used to produce biofuel and enzymes [11]. The decomposition period of sugarcane bagasse is around 45–60 d and these are eco-friendly materials that are produced around 80 MMT (million metric tonnes) annually in India [12]. Handling and disposal of sugarcane bagasse are one of the huge tasks and it consumes more time and money. Hence it is decided to use the sugarcane bagasse powder in water treatment, particularly for heavy metal ion removal. Many research works have been conducted to remove the heavy metal ions present in the aqueous solutions using sugarcane bagasse. From Table 2, it was observed that the sugarcane bagasse has the ability to remove various heavy metal ions from an aqueous environment. In this experimental study, the adsorption efficiency of heavy metal ions (Cr, Pb, and Zn) has been investigated using sugarcane bagasse powder in activated carbon form.

2. Materials and methods

2.1. Preparation of adsorbent

Sugarcane bagasse was collected and washed with de-ionized water and then dehydrated for 24 h at 60°C to eliminate tannins and other solvent organic complexes. The stock was cooled, purified, and washed several times with distilled water and eventually dried to 60°C. To boost stability, the bagasse was treated by 1% formaldehyde at a ratio of 1:4 (sugarcane bagasse: formaldehyde, weight/volume). The water is boiled for 6 h at 100°C and dried in an oven at 105°C for about 24 h. Table 3 represents the physical properties of sugarcane bagasse powder.

2.2. Pore distribution and Brunauer–Emmett–Teller surface area

The pore diameter, Brunauer–Emmett–Teller (BET) surface area, volume of micropore, and mesopore of SBP were obtained using adsorption–desorption isotherm process by nitrogen at –196°C temperature. Fig. 1 shows the nitrogen adsorption–desorption rate of SBP adsorbent and their characteristics are presented in Table 4. From Fig. 1, both the isotherms were seen to follow type II; which means that the SBP contains micropores and mesopores. The micropores are represented in the first part of the isotherm plot and mesopores are represented in the second part at high relative pressure, due to the adsorption of multilayers [17]. The BET surface area of SBP was about 558 m²/g which is smaller than other commercial activated carbons with a pore volume of 0.275 cm³/g. The BET surface area of 0.5 m²/g with a total pore volume of 2.70 × 10⁻³ cm³/g of white yam has been reported by Li et al. [18] for Cd²⁺ ion adsorption. Another study has been reported by Jing et al. [19] for activated

Table 1
Trace elements that may pose health hazards

| S. No | Element | Atomic No. | Sources | Health effects |
|-------|-----------|------------|---|--|
| (1) | Vanadium | 23 | Petroleum, chemicals and catalysts, steel, and other alloys | Probably no hazard at current levels |
| (2) | Chromium | 24 | Metal electroplating | A micronutrient to plant, chromates carcinogenic |
| (3) | Manganese | 25 | Mining and industrial wastes | Relatively non-toxic, and micro nutrient |
| (4) | Nickel | 28 | Fuel oil, coal, tobacco smoke, chemicals and catalysts, steel, and non-ferrous alloys | Lung cancer |
| (5) | Copper | 29 | Waste pipes, algae control, and industrial smoke | Possible liver damage with prolonged exposure, toxic to plants |
| (6) | Zinc | 30 | Metal electroplating, mining, and industrial smoke | Possible lung effects, low toxicity in solution |
| (7) | Arsenic | 33 | Coal, petroleum, detergents, pesticides | Arsenic poisoning |
| (8) | Selenium | 34 | Coal, sulfur | May cause dental caries, carcinogenic to rats, essential to mammals in low doses |
| (9) | Cadmium | 48 | Coal, Zinc mining, water mains and pipes, tobacco smoke | Cardiovascular disease and hypertension in human suspects interferes with zinc and copper metabolism |
| (10) | Lead | 82 | Auto exhaust, paints, and water through the lead joint of pipes | Brain damage, convulsions, behavioral disorder, and fatality |

Table 2
List of research work conducted for metal ion removal using sugarcane bagasse

| S. No. | Type of heavy metal ion | Adsorption efficiency | Concentration of heavy metal ions (mg/L) | Optimum pH | Adsorbent dose | Contact time (min) | Reference |
|--------|--|----------------------------|--|------------|----------------|--------------------|---------------|
| (1) | Cr ⁶⁺ , Pb ²⁺ , and Zn ²⁺ | 95.65%, 87.26%, and 83.32% | 25 | 6.0 | 2.5 g/L | 60 | In this study |
| (2) | Fe ²⁺ | 95.11% | 10.87 | 6.0 | 0.2 g | 100 | [13] |
| (3) | Cu ²⁺ , Ni ²⁺ , and Pb ²⁺ | 66.4%, 90%, and 99.9% | 25 | 4.0 | 7.5 g/L | 70 | [14] |
| (4) | Cu ²⁺ , Ni ²⁺ , and Pb ²⁺ | 268, 700, and 320 mg/g | 100 | 3.0 | 5 g/L | 180 | [15] |
| (5) | Cr ⁶⁺ and Zn ²⁺ | 98.84% and 94.67% | 30 | 4.0 | 1.5 g | 30 | [11] |
| (6) | Cu ²⁺ | 94.4% | 5 | 5.0 | 2 g | 120 | [16] |
| (7) | Cd ²⁺ | 90.9 mg/g | 200 | 6.5 | 1.5 g | 60 | [5] |

Table 3
Physical properties of treated sugarcane bagasse powder

| S. No | Physical parameters | Sugarcane bagasse powder |
|-------|----------------------------------|--------------------------|
| (1) | Specific gravity | 0.53 |
| (2) | Bulk density (g/cc) | 0.37 |
| (3) | Porosity (%) | 84 |
| (4) | Surface area (m ² /g) | 27 |
| (5) | Average particle size | 0.6–0.425 μm |
| (6) | Moisture content (%) | 39.2 |
| (7) | Loss in ignition | 93.22 (w/w %) |
| (8) | Al ₂ O ₃ | 2.2 (w/w %) |
| (9) | SiO ₂ | 1.05 (w/w %) |

carbon derived from bamboo and observed the BET surface area of 2,024 m²/g with a pore volume of 0.999 cm³/g.

2.3. Preparation of metal ion solutions

100 mg of potassium di-chromate (K₂Cr₂O₇), lead sulfate (PbSO₄), and zinc chloride (ZnCl₂) has been mixed with 1 L of double distilled water separately and the stock solutions were prepared. To obtain various design concentrations, the prepared water was diluted with double distilled water. Using 0.1 M of HCL, pH adjustments were carried out for the entire studies.

2.4. Batch adsorption studies

Adsorption of heavy metal ions onto the adsorbent has been conducted through batch mode by adjusting the parameters such as pH, contact time, adsorbent dose, metal ion concentrations, and temperature. The impact of the removal of heavy metal ions onto the adsorbent were examined with 50 mg/L of metal ion concentration, 100 mL of solution with pH of 2.0–7.0 at 30°C, for the equilibrium period of 60 min. By varying the adsorbent dosage from 0.5 to 2.5 g/L, the impact was analysed. The conical flasks were kept on the rotary shaker, shaken for 60 min

Table 4
Pore characteristics of SBP adsorbent

| Parameter | Value |
|--------------------------------------|---------|
| BET surface area, m ² /g | 558 |
| Pore volume, cm ³ /g | 0.275 |
| Micropore volume, cm ³ /g | 0.181 |
| Mesopore volume, cm ³ /g | 0.075 |
| Micropore area, m ² /g | 385 |
| Average pore diameter, Nm | 0.8–2.4 |

for ensuring attainment of equilibrium. The impact of contact time has been investigated by varying the time from 10 to 120 min. The total amount of metal ions adsorbed onto the adsorbent was calculated at various time intervals using Eq. (1) as given below:

$$q_t = \frac{(C_0 - C_t)V}{m} \text{ mg/g} \quad (1)$$

where q_t is the amount of metal ion adsorbed onto the adsorbent at any time “ t ” (mg/g), C_t is the concentration of batch adsorption processes.

The metal ion solutions were centrifuged for 5 min and residual concentration of metal ions was found out using atomic adsorption spectroscopy (AAS). Each analysis was repeated twice and the results were taken as an average of these values. The percentage removal of metal ions can be calculated by using the data obtained from batch studies. The mass balance relationship was obtained from Eq. (2):

$$\% \text{Removal} = \left(\frac{C_0 - C_e}{C_0} \right) \times 100 \quad (2)$$

where C_0 is the initial concentration of the metal ion solution (mg/L); C_e is the equilibrium concentration of the metal ion solution (mg/L); V is the volume of the metal ion solution; m is the mass of the adsorbent used.

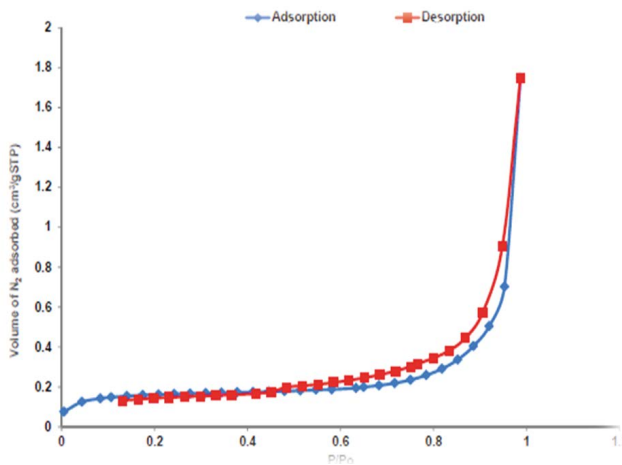


Fig. 1. Nitrogen adsorption–desorption isotherm.

2.5. Kinetic studies

For equilibrium studies, the initial concentration of metal ion solution has been varied from 25 to 150 mg/L with 2 g/L of SBP dose at the pH of 6.0 in 100 mL solutions. The temperature was fixed at 30°C and solutions were shaken up to 60 min at a speed of 120 rpm in a rotary shaker. Then the supernatant was taken for analysis using filter paper. The amount of metal ions adsorbed onto the adsorbent at equilibrium, q_e (mg/g) was calculated by Eq. (3):

$$q_e = \frac{(C_0 - C_e)V}{m} \text{ mg/g} \quad (3)$$

where C_0 and C_e are the initial and equilibrium metal ion concentrations respectively (mg/L); V is the volume of the metal ion solution; m is the mass of the adsorbent used.

Table 5 represents the various types of kinetic model studies used for this experimental work along with their constants and equations.

2.6. Temperature and thermodynamic studies

Batch adsorption experiments were performed at different temperatures of 15°C, 30°C, 45°C, and 60°C with the sugarcane bagasse powder for fixed initial metal ion concentration, adsorbent dose, and pH. To confirm whether the adsorption process is exothermic or endothermic in nature, thermodynamic parameters such as Gibbs free energy (ΔG°), enthalpy change (ΔH°), and entropy change (ΔS°), were calculated from the experimental data obtained from batch studies. The thermodynamic parameters, such as Gibbs free energy (ΔG°), enthalpy change (ΔH°), and entropy change (ΔS°) were calculated from Eqs. (4)–(6) as follows:

$$K_c = \frac{C_{Ae}}{C_e} \quad (4)$$

$$\Delta G^\circ = -RT \ln K_c \quad (5)$$

$$\log K_c = \frac{\Delta S^\circ}{2.303R} - \frac{\Delta H^\circ}{2.303RT} \quad (6)$$

where K_c is the equilibrium constant; C_e is the equilibrium of the metal ion concentration in solution (mg/L); C_{Ae} is the amount of metal ions adsorbed on the adsorbent per litre solution at equilibrium (mg/L); R is the gas constant – 8.314 J/mol K; T is the temperature.

2.7. Adsorption isotherm studies

2.7.1. Langmuir isotherm model

It is based mainly on few assumptions such as monolayer adsorption, and that adsorbate binding to the surface of adsorbent happens predominantly by chemical reactions, and all sites have an equivalent preference for adsorbate [25]. This isotherm follows few assumptions such as, all the process are homogeneous, only one adsorbate is used and these adsorbate molecules react with only one active site, in the adsorbate species no interactions are formed and the surface phase is monolayer [26].

The Langmuir isotherm equation can be expressed as in Eq. (7).

$$\frac{C_e}{q_e} = \frac{1}{Kq_{\max}} + \frac{C_e}{q_{\max}} \quad (7)$$

where C_e is the equilibrium concentration of the adsorbate solution; q_e is the amount of adsorbate adsorbed per g; K and q_{\max} is the Langmuir constants related to the adsorption capacity and intensity.

2.7.2. Freundlich isotherm model

This isotherm model allows the multi-layer adsorption on the adsorbent surface [27]. The Freundlich isotherm equation is expressed in Eq. (8):

$$\ln q_e = \ln k_f + \frac{1}{n} \ln C_e \quad (8)$$

where q_e is the amount of adsorbate adsorbed per g; n is the energy of intensity or adsorption; k_f is the Freundlich constant related to adsorption capacity; C_e is the equilibrium concentration of the adsorbate solution.

2.8. Interaction of heavy metal ions

The interaction of heavy metal ions onto the adsorbent in the adsorption process is of great significance for investigation, because of the presence of various metal ions in the effluent [28]. The impact of interaction with heavy metal ions on the adsorption of Cr(VI), Pb(II), and Zn(II) were tested by varying the initial metal ion concentrations from 25 to 150 mg/L.

3. Results and discussion

3.1. Scanning electron microscopy and energy dispersive X-ray analysis

The scanning electron microscopy (SEM) and energy dispersive X-ray (EDX) images of sugarcane bagasse powder and Cr(VI), Pb(II), and Zn(II) metal ions – adsorbed

Table 5
Different types of kinetic plots and their constants

| Type of kinetic model | Kinetic equation | Kinetic constants | References |
|--------------------------|---|--------------------|------------|
| Pseudo-first-order | $\log(q_e - q) = \log q_e - \frac{k}{2.303}t$ | k and q_e | [20] |
| Pseudo-second-order | $\frac{t}{q} = \frac{1}{h} + \frac{1}{q_e}t$ | k, q_e , and h | [21] |
| Elovich | $q_t = \frac{1}{b} \ln(1 + abt)$ | a and b | [22] |
| Boyd | $B = \frac{\Pi^2 D_i}{r^2}$ | B and D_i | [23] |
| Intra-particle diffusion | $q_t = k_p t^{1/2} + C$ | k_p and C | [24] |

sugarcane bagasse powder are shown in Figs. 2a and b. From Fig. 2a, it was observed that the surface was porous and the material was irregular in shape. From this, the surface morphology of the adsorbent can be identified for the adsorption of metal ions. Fig. 2b shows the loaded metal ions of Cr(VI), Pb(II), and Zn(II) onto the sugarcane bagasse powder. At the time of adsorption process,

permeable surface on the adsorbent (sugarcane bagasse) gets occupied by metal ions. On account of the presence of functional groups in bagasse surface inner walls, the metal ions were adsorbed onto them [29]. From the above surface morphology study, it becomes evident that sugarcane bagasse powder has been loaded with metal ions and can be absorbed in the internal walls of the adsorbent surface.

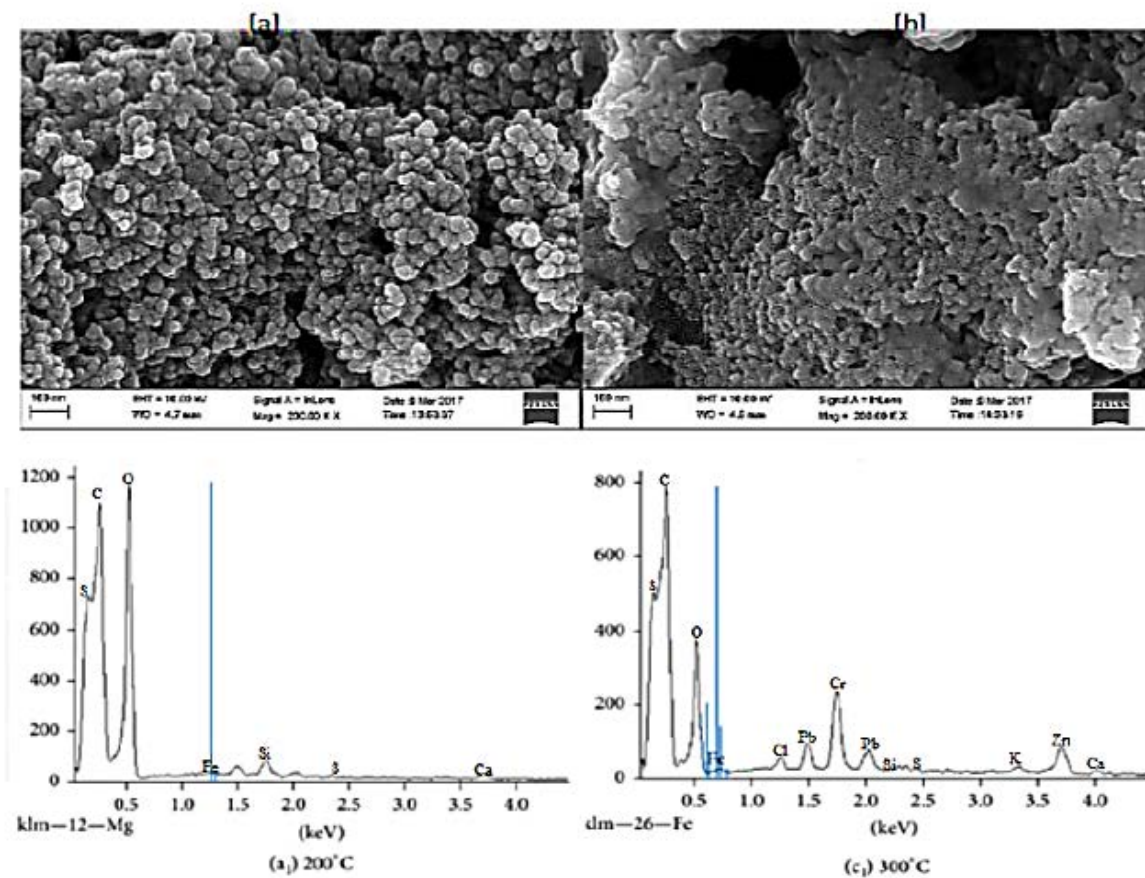


Fig. 2. (a and b) SEM/EDX images of raw and metal ion loaded adsorbent.

The EDX image of raw and metal ion loaded adsorbent has been shown in Figs. 2a and b and it confirm the presence of chromium, lead, and zinc ions after the adsorption process. Apart from chromium, lead, and zinc ions adsorbed, the raw adsorbent such as carbon, oxygen, and sulfur are seen in the EDX image of metal ion loaded adsorbent. Referring to Figs. 2a and b, additional metal elements/components were identified such as calcium, silica, and chlorine. Due to the acid treatment of sugarcane bagasse powder, the concentrated sulphuric acid reacts with hydroxyl groups present in the adsorbent and forms sulphuric esters as a non-ionic functional group, which may complex cations [30]. Hence it is concluded that the treatment with sulfuric acid solution protonates the charged sites and does not destroy any significant functional groups of raw adsorbent [31].

3.2. Fourier transmission infrared studies

The Fourier transmission infrared (FT-IR) analysis of sugarcane bagasse and Cr(VI), Pb(II), and Zn(II) metal ion loaded sugarcane bagasse is shown in Fig. 3. From Fig. 3, it is clear that there is a presence of $-OH$ and $-CH_2$ groups ($3,420$ and $2,860\text{ cm}^{-1}$) in the high energy region. The band range of $1,000\text{--}1,800\text{ cm}^{-1}$ represents the vibrations of aqueous solutions such as water ($1,620\text{ cm}^{-1}$), around $1,600\text{--}1,460\text{ cm}^{-1}$ for aromatic vibrations, and in the range of $1,380$ and $1,400\text{ cm}^{-1}$ for $-CH_2$ bending vibrations and $1,080\text{ cm}^{-1}$ for C–O vibrations. If the aromatic vibrations are due to C–H bending, the peak goes below $1,000\text{ cm}^{-1}$ [32]. The spectrum of sugarcane bagasse and metal ions (Cr, Pb, and Zn) loaded sugarcane bagasse indicates the aromatic ring vibrations and $-OH$ stretching because of low frequencies [33]. Further, the $-CH_2$ stretching vibrations will disappear at $2,860\text{ cm}^{-1}$. From sugarcane bagasse and metal ion loaded sugarcane bagasse FT-IR spectra, it was observed that there is a presence of active sites in the adsorbent for the adsorption process.

3.3. Impact of solution pH on metal ion adsorption

The impact of solution's pH with the removal of metal ions such as Cr(VI), Pb(II), and Zn(II) were investigated using sugarcane bagasse as an adsorbent, which had a pH range of 2.0 to 7.0. This study shows that the percentage of metal ions diminishes with the rise in pH of the solution up to 6.0 and their effect reduces subsequently. The surface of the adsorbent was enclosed by hydronium ions at a lower pH level as shown in Fig. 4. The interaction between metal ions and adsorbent has been reduced due to the presence of a high positively charged adsorbent surface. When the pH is low, the surface of the adsorbent becomes less positively charged stimulating faster removal of heavy metal ions [26]. The metal ions become less stable when the pH of the solution increases and adsorption reaches equilibrium at 6.0. At higher pH values, there is a decrease in metal ion removal, which can be attributed to precipitation into their respective hydroxides. The sugarcane bagasse powder can be used as an adsorbent up to a maximum range of 93.55% for Cr(VI), 87.56% for Pb(II), and 83.27% for Zn(II) at an optimum pH.

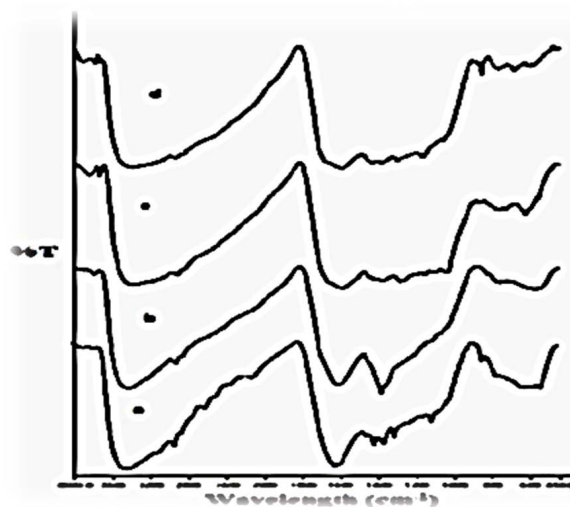


Fig. 3. FTIR spectra of (a) raw and (b–d) metal ions (Cr(VI), Pb(II), and Zn(II)) loaded adsorbent.

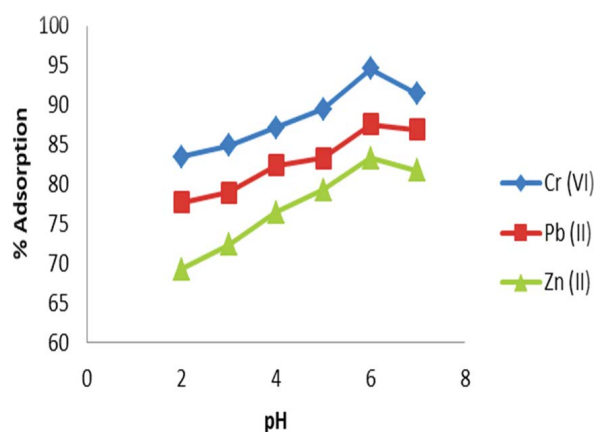


Fig. 4. Impact of pH on metal ion removal by SBP (concentration = 50 mg/L , dose = 1.5 g/L , time = 90 min , and temperature = 30°C).

3.4. Impact of adsorbent dosage on metal ion adsorption

Adsorbent dosage specifies the amount of active sites and serves as a significant criterion for the initial and final concentration of metal ions [34]. Fig. 5 indicates the influence of various dosages of sugarcane bagasse on the removal efficiency of metal ions such as Cr(VI), Pb(II), and Zn(II), respectively. The maximum range of eliminating various heavy metals was found to be 93.55% for Cr(VI), 77.7% for Pb(II), and 59.2% for Zn(II) at a dosage of 2.5 g/L . The number of available active sites in the adsorbent has been found to be proportional to the efficiency of heavy metals removal from an aqueous solution.

3.5. Impact of contact time on metal ion concentration

The impact of contact time between adsorbent particles and adsorbate is a vital parameter. Investigations were carried out by varying the concentration ($25\text{--}150\text{ mg/L}$)

of metal ion solution with different contact periods. The contact time was adjusted from 10 to 120 min for each adsorption study (Figs. 6a–c). During the initial stages, the removal rate of metal ions (Cr, Pb, and Zn) was found to be rapid due to the higher availability of vacant sites. In

addition, heavy metal ions were adsorbed into the mesopores that get saturated [22]. After a period of 60 min, there was no significant change in metal ion removal, that is, it remains constant. This is due to repulsive forces on the solid surface between adsorbate molecules [35]. Hence, the mass transfer between the phase of solids and liquids decreases with the passage of time. Furthermore, the metal ions have to travel farther and deeper through the pores with even higher strength. This results in slowing down the adsorption during the later phases.

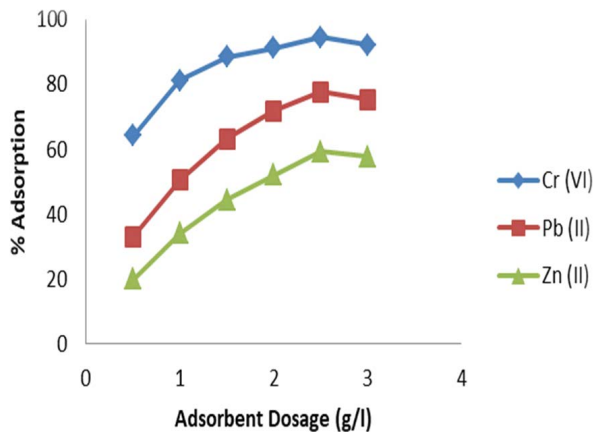


Fig. 5. Impact of adsorbent dose on metal ion removal by SBP (concentration = 50 mg/L, pH = 6.0, time = 90 min, and temperature = 30°C).

3.6. Impact of temperature on metal ion adsorption

The impact of temperature on the metal ion adsorption process is related to many thermodynamic parameters. Studies are conducted for varying temperatures (15°C–60°C) while the other parameters like adsorbate concentration, adsorbent concentration, and equilibrium time were kept constant and had values of 50 mg/L, 2.5 g/25 mL, and 60 min, respectively. The efficiency of the adsorption process gradually increased and attained a constant rate at a temperature of 45°C (optimum). Beyond this temperature, the efficiency gradually decreased due to an increase in desorption rate. Fig. 7 shows the adsorption efficiency of adsorbents in various temperature levels from 15°C to 60°C.

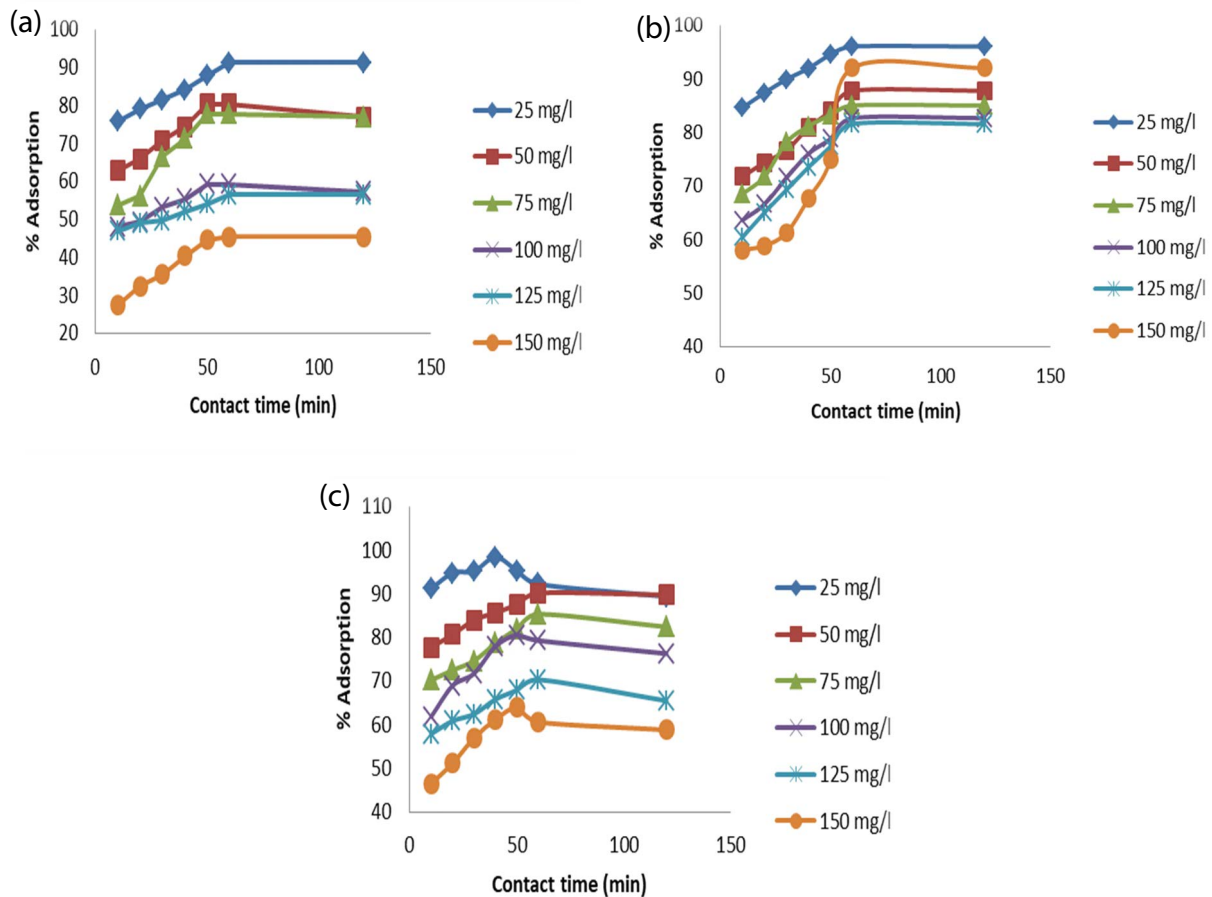


Fig. 6. (a–c) Impact of contact time on Cr(VI), Pb(II), and Zn(II) adsorption (pH 6.0, dose = 2.5 g/L, temperature = 30°C, and concentration = 25–150 mg/L).

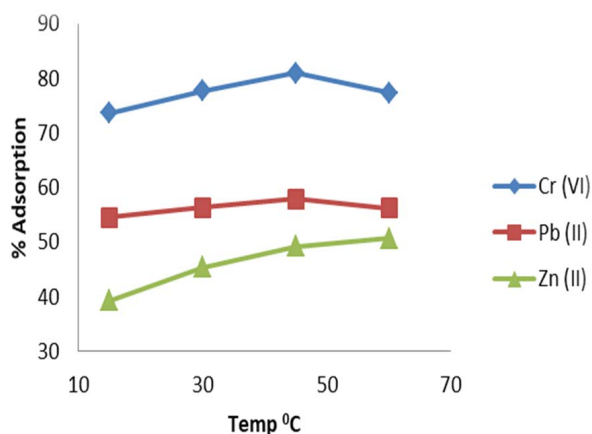


Fig. 7. Impact of temperature on metal ion adsorption (pH = 6.0, time = 60 min, concentration = 50 mg/L, and dose = 2.5 g/L).

3.7. Interaction of metal ions

From Fig. 8, the percentage of adsorption was gradually decreased from 98.63% to 64.26% for Cr(VI), 89.27% to 68.92% for Pb(II), and 82.39% to 68.82% for Zn(II), with an increase in metal ion concentration (25–150 mg/L). The amount of adsorption decreases uniformly with an increase in metal ion concentration indicates that adsorbent material has not reached the saturation level. Based on the above results, it was identified that the amount of adsorption shows a substantial difference between single and multi-heavy metal ions adsorbed onto sugarcane bagasse powder. Compared to Pb(II) and Zn(II), the adsorption rate of Cr(VI) is high. For 150 mg/L of multi-metal ion concentration, the sugarcane bagasse powder adsorbed 68.92% of Cr(VI) single metal ions. A similar trend has been followed for Pb(II) and Zn(II) metal ions also. Compared to multi-metal ion analysis, single metal ion adsorption provides the maximum adsorption efficiency due to the availability of a large amount of strived adsorption sites. The competitions between heavy metal ions are very much stronger in the higher concentration level of the mixed stage. The Pb(II) and Zn(II) can be replaced by Cr(VI) to be adsorbed on the surface of the adsorbent. Therefore, Cr(VI) was more competitive than Pb(II) and Zn(II). The performance of adsorbent for removal of Cr(VI) is higher due to high atomic weight and paramagnetic of metal ions, reduction potential is very high as compared to Pb(II) and Zn(II). Also, the hydration energy and ionic radius of metal ions are very low which are easier for adsorption [36,37].

3.8. Kinetic studies on metal ion adsorption

3.8.1. Pseudo-first-order kinetic model

Pseudo-first-order reaction model explains the adsorption kinetics. The plot of $\log(q_e - q)$ vs. time " t " is shown in Fig. 9 for various metal ions (Cr(VI), Pb(II), and Zn(II)). The rate constant " k " and correlation coefficient " R^2 " were calculated for different metal ion concentrations (25–150 mg/L) and represented in Table 6. From those observations, it was found that the regression coefficients (R^2) are in good contract with pseudo-first-order kinetics [38].

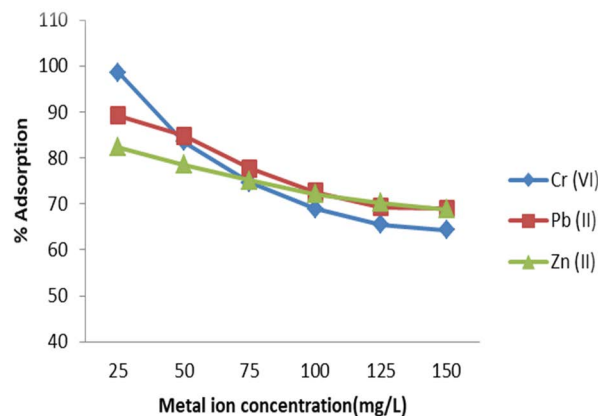


Fig. 8. Interaction of metal ions in multi-stage level (adsorbent dose = 2.0 g/L, pH = 2.0, time = 90 min, and temperature = 30°C).

3.8.2. Pseudo-second-order kinetic model

Pseudo-second-order kinetic model may also be relevant to the kinetics of sorption. The kinetics plots of t/q vs. t were made at different adsorbate concentrations is shown in Fig. 10 for different metal ions (Cr(VI), Pb(II), and Zn(II)). The correlation coefficients, h and k values are calculated from the plot as shown in Table 6. From the above discussion, it can be decided that the adsorption process of Cr(VI), Pb(II), and Zn(II) metal ions using sugarcane bagasse powder is best suited to pseudo-second-order kinetics study as " R^2 " values matched well [39]. The findings suggested that the observed R^2 values for pseudo-first-order equation were small and the discrepancy between measured q_e and experimental q_e values was large. Hence, the pseudo-second-order model was a better choice of study for metal ion adsorption using sugarcane bagasse powder [27].

3.8.3. Elovich kinetic model

Using sugarcane bagasse powder, the kinetics of metal ion adsorption is verified with Elovich kinetic model by plotting q_i vs. $\ln(t)$ and the findings are seen in Fig. 11 for Cr(VI), Pb(II), and Zn(II) metal ions respectively with various concentrations. The Elovich kinetic constants (a and b) were estimated from the plot and are mentioned in Table 6. The obtained values of regression coefficient (R^2) from the Elovich kinetic model were lower than the values obtained from pseudo-second-order kinetic model. But, the Elovich model is used for explaining the heterogeneous adsorbents in the adsorption process [18].

3.8.4. Boyd kinetic model

B_i vs. t plots were used to check the experimental variables and their linearity. If the plots passes through the origin and are of linear type, then the intra-particle diffusion is the slowest stage in the adsorption method, and vice versa. Fig. 12 shows the metal ion adsorption of Cr(VI), Pb(II), and Zn(II) using sugarcane bagasse powder, for various concentrations, which indicates that the plots are circular but do not move into the origin [40]. It shows

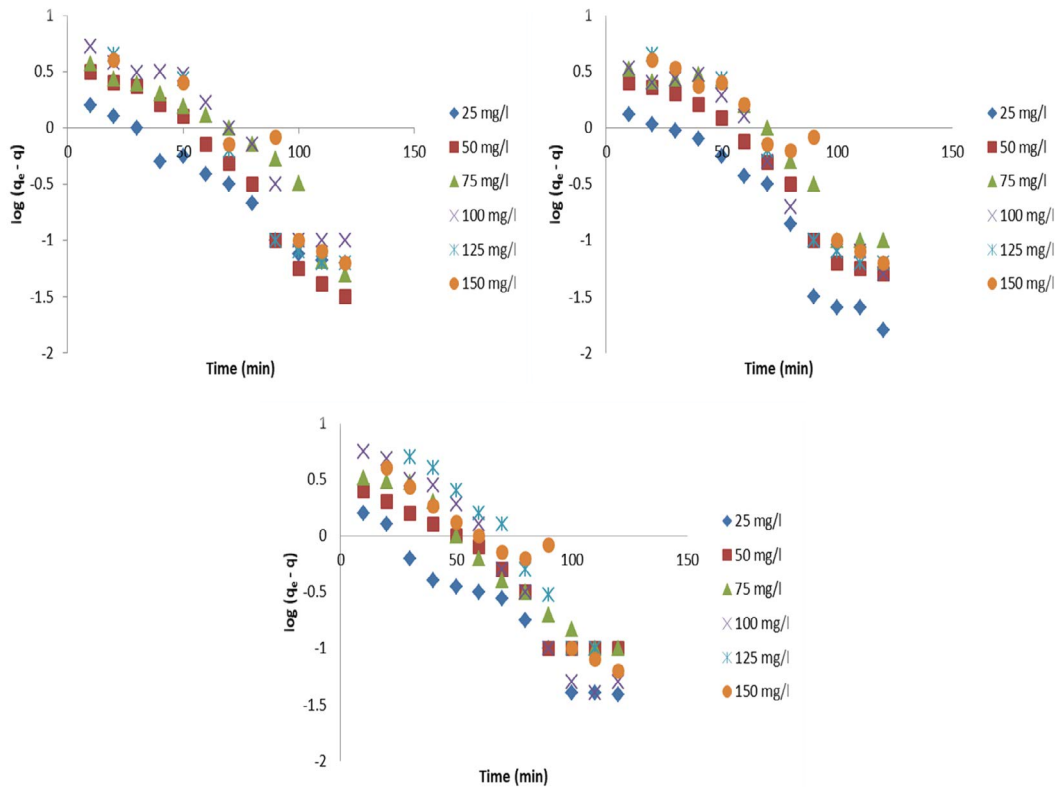


Fig. 9. Pseudo-first-order kinetic plots of Cr(VI), Pb(II), and Zn(II) metal ions.

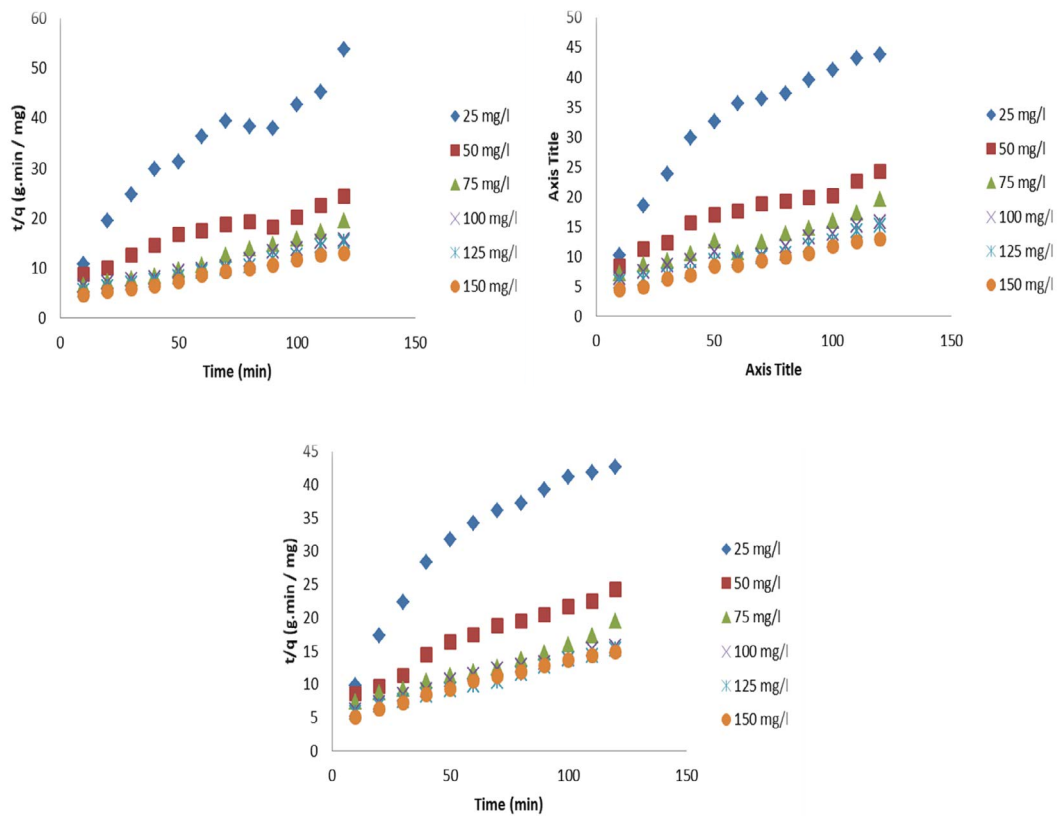


Fig. 10. Pseudo-second-order plots of Cr(VI), Pb(II), and Zn(II) metal ions.

Table 6
Kinetic constants and its parameters of different kinetic models

| S.No. | Type of metal | Conc. (mg/L) | Pseudo-first-order | | | Pseudo-second-order | | | Elovich | | | Boyd | | | Intra-particle diffusion | | | |
|-------|---------------|--------------|---------------------------|---------------------------|-------|---------------------------------|---------------------------|----------------|---------|----------------|------------|-------|-------|--|--------------------------|----------------------------------|-------|-------|
| | | | k (min^{-1}) | $q_{e,\text{cal}}$ (mg/g) | R^2 | k (g/mg min) $\times 10^{-3}$ | $q_{e,\text{cal}}$ (mg/g) | h (mg/g min) | R^2 | a (mg/g min) | b (g/mg) | R^2 | B | D_i ($\times 10^{-3}$ m^2/s) | R^2 | k_p (mg/g $\text{min}^{1/2}$) | C | R^2 |
| 1 | | 25 | 0.034 | 2.640 | 0.950 | 16.690 | 2.150 | 0.100 | 0.960 | 0.244 | 1.550 | 0.940 | 0.034 | 5.472 | 0.915 | 0.197 | 0.410 | 0.950 |
| 2 | | 50 | 0.043 | 7.020 | 0.930 | 5.730 | 5.190 | 0.180 | 0.980 | 0.988 | 0.740 | 0.930 | 0.044 | 7.340 | 0.973 | 0.399 | 0.580 | 0.950 |
| 3 | Cr(VI) | 75 | 0.041 | 10.000 | 0.930 | 3.340 | 8.300 | 0.220 | 0.980 | 0.978 | 0.550 | 0.920 | 0.043 | 7.621 | 0.963 | 0.517 | 0.600 | 0.980 |
| 4 | | 100 | 0.039 | 11.360 | 0.940 | 5.070 | 10.750 | 0.260 | 0.980 | 0.948 | 0.440 | 0.950 | 0.039 | 6.856 | 0.914 | 0.783 | 0.540 | 0.990 |
| 5 | | 125 | 0.048 | 17.470 | 0.920 | 2.000 | 12.910 | 0.290 | 0.970 | 0.977 | 0.350 | 0.920 | 0.049 | 8.725 | 0.982 | 0.804 | 0.370 | 0.930 |
| 6 | | 150 | 0.045 | 19.430 | 0.930 | 3.120 | 13.820 | 0.300 | 0.960 | 0.934 | 0.260 | 0.920 | 0.045 | 7.452 | 0.952 | 0.856 | 0.320 | 0.960 |
| 7 | | 25 | 0.046 | 3.680 | 0.910 | 12.620 | 2.700 | 0.100 | 0.970 | 0.263 | 1.160 | 0.910 | 0.046 | 7.678 | 0.943 | 0.190 | 0.380 | 0.980 |
| 8 | | 50 | 0.041 | 6.540 | 0.930 | 5.320 | 5.470 | 0.120 | 0.980 | 0.328 | 0.810 | 0.940 | 0.041 | 6.294 | 0.983 | 0.348 | 0.510 | 0.940 |
| 9 | Pb(II) | 75 | 0.043 | 9.950 | 0.920 | 3.560 | 8.600 | 0.230 | 0.980 | 0.541 | 0.650 | 0.930 | 0.045 | 7.959 | 0.924 | 0.554 | 0.680 | 0.950 |
| 10 | | 100 | 0.046 | 12.380 | 0.940 | 2.770 | 10.100 | 0.280 | 0.970 | 0.611 | 0.470 | 0.920 | 0.046 | 7.678 | 0.983 | 0.686 | 0.670 | 0.950 |
| 11 | | 125 | 0.052 | 20.550 | 0.920 | 2.300 | 11.560 | 0.310 | 0.980 | 0.618 | 0.310 | 0.900 | 0.053 | 8.590 | 0.941 | 0.772 | 0.510 | 0.970 |
| 12 | | 150 | 0.050 | 25.480 | 0.940 | 2.650 | 12.730 | 0.330 | 0.960 | 0.596 | 0.200 | 0.920 | 0.055 | 8.972 | 0.922 | 0.805 | 0.320 | 0.960 |
| 13 | | 25 | 0.039 | 2.840 | 0.950 | 15.430 | 2.140 | 0.090 | 0.960 | 0.230 | 1.630 | 0.940 | 0.037 | 6.428 | 0.993 | 0.189 | 0.300 | 0.970 |
| 14 | | 50 | 0.032 | 6.510 | 0.930 | 5.310 | 5.550 | 0.150 | 0.990 | 0.323 | 0.870 | 0.960 | 0.041 | 6.492 | 0.939 | 0.318 | 0.370 | 0.980 |
| 15 | Zn(II) | 75 | 0.041 | 10.400 | 0.910 | 3.640 | 7.740 | 0.210 | 0.970 | 0.484 | 0.650 | 0.940 | 0.046 | 7.687 | 0.920 | 0.514 | 0.420 | 0.940 |
| 16 | | 100 | 0.044 | 13.270 | 0.950 | 2.860 | 9.400 | 0.270 | 0.980 | 0.572 | 0.460 | 0.930 | 0.050 | 8.344 | 0.942 | 0.677 | 0.410 | 0.950 |
| 17 | | 125 | 0.044 | 18.350 | 0.940 | 1.350 | 12.640 | 0.230 | 0.980 | 0.652 | 0.360 | 0.920 | 0.049 | 8.725 | 0.941 | 0.746 | 0.320 | 0.960 |
| 18 | | 150 | 0.051 | 23.470 | 0.950 | 1.110 | 15.230 | 0.250 | 0.950 | 0.549 | 0.250 | 0.910 | 0.045 | 8.921 | 0.955 | 0.840 | 0.270 | 0.970 |

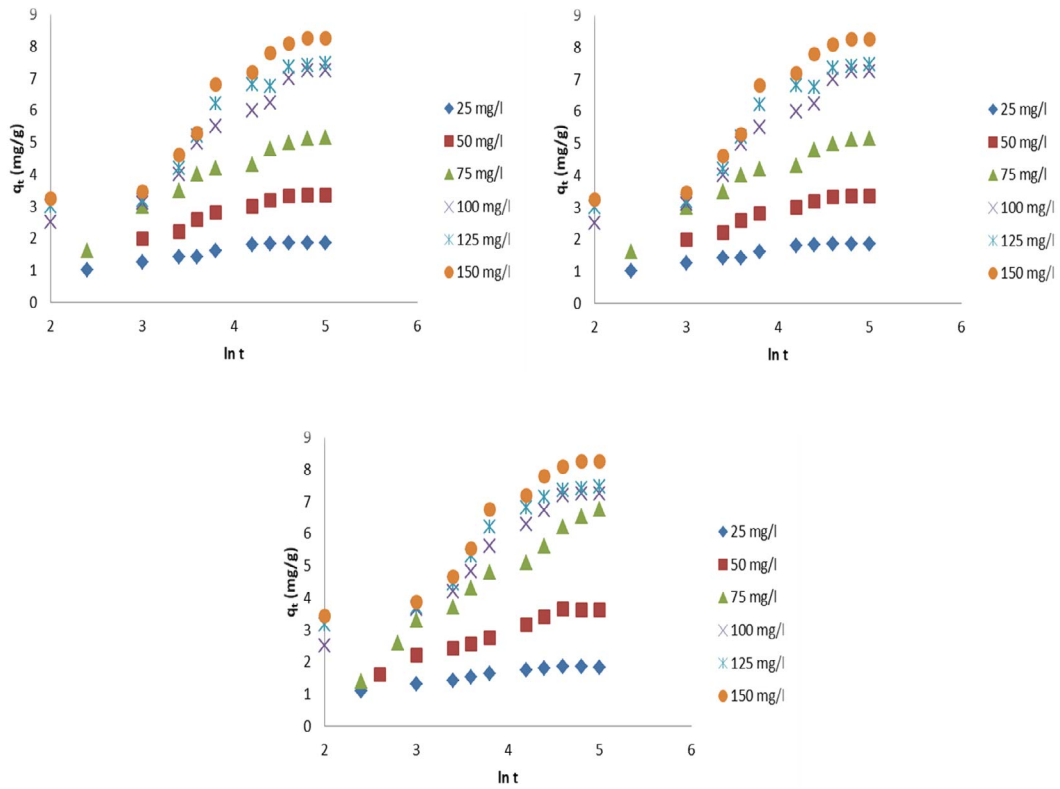


Fig. 11. Elovich kinetic plots of Cr(VI), Pb(II), and Zn(II) metal ions.

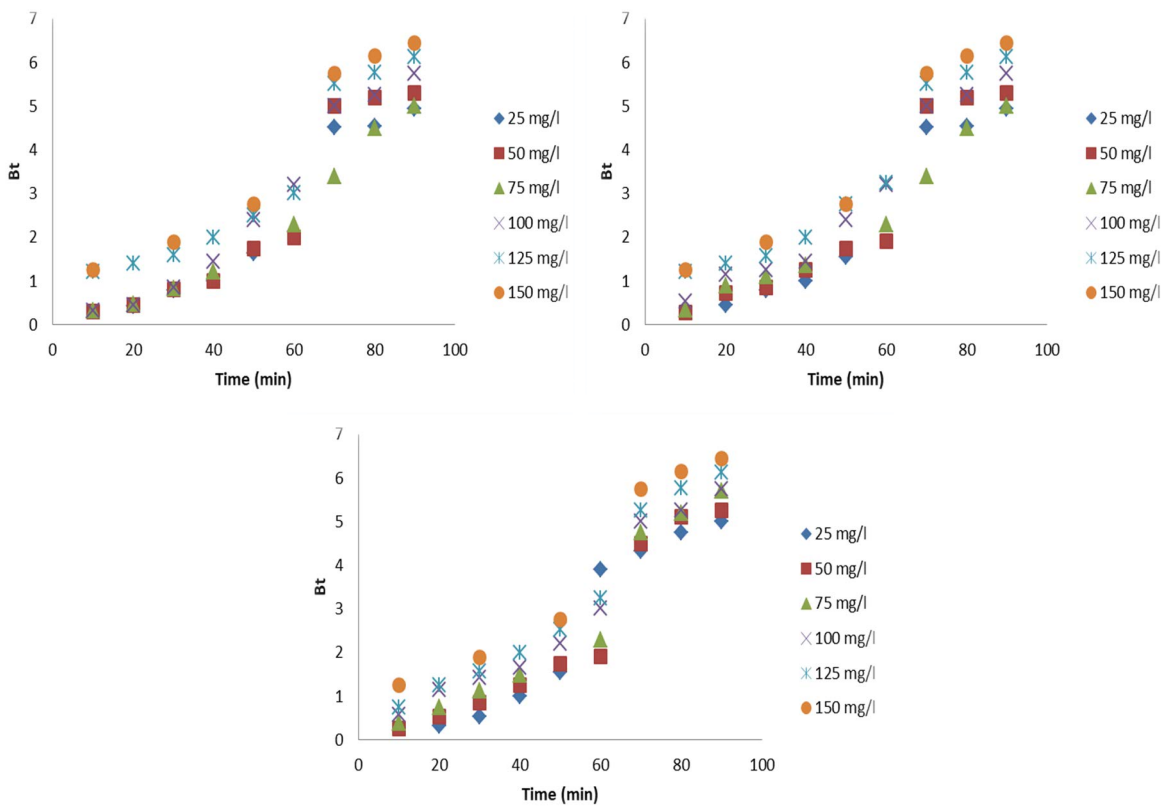


Fig. 12. Boyd kinetic plots of Cr(VI), Pb(II), and Zn(II) metal ions.

that the adsorption of metal ions to sugarcane bagasse powder is regulated by external or film diffusion [41]. The values of D_i and B were calculated from the plot and shown in Table 6 along with regression coefficient “ R^2 ” values.

3.8.5. Intra-particle diffusion model

The plot of q_t vs. $t^{0.5}$ is shown in Fig. 13 for Cr(VI), for Pb(II) and Zn(II) metal ions, respectively, in different concentrations and the values of IPD constants (k_p , C , and R^2) were calculated (Table 6). The IPD plots pass through the origin, then metal ions adsorption to the adsorbents is regulated by intra-particle diffusion, whereas if the data display a multi-linear curve pattern, two or more phases influence the adsorption cycle [16,42]. The dual structure of the curve is characterized by the differing degrees of adsorption in the preliminary and final stages of adsorption with sugarcane bagasse powder. It may be recognized to the assumption that at the early stage, adsorption was due to diffusion effect of boundary layer and, at a later point to diffusion impact of intraparticle [40]. By referring to Fig. 13, it can be shown that the plots are linear and the mechanism is driven by film diffusion.

3.9. Adsorption isotherm studies

3.9.1. Langmuir adsorption isotherm

The Langmuir isotherm plots of C/q_e vs. C_e is shown in Fig. 14 for Cr(VI), Pb(II), and Zn(II) metal ions adsorption using sugarcane bagasse powder. The constants of the

Langmuir isotherm study (“ k ,” q_{max} , and R^2) were obtained from Fig. 14 and listed in Table 7 at the temperature of 30°C. From that Table 7, it was observed that the regression coefficients (R^2) are <1 for all three heavy metal ions with 100 mg/L concentrations. The reported R^2 values are between 0 and 1 for all metal ions tested with sugarcane bagasse powder, which means that the adsorption is favorable [35].

3.9.2. Freundlich adsorption isotherm

The Freundlich isotherm constants k_f and n were calculated and listed in Table 7, by referring to the plots of $\ln q_e$ vs. $\ln C_e$ (Fig. 15). The degree of non-linearity between adsorption and metal ion concentration was indicated based on the “ n ” values [13]. The adsorption process is in the linear stage the value of “ n ” equal to 1. If the value of “ n ” is less than 1, it indicates that the process is chemically oriented and “ n ” value is greater than 1, it is physically oriented [14]. The Freundlich constants (k_f and n) with R^2 values were derived from the plot of $\ln q_e$ vs. $\ln C_e$ at 30°C, and are listed in Table 7. The “ n ” values were found to be 2.936, 2.351, and 2.062 for Cr(VI), Pb(II), and Zn (II) ions, respectively, for sugarcane bagasse powder adsorbent. Since the “ n ” values are greater than 1 it indicates that the metal ion adsorption follows the physical adsorption process [6]. Experimental results from the analysis of the preliminary accumulation of metal ions onto sugarcane bagasse powder surfaces are fitted with Langmuir and Freundlich adsorption isothermal models, and the plots of these models are shown in Figs. 14 and 15. From the obtained R^2 values, the best fit of the adsorption isotherms

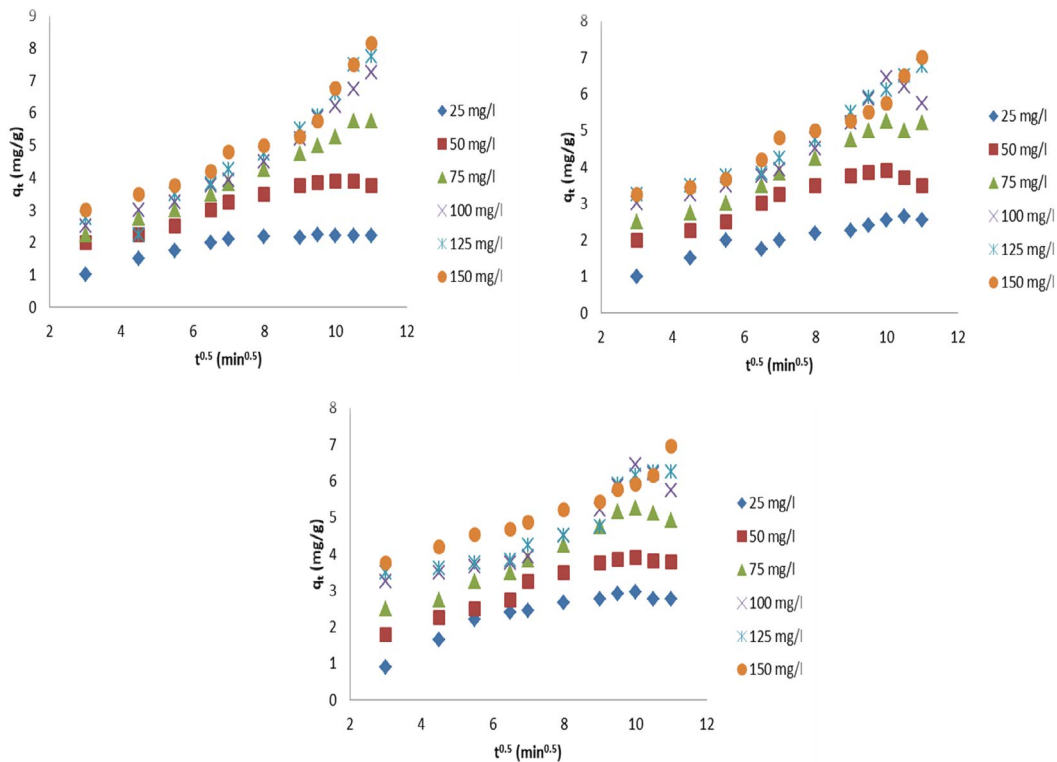


Fig. 13. IPD kinetic plots of Cr(VI), Pb(II), and Zn(II) metal ions.

Table 7
Isotherm constants for the removal of heavy metal ions by sugarcane bagasse

| Isotherm model | Metal ion solution | | | |
|----------------|---------------------------------------|--------|--------|--------|
| | Parameters | Cr(VI) | Pb(II) | Zn(II) |
| Langmuir | q_{\max} (mg/g) | 8.304 | 8.992 | 9.344 |
| | K_L (L/mg) | 0.334 | 0.147 | 0.104 |
| | R^2 | 0.9715 | 0.9982 | 0.9866 |
| Freundlich | K_f ((mg/g) (L/mg) ^{1/n}) | 2.514 | 1.823 | 1.399 |
| | n (g/L) | 2.936 | 2.351 | 2.062 |
| | R^2 | 0.9508 | 0.9785 | 0.9400 |

studied with different heavy metal ions with sugarcane bagasse powder, for Cr(VI): Langmuir > Freundlich isotherm, for Pb(II): Langmuir > Freundlich isotherm, for Zn(II): Langmuir > Freundlich isotherm.

Referring to regression coefficient (R^2) values, adsorption of metal ions using sugarcane bagasse powder was fitted well with Langmuir isotherm model, than by Freundlich isotherm model. This is established on adsorption of metal ions by bagasse in the monolayer [9]. Based on the above observations, metal ions adsorption with the adsorbent follows heterogeneous and monolayer adsorption mechanism [43].

3.10. Impact of temperature and thermodynamic studies

The impact of temperature on metal ion removal has been investigated at different temperatures (15°C, 30°C,

45°C, and 60°C) and represented in Fig. 16, which indicates the importance of temperature in the adsorption process. At 20°C, the maximum efficiency rate was attained and after that, a gradual decrease in metal ion removal were identified with an increase in temperature. This may suggest that adsorption process between the adsorbent and heavy metal ions is an exothermic process and this was mainly due to a decrease in the active sites [44]. The values of ΔH° and ΔS° were determined from the slope and intercept of the plot of $\log K_c$ vs. $1/T$ as shown in Fig. 17. Table 8 indicates the values of ΔG° , ΔH° , and ΔS° in different temperatures along with varying initial concentrations (25–150 mg/L) of heavy metal ions (Cr, Pb, and Zn). From the table, negative ΔG° values with positive ΔH° indicate the performance of metal ions with sugarcane bagasse powder and the spontaneous nature of adsorption. From the above discussion, it was observed that the adsorption reaction follows endothermic nature [45]. Furthermore, the positive value of ΔS° suggests the increases in the uncertainty of solid/liquid interface at the time of metal ions adsorption on sugarcane bagasse powder in the aqueous medium.

4. Conclusion

The performance of biosorption of heavy metal ions (Cr, Pb, and Zn) was investigated using sugarcane bagasse powder in the batch adsorption process. By varying the metal ion concentration, pH of the solution, contact time, adsorbent dosage level, and temperature, the impact of adsorption efficiency were investigated. From batch studies, it was observed that the adsorption of metal ions onto adsorbent was increased with an increase in contact time and adsorbent dosage level. On the other hand, the process

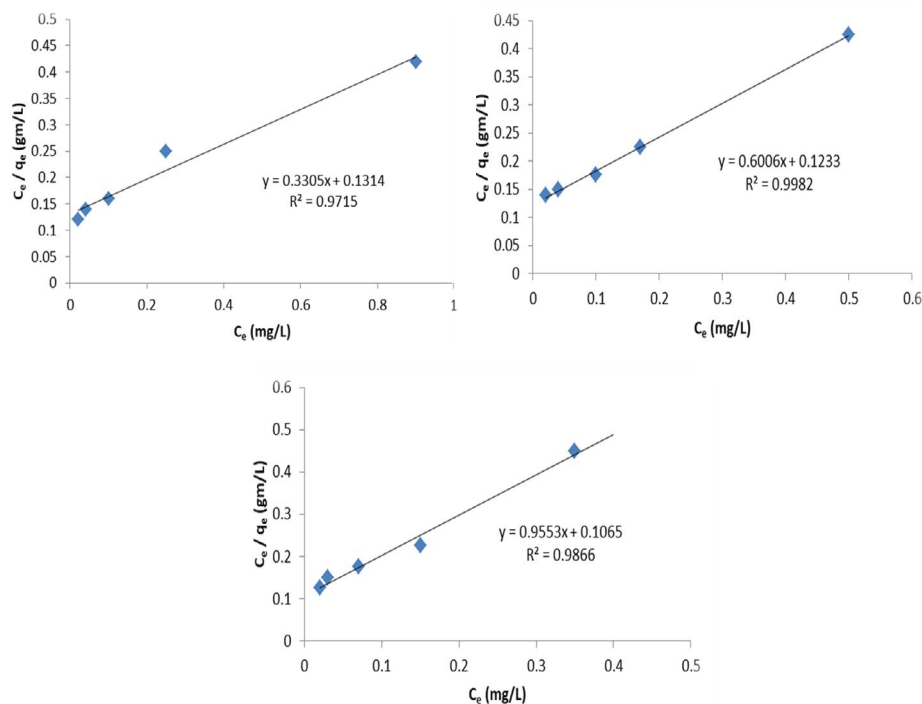


Fig. 14. Langmuir isotherm plots of Cr(VI), Pb(II), and Zn(II) metal ions.

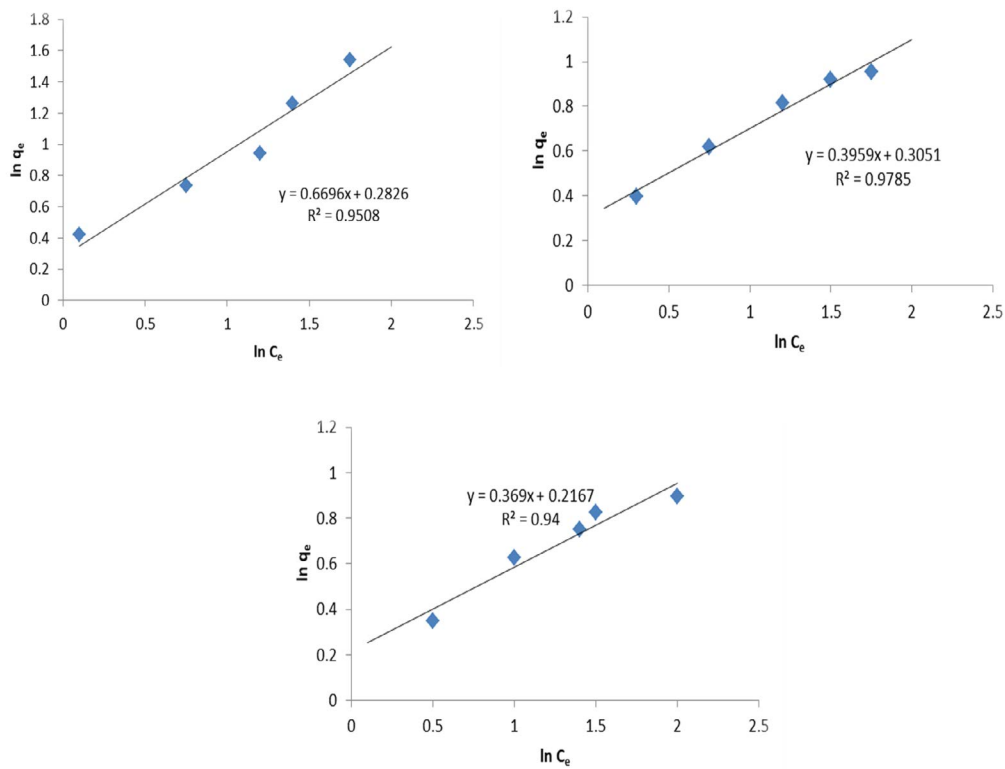


Fig. 15. Freundlich isotherm plots of Cr(VI), Pb(II), and Zn(II) metal ions.

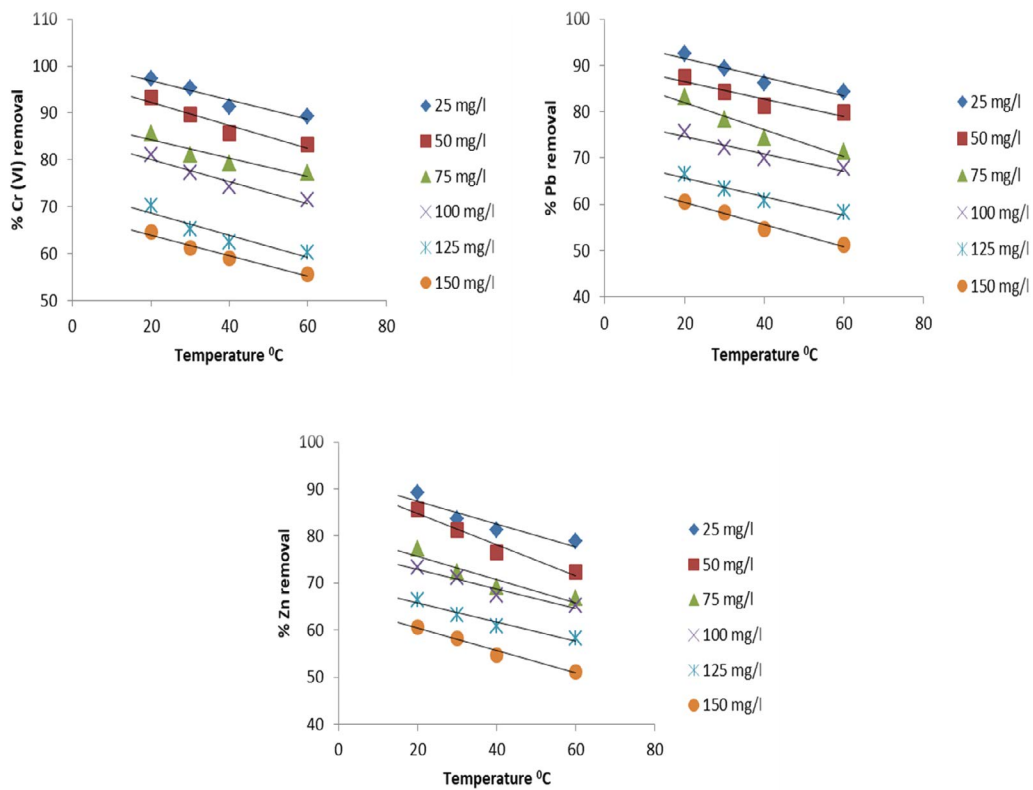


Fig. 16. Effect of temperature on the adsorption Cr(VI), Pb(II), and Zn(II) ions by SBP (initial ion concentration = 25–150 mg/L, dose = 1.2 g/L, pH = 6.0, and time = 60 min).

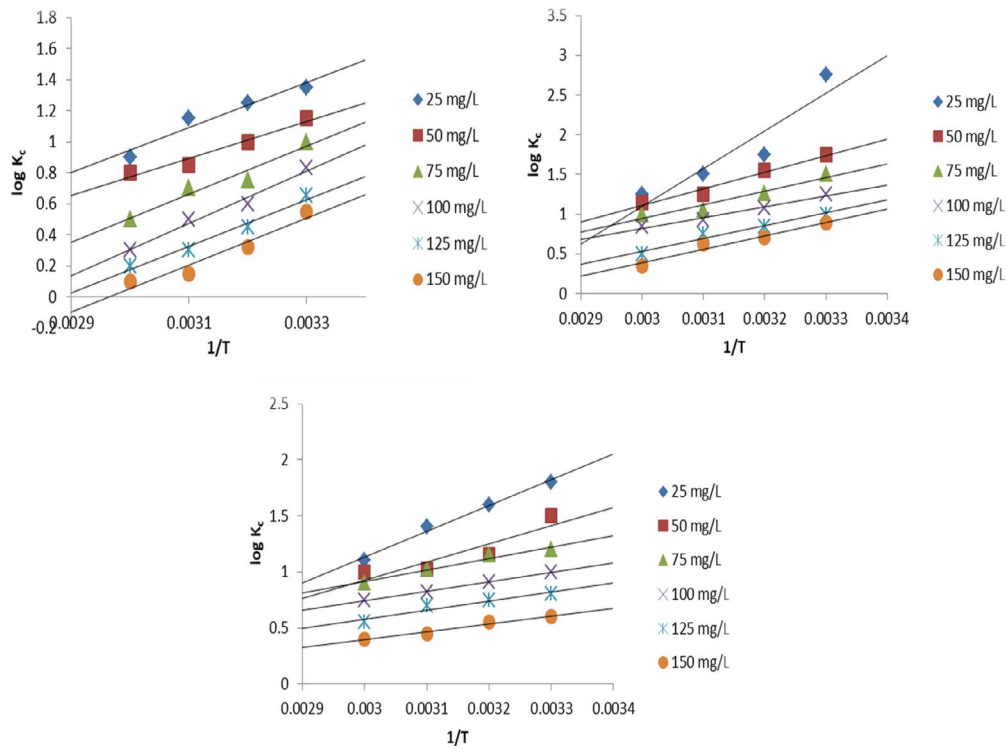


Fig. 17. Thermodynamic plots for the adsorption of Cr(VI), Pb(II), and Zn(II) ions using SBP.

Table 8
Thermodynamic parameters for the adsorption of metal ions using sugarcane bagasse

| Initial ion concentration of Cr(VI) ions | ΔH° (kJ/mol) | ΔS° (J/mol) | ΔG° (kJ/mol) | | | |
|--|---------------------------|--------------------------|---------------------------|--------|--------|--------|
| | | | 15°C | 30°C | 45°C | 60°C |
| 25 | 32.764 | 80.723 | -8.299 | -7.587 | -7.116 | -6.942 |
| 50 | 20.134 | 42.542 | -6.396 | -6.244 | -6.102 | -5.632 |
| 75 | 18.678 | 36.923 | -5.809 | -5.568 | -5.314 | -3.449 |
| 100 | 13.452 | 25.766 | -5.079 | -3.888 | -3.637 | -3.135 |
| 125 | 11.823 | 21.145 | -3.233 | -3.214 | -3.924 | -3.752 |
| 150 | 9.345 | 19.623 | -3.921 | -3.786 | -3.582 | -3.245 |
| Initial ion concentration of Pb(II) ions | | | | | | |
| 25 | 19.277 | 45.892 | -6.810 | -5.563 | -5.238 | -3.381 |
| 50 | 13.723 | 27.922 | -3.892 | -3.528 | -3.347 | -3.082 |
| 75 | 10.984 | 21.114 | -3.962 | -3.153 | -3.196 | -3.080 |
| 100 | 9.823 | 22.276 | -2.183 | -2.160 | -2.377 | -2.119 |
| 125 | 7.546 | 20.091 | -1.182 | -1.363 | -1.246 | -1.204 |
| 150 | 5.984 | 18.472 | -0.892 | -0.817 | -0.820 | -0.798 |
| Initial ion concentration of Zn(II) ions | | | | | | |
| 25 | 13.376 | 28.491 | -5.807 | -3.671 | -3.398 | -3.162 |
| 50 | 11.036 | 23.881 | -3.099 | -3.484 | -3.425 | -3.298 |
| 75 | 9.311 | 20.392 | -3.154 | -2.920 | -2.771 | -2.458 |
| 100 | 8.741 | 21.923 | -2.283 | -2.056 | -1.845 | -1.648 |
| 125 | 7.293 | 18.467 | -1.957 | -1.414 | -1.265 | -1.029 |
| 150 | 6.756 | 16.873 | -1.458 | -1.037 | -1.004 | -1.036 |

of metal ion adsorption was decreased with an increase in solution's pH, the concentration of the adsorbate, and changes in temperature. The Langmuir and Freundlich isotherms fitted with the equilibrium data and pseudo-second-order kinetic model data has been followed in this adsorption process. From the above discussions, this work concluded that the sugarcane bagasse powder is the most efficient adsorbent to remove the heavy metal contaminations from the aqueous solutions.

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