



Assessment of photocatalytic n-TiO₂/UV and n-TiO₂/H₂O₂/UV methods to treat DB 86, RY 145 and AV 90 dye mix containing wastewater

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ABSTRACT

This study examined and compared the n-TiO₂/H₂O₂/UV and n-TiO₂/UV processes to remove the content of synthetic wastewater containing Direct Blue (DB86), Reactive Yellow (RY145) and Acid Violet (AV90) dye mix. The optimum parameters for both processes have been determined as pH 3, 125 mg/L n-TiO₂ and 45 min, and the H₂O₂ dose has been specified as 750 mg/L in the n-TiO₂/H₂O₂/UV process. While the removal efficiencies were 54.38% at 436 nm, 44.20% at 525 nm and 85.72% at 620 nm for the n-TiO₂/UV process, 90.63% at 436 nm, 92.27% at 525 nm and 99% at 620 nm were obtained for n-TiO₂/H₂O₂/UV process. With H₂O₂ addition to the process, treatment efficiencies have been increased drastically. Moreover, due to the results of the kinetic analyses for both processes, the pseudo-second-order was observed to be the most appropriate kinetic model, depending on the regression coefficient in color removal.

Keywords: Direct Blue (DB86); Reactive Yellow (RY145); Acid Violet (V90); n-TiO₂/H₂O₂/UV; n-TiO₂/UV

1. Introduction

The textile industry is one of the oldest sectors, and the dimensions of the sector are expanding day by day due to the increasing population [1]. The total budget of the global textile industry has been estimated to be 920 billion USD. Until 2024 it has expected to reach 1,230 billion USD [2]. Despite the economic contribution of the textile industry, textile production contributes the environmental pollution, with the usage of vast amounts of water, fuel and chemicals [3].

Textile industry wastewaters include a mixture of dyes and a variety of chemicals [4]. Dyes are one of the larger groups of pollutants in wastewater released from textile industries. Discharged dyes to water bodies prevent light penetration and disturb aquatic life [5]. According to the

application types, the dyestuffs are classified as reactive dyes, disperse dyes, acid dyes, basic dyes, direct dyes, and vat dyes [6,7]. Most of the dyestuffs used are toxic, non-biodegradable and carcinogenic, and direct discharge to the receiving environment is hazardous for the environment and human health [8]. Many methods have been used in the literature to degrade textile dyestuffs and the purification of water.

In studies on the degradation of Direct Blue 86, Reactive Yellow 145 and Acid Violet 90 dyes (O₃ and O₃/UV), photocatalytic oxidation [9,10], adsorption [11–14], coagulation [15], persulfate degradation [16], electrocoagulation [17] methods were observed.

Within the scope of this study, the treatment of synthetic wastewater samples containing 150 mg/L dyestuffs (DB 86, RY 145 and AV 90 50 mg/L each) with n-TiO₂/UV

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and n-TiO₂/H₂O₂/UV processes was investigated. Although there are limited studies on the removal of these dyes in the literature, textile wastewater contains dyestuff mixtures.

Depending on the product and dye variety, textile industry wastewater contains a mixture of many dyes. For this reason, the sector demands to perform the removal of the mixed dyes used, rather than the removal of a single dye. In our study, the efficient removal of three different dye mixtures with the specified methods was investigated in a way that would meet the needs of the factory in which the dyes were supplied. Due to the carcinogenic, mutagenic and toxic properties of dyes used in the textile industry, it holds great importance to treat them in accordance with discharge regulations

In this context, this study is an original contribution to the literature by investigating the treatment of synthetic wastewater containing three different dyestuffs by the n-TiO₂/H₂O₂/UV and n-TiO₂/UV processes. In both processes, pH (3–11), n-TiO₂ dose 25–150 mg/L, and reaction time 5–60 min parameters were studied. For the n-TiO₂/H₂O₂/UV process, the H₂O₂ dose 250–1,500 mg/L parameter was also examined. By using all these parameters, the effects of the processes on the color removal were evaluated in the synthetic wastewater sample containing the mixture of three different dyestuffs (DB 86, RY 145, AV 90) in equal amounts.

2. Materials and methods

2.1. Materials

Direct Blue 86 (DB 86), Reactive Yellow 145 (RY 145) and Acid Violet 90 (AV 90) dyestuffs were provided from a textile factory in Sakarya. Hydrochloric acid (HCl) and sodium hydroxide (NaOH) at brand MERCK has been used to adjust the pH of the solutions. pH measurements have been performed pH meter at brand HANNA. Magnetic stirrers at brand MTOPS and NUVE brand centrifuge device has been used to remove n-TiO₂. 21 nm diameter n-TiO₂ was used as a catalyst at brand SIGMA ALDRICH. The diameter of commercially available n-TiO₂ was determined as 21 nm from the scanning electron microscopy (SEM) images given in Fig. 1. It can be seen clearly from the SEM image, n-TiO₂ that was used in the experiment has a homogeneous

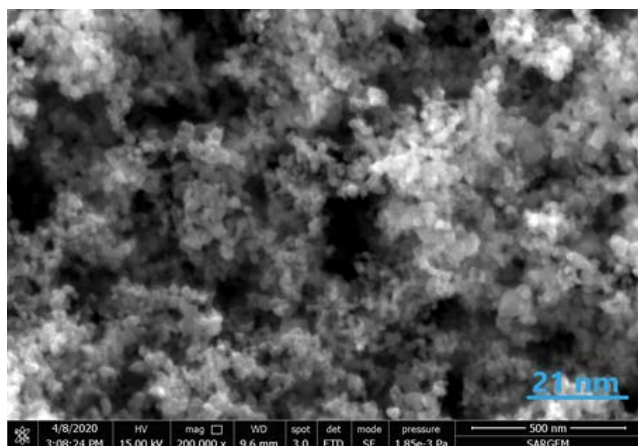


Fig. 1. SEM image for n-TiO₂.

structure. Dye mix concentrations have been measured with a Pharo 3600 spectrophotometer at brand MERCK. Zeta potential analyses have been measured with the Zeta sizer brand Nano-ZS model device.

2.2. Dyestuff mixture preparation and color measurement method

In this study, synthetic wastewater was prepared by mixing DB 86, RY 145, and AV 90 (each 50 mg/L) dyestuffs with the total 150 mg/L initial concentration. The color limit values according to European norms are given in Table 1. The initial values according to European norms in synthetic water containing dyestuffs used in the experiments are given in Table 2. Experiments were repeated three times, and the samples were centrifuged at 4,500 rpm for 15 min before each measurement for the removal of TiO₂ nanoparticles used. Color measurements were performed as EN ISO 7887-B. According to the method, the color is specified with measurements of attenuation of light absorption by using a spectrophotometer at a minimum of three different wavelengths. The measurements were performed with 436, 525 and 620 nm wavelengths of the visible spectrum with a 50 mm optical pathlength cell [18].

2.3. Reactor design

The reactor design of the n-TiO₂/UV and n-TiO₂/H₂O₂/UV processes used in color removal experiments is shown in Fig. 2. A photoreactor consisting of six pieces of 6-W lamps with a wavelength of 254 nm was used for the photocatalytic oxidation experiment. A cylindrical sample container made of quartz glass with a volume of 140 mL was used inside the reactor. A homogeneous mixture was achieved by placing the reactor on the magnetic stirrer. In order to keep the system temperature constant, a pump is used to provide cold air to the system.

2.4. Photocatalytic reactions

Various methods such as ozonation, irradiation, H₂O₂ and oxidation are used for the treatment of dye wastewater [19,20]. Due to the non-toxic, insoluble, inexpensive and highly reactive nature of TiO₂, it has been used to treat wastewater containing TiO₂/UV dyes. The purification mechanism of TiO₂ is given in Eqs. (1)–(3) [21].

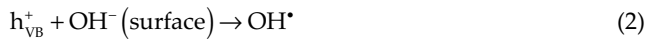


Table 1
European norm color limit values

436 nm	525 nm	620 nm
7 m ⁻¹	5 m ⁻¹	3 m ⁻¹

Table 2
Initial color values of synthetic wastewater of dye mix

436 nm	525 nm	620 nm
32 m ⁻¹	36.2 m ⁻¹	30.1 m ⁻¹



where $h\nu$ represents UV radiation; h_{VB}^+ is valence-band holes; e_{CB}^- is conduction-band electrons.

The OH^{\bullet} radical, which provides the removal of pollutants, can occur in various forms, as seen in the equations. Electrons that emerge as a result of photon irradiation can interact with other molecules as well as release heat by interacting with valence-band holes. These holes can provide oxidation by direct electron transfer. The mechanism of the n-TiO₂/H₂O₂/UV photocatalytic treatment system is briefly shown by the Eqs. (4)–(8) [21–23];



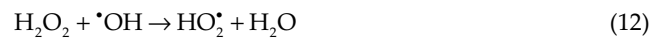
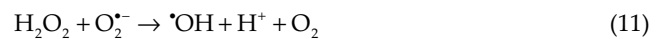
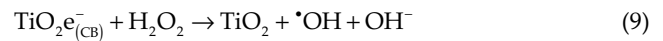
where h_{ν} represents UV radiation; h_{VB}^+ is valence-band holes; e_{CB}^- is conduction-band electrons.

3. Results and discussion

3.1. Effect of pH on color removal

The pH parameter affects the removal efficiency for dye removal in wastewater containing dyestuffs [24]. The experiments were carried out at pH 3, 5, 9 and 11 to investigate the effects of pH on TiO₂/UV and TiO₂/H₂O₂/UV processes for dye removal efficiency. In the beginning, synthetic textile wastewater with a total dye concentration of 150 mg/L was prepared, with 50 mg/L of each of three dyes. The initial color values of the Direct Blue 86, Reactive Yellow 145, and Acid Violet 80 dye mixture measured as 32 m⁻¹ at 436 nm, 36.2 m⁻¹ at 525 nm and 30.1 m⁻¹ at 620 nm. The experiments

have been performed with 100 mg/L n-TiO₂ dose, 30 min reaction time and 36 W light intensity to investigate the optimum pH value for both TiO₂/UV and TiO₂/H₂O₂/UV processes. Additionally, 1,000 mg/L H₂O₂ was used in the TiO₂/H₂O₂/UV process. The effect of pH on the color removal efficiency in photocatalytic processes using n-TiO₂ and photocatalytic processes with H₂O₂ is given in Fig. 3.



As can be seen in Fig. 3, the color removal efficiency decreases with the increase in pH in the n-TiO₂/UV and n-TiO₂/H₂O₂/UV processes. The highest color removal efficiency is obtained at pH 3 in both processes. Color removal efficiencies show a decreasing trend with the increase in

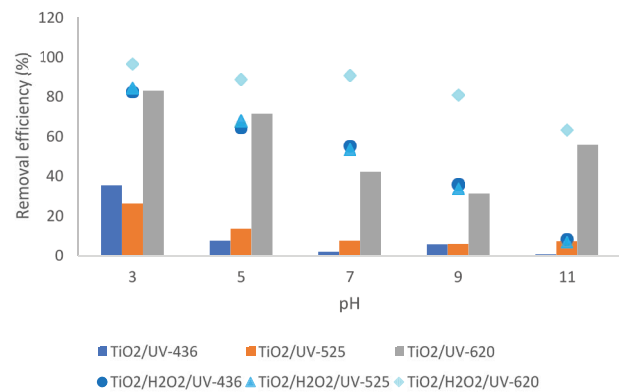


Fig. 3. Effect of pH on the color removal ($C_{0,436 \text{ nm}} = 32 \text{ m}^{-1}$; $C_{0,525 \text{ nm}} = 36.2 \text{ m}^{-1}$; $C_{0,620 \text{ nm}} = 30.1 \text{ m}^{-1}$; n-TiO₂ dose = 100 mg/L; H₂O₂ dose = 1,000 mg/L; $t = 30 \text{ min}$).

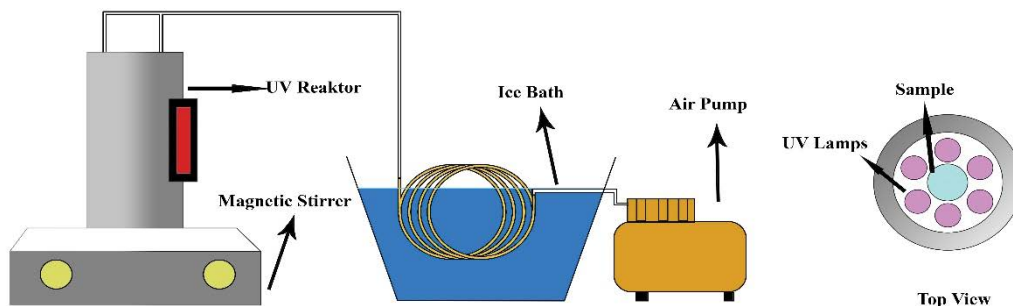


Fig. 2. Experimental set-up.

pH. Therefore, the optimum pH was determined as 3 in both processes. Although the pH decrease in both processes increases the color removal efficiency, it has also been observed that the n-TiO₂/H₂O₂/UV process has better removal performance. Unlike the n-TiO₂/UV process, in the n-TiO₂/H₂O₂/UV process, with the addition of H₂O₂ formation of OH[•] radicals multiply and correlatively the oxidation capacity increases [25].

In the n-TiO₂/UV process, the color removal efficiencies for 436, 525 and 620 nm at pH 3 are 35.31%, 26.24% and 83.06%, respectively, while in the n-TiO₂/H₂O₂/UV process, the removal efficiencies at the same wavelengths are 82.5%, 84.53% and 96.35%, respectively. When the situation at pH 3 is examined for both processes, the difference in color removal efficiency is approximately 48% at 436 nm, approximately 58% at 525 nm and approximately 13% at 620 nm. In the literature, n-TiO₂/UV and n-TiO₂/H₂O₂/UV processes have been used for color removal using different azo dyes, and dye removal has been achieved with the highest efficiency in acidic conditions. In these studies, high color removal was found between pH 2.5 and 4.5 [26–28].

In the literature, TiO₂ catalysed photocatalytic studies are examined, and the dye removal efficiency is generally high at acidic pHs. As a result of the positive charge of the surface of TiO₂ in an acidic solution, it absorbs negatively charged organic compounds. Therefore, high photocatalytic activity is expected to occur [29]. The zeta potential of the catalysts varies depending on the pH of the solution [30]. The zeta potential change of n-TiO₂ used in the experiments depending on pH is given in Table 3. Table 3 shows, while the zeta potential of TiO₂ is high at low pH, it decreases as the pH rises and even becomes negative at high pH. The zero point of charge for n-TiO₂ is at an almost neutral pH. The TiO₂ surface is positively charged below approximately neutral pH; however, the particle surface becomes negatively charged above neutral pH [31].

3.2. Effect of catalyst dose on color removal

The catalyst dose is one of the essential parameters in photocatalytic processes [32]. Since TiO₂ is a low cost, non-toxic, easy to procure material, has been preferred catalyst used in heterogeneous photocatalytic processes [33]. The particle size, shape, and crystal structure of TiO₂ are suitable for changes at the desired level in order to increase the degradation rate and amount of the substances desired to be purified in photocatalytic processes [34].

In this study, studies have been taken place in the range of 25–150 mg/L TiO₂ dose to examine the effect of catalyst

Table 3
Zeta potentials of n-TiO₂ depending on the pH

pH	Zeta potential (mV)
2	2.61
3	1.388
5	0.691
7	0.218
9	−3.33

dose on color removal in synthetic wastewater containing dyestuff (DB 86, RY 145, AV 90). The effect of catalyst dose on color removal is seen in Fig. 4. In the n-TiO₂/UV process, color removals have been obtained as 13.44% at 436 nm, 16.02% at 525 nm and 50.17% at 620 nm at a dose of 25 mg/L n-TiO₂. In this process, color removal efficiencies at 150 mg/L n-TiO₂ doses were 43.44% at 436 nm, 32.87% at 525 nm and 83.39% at 620 nm. In the n-TiO₂/H₂O₂/UV process, color removals have been achieved as 51.56%, 65.75% and 79.40% at 436, 525 and 620 nm for 25 mg/L n-TiO₂ dose, meanwhile at 436, 525 and 620 nm for 150 mg/L n-TiO₂ dose removal efficiencies were 89.38%, 92.54% and 98.01% respectively. While the color removal increases with the increase of the catalyst dose, there is no high increase in color removal after 125 mg/L n-TiO₂ dose, therefore 125 mg/L has been determined as optimum n-TiO₂ dose. Similarly, in the study of Amoli et al., the photocatalytic degradation of synthetic wastewater containing Direct Blue 199 and Basic Yellow 28 was investigated, and with the increase of the TiO₂ dose, the removal efficiency of the dyestuffs increased up to a certain point, but then a considerable increase was not observed [35]. The reduction in removal efficiency at a very high catalyst dose may be due to the low transmittance of the reaction mixture and light scattering. Another reason is thought to be the deactivation of reactive molecules, which may occur due to collision with non-reactive molecules [32]. As the amount of catalyst used increases, the contact surface area increases, and the treatment efficiency generally increases because it creates more catalyst active sites for the substances intended to degrade [36,37].

3.3. Effect of H₂O₂ dose on color removal

Experiments have been performed at pH 3, 125 mg/L n-TiO₂ dose, 30 min reaction time and 36 W light intensity while examining the effect of H₂O₂ dose on color removal in n-TiO₂/H₂O₂/UV process. It is known that the OH[•] radicals formed when H₂O₂ is added cause the removal of dyestuffs, and the effect of H₂O₂ on color removal efficiencies has been investigated between 250–1,500 mg/L of H₂O₂ doses and has been shown in Fig. 5. While color removal is fewer at 250 mg/L H₂O₂ dose, the increase in color removal is slightly

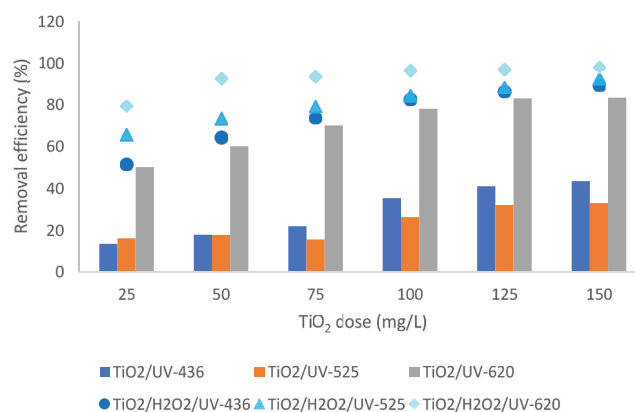


Fig. 4. Effect of n-TiO₂ dose on the color removal ($C_{0,436\text{ nm}} = 32\text{ m}^{-1}$; $C_{0,525\text{ nm}} = 36.2\text{ m}^{-1}$; $C_{0,620\text{ nm}} = 30.1\text{ m}^{-1}$; pH = 3; H₂O₂ dose = 1,000 mg/L; $t = 30\text{ min}$).

raised up after 750 mg/L H_2O_2 dose, though H_2O_2 dose does not affect color removal at 620 nm, only around 2%. On the contrary, the effect of H_2O_2 dose on color removal at 436 nm and 525 nm is quite evident. Color removals at 750 mg/L have been achieved as 85.63% at 436 nm, 86.74% at 525 nm and 97.34% at 620 nm.

Due to the rise in H_2O_2 dose, Chu also stated that treatment efficiencies are affected by the autodegradation of H_2O_2 and the reaction of hydroxyl radicals with excessive H_2O_2 concentrations [38]. This can be explained as the reason why the increase in removal efficiency slows down after 750 mg/L.

3.4. Effect of reaction time on color removal

In order to investigate the effect of reaction time on color removal from synthetic wastewater samples containing dyes in n-TiO₂/UV and n-TiO₂/H₂O₂/UV processes, pH 3, 125 mg/L n-TiO₂ dose and 36 Watt light intensity conditions have been studied. Furthermore, 750 mg/L H_2O_2 was used for the n-TiO₂/H₂O₂/UV process. Fig. 6 shows the effect of reaction time on color removal in both processes. According to

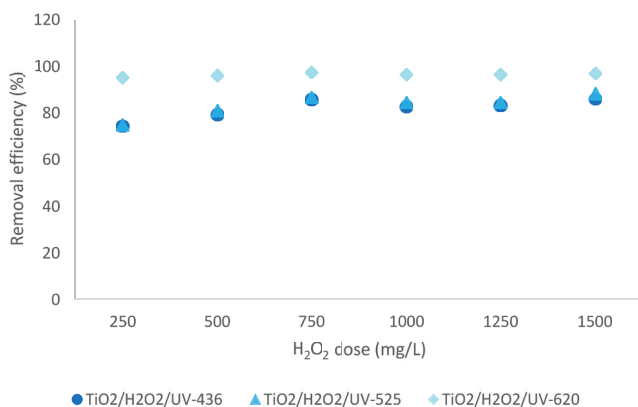


Fig. 5 Effect of H_2O_2 dose on the color removal ($C_{0,436\text{ nm}} = 32\text{ m}^{-1}$; $C_{0,525\text{ nm}} = 36.2\text{ m}^{-1}$; $C_{0,620\text{ nm}} = 30.1\text{ m}^{-1}$; pH = 3; n-TiO₂ dose = 125 mg/L; $t = 30\text{ min}$).

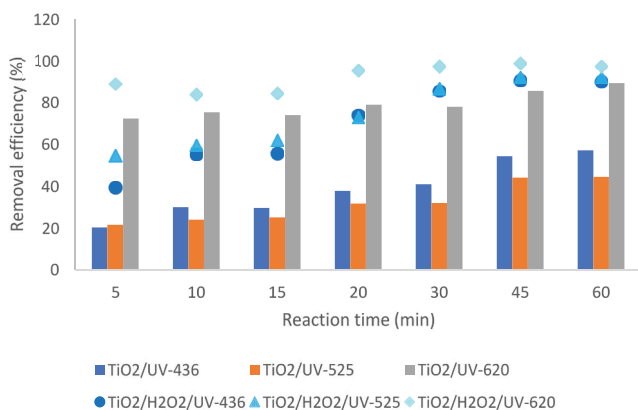


Fig. 6 Effect of reaction time on the color removal ($C_{0,436\text{ nm}} = 32\text{ m}^{-1}$; $C_{0,525\text{ nm}} = 36.2\text{ m}^{-1}$; $C_{0,620\text{ nm}} = 30.1\text{ m}^{-1}$; pH = 3; n-TiO₂ dose = 125 mg/L; H_2O_2 dose = 750 mg/L).

Fig. 6, color removal increased until the 45th min and then the increase slowed down for both processes. In order to examine the effect of reaction time on dye removal, removal efficiencies have been measured between 5 min and 60 min. Color removal efficiencies for n-TiO₂/UV at 5 min were found to be 20.31%, 21.55% and 72.43% at wavelengths of 436, 525 and 620 nm, respectively. Moreover, the removal efficiencies at 5 min in the n-TiO₂/H₂O₂/UV process are 39.38%, 54.70% and 89.04% at 436, 525 and 620 nm wavelengths, respectively. From the 5th to the 60th min, it was observed that the color removal efficiencies increased, but the difference between the color removal efficiencies between 45 min and 60 min were less than about 1% at each wavelength in both processes. Therefore, the 45th min has been determined as the optimum reaction time for both processes.

When both processes are compared, the color removal efficiencies of the n-TiO₂/H₂O₂/UV process are higher than the results obtained with the n-TiO₂/UV process. n-TiO₂/H₂O₂/UV process is a considerable method for the treatment of dyestuffs from wastewater. The highest removal efficiency has been obtained at 620 nm wavelength in the n-TiO₂/UV process. Color removal efficiencies at 436, 525 and 620 nm for 45th min have been obtained as 54.38%, 44.20% and 85.72%, respectively. In addition, the color removal efficiencies that have been achieved at the same wavelengths in the n-TiO₂/H₂O₂/UV process are 90.63%, 92.27% and 99%, respectively, for 45th min. Among the studies examined for dye removal with the n-TiO₂/UV process; more than 90% of the color was removed in 120 min for Methylene Blue [39], in 40 min for Reactive Orange 4 [40], and 45 min for textile wastewater containing Reactive Red 239 [41]. Also, the color removal of Red-147 dye was achieved in 60 min with the n-TiO₂/H₂O₂/UV method, and the treatment efficiency was above 95% [42].

In Table 4, other advanced oxidation studies in the literature examining DB 86, AV 90 and RY 145 dyes alone or in combination with other dyes are examined. Optimum reaction times have been determined as 45 min in both methods used in the study. It has been observed that in Table 4, the processes using catalytic UV, the optimum reaction times vary due to the types of used catalysts to remove only a single dye with a period of 15–240 min. In this study, color removal of synthetic wastewater consisting of a mixture of 3 dyes has been achieved in 45 min.

In addition to the photocatalytic method, other advanced oxidation processes studies involving these dyes are also listed in Table 4. In the advanced oxidation methods presented in Table 4, color removal efficiencies were investigated in optimum parameters such as pH, oxidant doses, catalyst dose, initial concentration, and reaction time. Current study is critical because it is the only study which used n-TiO₂/UV and n-TiO₂/H₂O₂/UV methods with three different color tones as blue, yellow, red are studied together. In the literature, there is no color removal study with advanced oxidation processes containing Direct Blue 86, Reactive Yellow 145 and Acid Violet 90 dyestuffs as a mixture.

3.5. Kinetic analysis of color removal

Kinetic examinations have been used experimental data to obtain reaction rate constants and regression coefficients.

Table 4

Comparison of the current study to previously reported with Direct Blue, Reactive Yellow and Acid Violet dyes

Dyestuff	Treatment method	Optimum conditions	Removal efficiency	References
Direct Blue 86	O ₃ UV/O ₃	pH 11 100 ppm (initial dye conc.) 35 min	O ₃ : 99% UV/O ₃ : 94%	[43]
Direct Blue 86	Photocatalytic treatment with phosphoric acid-based geopolymer	20 W Xenon Lamp 0.3 g of PAG 1:1 50 mg/L (initial dye conc.) 15 min	87.31% with P/Al ratio 1:1	[44]
Direct Blue 129	Photocatalytic treatment with green synthesis of zinc oxide nanoparticles	20 mg/L (initial dye conc.) 0.05 g ZnO visible light irradiation <i>t</i> = 105 min	95%	[45]
Reactive Yellow 14	Photocatalytic oxidation UV/TiO ₂	5 × 10 ⁻⁴ mol/L (initial conc.) TiO ₂ -P25 4 g/L 32 W H ₂ O ₂ : 15 mM 20 min	90.4%	[46]
Reactive Yellow 14	UV/TiO ₂ , UV/H ₂ O ₂ UV/H ₂ O ₂ /Fe ²⁺ Fe ²⁺ / H ₂ O ₂	TiO ₂ = 4 g/L [H ₂ O ₂] = 10 mmol/l [Fe ²⁺] = 0.05 mmol/l Irradiation time = 40 min pH: 3	UV/TiO ₂ için: 91.3%	[47]
Reactive Yellow 86	UV/TiO ₂ UV/TiO ₂ / BiOCl UV/TiO ₂ / BiOCl/La ₂ O ₃	120 min	TiO ₂ : 67% TiO ₂ /BiOCl: 79% TiO ₂ /BiOCl/La ₂ O ₃ : 92%	[48]
Reactive Yellow 125	Photocatalysis natural zeolite modified with nitrogen-doped TiO ₂ photocatalysis (Z-TiO ₂ -N) UV-Vis irradiation	Photocatalyst = 1 g/L Dye = 25 mg/L pH: 3 240 min	UV/Z-TiO ₂ -N: 91.6% VIS/Z-TiO ₂ -N: 94.8%	[49]
Reactive Yellow dye	Photocatalysis	10 mg/L (initial conc.) pH: 3 TiO ₂ = 10 mg/L 60 min	UV-TiO ₂ : 87.5%	[50]
Reactive Yellow 17	Photocatalysis	10 mg/L (initial conc.) [Fe ²⁺] = 0.05 mM [S ₂ O ₈ ²⁻] = 1 mM pH: 3 20 min	S ₂ O ₈ ²⁻ /Fe ²⁺ : 63.3% O ₈ ²⁻ /UV: 81.0% S ₂ O ₈ ²⁻ /Fe ²⁺ /UV: 95.4%	[51]
Reactive Yellow 145 Reactive Orange 122 Reactive Black 5	UV/H ₂ O ₂ Fenton photo-Fenton	50 mg/L (initial conc.) [H ₂ O ₂] = 40 mg/L [Fe] = 1 mg/L pH = 3 and 4	98% for color 60 min 68% for aromatics 60 min 93% for COD 180 min	[52]
Reactive Red 194 Reactive Yellow 145	O ₃ H ₂ O ₂ /UV-C	Optimum pH = 11 O ₃ neutral pH (H ₂ O ₂ /UV-C) [H ₂ O ₂] = 40 mM <i>t</i> = 30 min	O ₃ process TOC = 43.8% for RY145 35.5% for RR 194 H ₂ O ₂ /UV-C process 26.1% for RY145 25.7% for RR 194	[53]
Reactive Red 120 Reactive Black 5 Reactive Yellow 84	UV/H ₂ O ₂	100 mg/L (initial conc.) [H ₂ O ₂] = 24.5 mmol/l Irradiation time = 60 min	Decolorization 99.6%. Mineralization rates were 58.5% for RY84 52.9% for RR84 81.6% for RB5	[54]
Reactive Black 5 Direct Red 28 Direct Yellow 12	UV/H ₂ O ₂ /Fe ²⁺	100 mg/L (initial conc.) 16 W low-pressure mercury vapour lamp <i>t</i> = 60 min (UV, UV/H ₂ O ₂) <i>t</i> = 5 min (UV/H ₂ O ₂ /Fe ²⁺)	UV/H ₂ O ₂ 99% for RB5 98% for DY12 in 60 min 70% for DR28 in 120 min UV/H ₂ O ₂ /Fe ²⁺ 98% for RB5 88% for DY12 85% for DR28 in 5 min	[55]

(Continued)

Table 4

Dyestuff	Treatment method	Optimum conditions	Removal efficiency	References
Acid Violet 7	Ultrasound-H ₂ O ₂	20 mg/L (initial conc.) 40 kHz US 100 W power dissipation pH: 3 [H ₂ O ₂] = 25 ppm <i>t</i> = 180 min	77%	[56]
Acid Violet 7	Ultrasound + Fenton	20 mg/L (initial conc.) 40 kHz US 100 W power dissipation pH: 3 [Fe ²⁺] = 10 ppm [H ₂ O ₂] = 50 ppm <i>t</i> = 2 min	99.9%	[56]
Acid Violet 12	Electrocatalytic oxidation	140 mg/L (initial conc.) 1 A/dm ² current density pH: 7 Electrolyte concentration of 0.58 g/L <i>t</i> = 30 min	78% COD reduction	[57]
Acid Violet 7	Photocatalytic	0.5 mM (initial conc.) Catalyst [ZnO] = 2 g/L pH: 9 Airflow rate = 8.1 mL/s <i>t</i> = 60 min	99%	[58]
Acid Violet 7	Photo-Fenton	0.5 mM (initial conc.) Fe ³⁺ loaded Al ₂ O ₃ UV-A light Catalyst loading of 1 g/L [H ₂ O ₂] = 10 mmol Airflow rate = 8.1 mL/s pH: 3 ± 0.1 <i>I</i> ₀ = 1.381 × 10 ⁻⁶ Einstein L ⁻¹ s ⁻¹ <i>t</i> = 90 min	99%	[59]

R^2 values obtained by kinetic studies show the relationship between reaction rate and time. For this reason, when trying kinetic models, the model with the highest R^2 is preferred. The reaction rate constants obtained are used to determine the concentration of the substances at any time t in the removal of wastewater with this characteristic by the determined method.

In the literature, the reaction rate constant (k) and regression coefficient (R^2) of 4 different kinetic models can be calculated using the reaction time. In this study, k and R^2 values have been calculated based on the reaction time by using four different kinetic models as the first-order, second-order, pseudo-first-order and pseudo-second-order in n-TiO₂/UV and n-TiO₂/H₂O₂/UV processes.

The kinetic equations are given below.

$$\text{First-order kinetic: } \ln \frac{C_0}{C} = kt \quad (14)$$

$$\text{Second-order kinetic: } \frac{1}{C} - \frac{1}{C_0} = kt \quad (15)$$

$$\text{Pseudo-first-order kinetic: } \ln \frac{C_0}{C} = k't \quad (16)$$

$$\text{Pseudo-second-order kinetic: } \frac{t}{C} = \frac{1}{k_2 C_{\max}^2} + \frac{1}{C_{\max}} t \quad (17)$$

where C_0 : the initial concentration of dyestuff; C : the dyestuff concentration in t time; C_{\max} : the maximum percentage removal efficiency; k : reaction rate constants.

k and R^2 values calculated with 4 different kinetic models for n-TiO₂/UV and n-TiO₂/H₂O₂/UV processes are given in Tables 5 and 6. According to Table 5, pseudo-second-order kinetics for 436, 525 and 620 nm have been determined as the most suitable models because of the highest R^2 in the n-TiO₂/UV process. The k and R^2 values have been calculated as 0.0026 mg/L min and 0.984 at 436 nm, 0.004 L/mg min and 0.986 at 525 nm, and 0.0128 and 0.996 at 620 nm, respectively. Pseudo-second-order kinetic has been found to be the most significant model for the n-TiO₂/H₂O₂/UV process at all wavelengths with the highest R^2 as seen in Table 6.

Table 5
R² and k values as per kinetic models in TiO₂/UV process

Kinetic model	R ²			k		
	436 nm	525 nm	620 nm	436 nm	525 nm	620 nm
First-order ^a	0.948	0.982	0.909	0.0167	0.0122	0.0467
Second-order ^b	0.979	0.938	0.947	0.0008	0.0004	0.0048
Pseudo-first-order ^c	0.926	0.860	0.854	0.0742	0.0626	0.0817
Pseudo-second-order ^b	0.984	0.986	0.996	0.0026	0.0040	0.0128

^ak unit: 1/min; ^bk unit: L/mg min, ^ck unit: mg/L min.

Table 6
R² and k values as per kinetic models in TiO₂/H₂O₂/UV

Kinetic model	R ²			k		
	436 nm	525 nm	620 nm	436 nm	525 nm	620 nm
First-order ^a	0.947	0.823	0.852	0.0485	0.1108	0.0897
Second-order ^b	0.966	0.966	0.933	0.0059	0.0059	0.0553
Pseudo-first-order ^c	0.937	0.866	0.891	0.2142	0.1769	0.1407
Pseudo-second-order ^b	0.988	0.991	0.996	0.0030	0.0037	0.0186

^ak unit: 1/min; ^bk unit: L/mg min, ^ck unit: mg/L min.

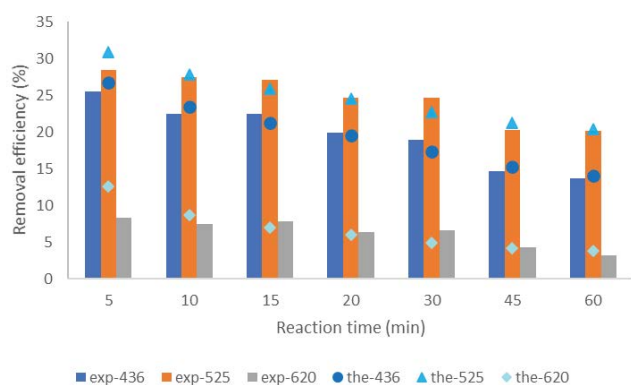


Fig. 7. Theoretical and experimental color removal as second-order kinetics for TiO₂/UV process.

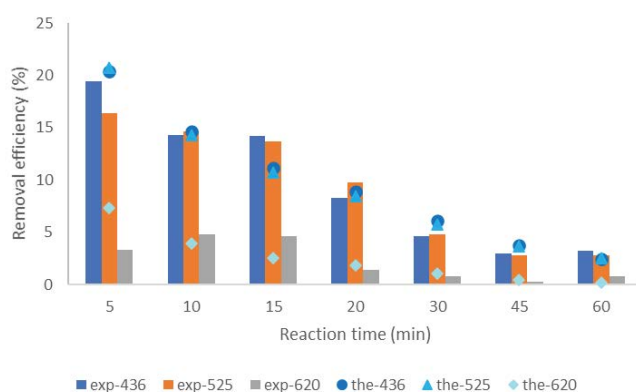


Fig. 8. Theoretical and experimental color removal as second-order kinetics for TiO₂/H₂O₂/UV process.

The k values for 436, 525 and 620 nm wavelengths have been calculated as 0.0030, 0.0037 and 0.0186 L/mg min, respectively, while the R² values were calculated as 0.988, 0.991 and 0.996, respectively.

The theoretical color removals have been calculated using the reaction rate constants obtained from the pseudo-second-order model for n-TiO₂/UV and n-TiO₂/H₂O₂/UV processes. Thereafter, the comparison of the theoretical and the experimental results have shown in Fig. 7 for the n-TiO₂/UV process and in Fig. 8 for the n-TiO₂/H₂O₂/UV process. Depending on the results obtained from the calculations at wavelengths of 436 nm, 525 nm and 620 nm, significant results were found that matched the experimental results with the pseudo-second-order model. Figs. 7 and 8 show that except for the theoretical results, especially at 620 and up to 5 min, the theoretical results at the other wavelengths coincide exactly with the experimental results for both processes.

4. Conclusions

Two different processes, n-TiO₂/UV and n-TiO₂/H₂O₂/UV have been researched on the synthetic wastewater containing dyestuffs (DB 86, RY 145 and AV 90) treatment. The effect of pH, n-TiO₂ dose, H₂O₂ dose, and reaction times as operational parameters have been investigated. pH 3, n-TiO₂ 125 mg/L and 45 min have been determined as optimum conditions for both processes, and also 750 mg/L H₂O₂ have been found for n-TiO₂/H₂O₂/UV. Under optimum conditions, decolorization efficiencies were 54.38% at 436 nm, 44.20% at 525 nm and 85.72% at 620 nm for the n-TiO₂/UV process, while 90.63% at 436 nm, 92.27% at 525 nm and 99% at 620 nm color removal efficiencies have been achieved in the n-TiO₂/H₂O₂/UV process. While dye removal

was carried out under the same conditions in both processes, when the color removals are examined, the n-TiO₂/H₂O₂/UV process performs much more efficient treatment than the n-TiO₂/UV process. As can be understood from here, it is revealed that hydrogen peroxide advances the treatment efficiency. The results revealed that the pseudo-second-order kinetic model was in concordance with the experimental results for both processes.

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