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# Fabrication removal method of gallic acid and Pb<sup>2+</sup> ions in real samples by a new adsorbent chitosan grafted with a mixture of itaconic acid-methacrylamide

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# ABSTRACT

In this study, chitosan (CTS) with medium molecular weight, using a mixture of monomers, itaconic acid (IA) and methacrylamide (MAm) were grafted in the presence of radical initiator, 4,4'-azobis(4-cyanovaleric acid) (ACV) and ethylene glycol dimethacrylate (EGDMA) as cross-linkage. The best graft yield was obtained by gravimetric method (EGDMA = 2 mL, ACV = 0.01 g, IA = 0.01 g, MAm = 0.09 g, *t* = 45 min, CTS = 0.2 g. The characterization was performed by X-ray diffraction, Fourier-transform infrared spectroscopy, thermogravimetric analysis and scanning electron microscopy methods. The composite was used to adsorb Pb<sup>2+</sup> and gallic acid compounds and important adsorption parameters including pH, adsorption time, adsorbate concentration and adsorbent amount were optimized. The best adsorption conditions for Pb<sup>2+</sup> (time = 75 min, pH = 4.0, adsorbent = 0.05 g, adsorbate concentration = 150 mg/L and  $q_{max}$  = 276 mg/g) and for gallic acid (time = 60 min, pH = 3.0, adsorbent = 0.02 g, adsorbate concentration = 300 mg/L and  $q_{max}$  = 576 mg/g) were recorded. The adsorption data's were fitted by Langmuir isotherm and pseudo-second-order kinetic model, as well. Thermodynamic studies showed that the adsorption process is endothermic and spontaneous and due to the increase in entropy confirms the proper interaction of the adsorbent and adsorbate. Meantime, due to the high adsorption capacity of the adsorbent and good efficiency after several times of adsorption–desorption, it can be used as an industrial adsorbent.

Keywords: Chitosan; Itaconic acid; Methacrylamide; Adsorption

# 1. Introduction

Chitosan is obtained by N-deacetylation of chitin with strong alkali [1]. It is a natural adsorbent for metal ions such as zinc, nickel, copper, chromium and lead. This could be for the following reasons: Large group of hydroxyl in chitosan, existence of primary amino groups that are considered as adsorption sites, and the flexibility of a polymer chain structure that works as a convenient provider for the formation of complexes with metal ions [2]. All toxic metals can be absorbed by living organisms. Large amounts of these metals accumulate in the human body through the food chain. Due to the digestion of these metals in the body, they can cause some health problems and disorders in humans and animals [3]. Pb is a toxic metal that accumulates and enters the vital organs of humans and animals. It enters the body through air, water and food. According to the World Health Organization's (WHO) standards, its maximum limit in drinking water is 0.05 mg/L and in wastewater is 0.5 mg/L [4]. The presence of small amounts of toxic metal ions in liquid solutions is challenging. In addition, the method of separating metal ions from wastewater is a complex process by pH, competing ions, and the presence of organic substances. Various methods such as filtration, ion exchange, precipitation and adsorption have been used to remove toxic species from liquid solutions [5-8]. Chemical modification of natural polymers is a good way to produce new, high-performance biomaterials. Various articles have been published on graft copolymerization of acrylonitrile, dimethylaminoethyl methacrylate, acrylamide, and vinylpyrrolidone onto chitosan using ceric ammonium nitrate (CAN) as an initiator [9–13]. On the other hand,  $\gamma$ -Ray and photoinduced graft-copolymerization of acrylamide, methyl methacrylate, 2-hydroxyethyl methacrylate, styrene and acrylonitrile on chitosan has also been reported [10-14]. Chitosan is grafted using potassium persulfate as a redox initiator of methyl methacrylate, methyl acrylate, butyl acrylate, 2-acrylamido-2-methylpropane sulfonic acid, and acrylic acid [11,15–19]. In recent years, important researches have been reported on chitosan-clay composites and grafting with monomers for the removal of pollutants [20-30]. So far, no report has been published about chitosan grafting with itaconic acid-methacrylamide (IA-MAm) monomer mixture in the presence of 4,4'-azobis(4-cyanovaleric acid) (ACV), and in the published reports, common radical initiators benzoyl peroxide (BPO) and azobisisobutyronitrile (AIBN) have been used.

In my recent study, chitosan was grafted with a mixture of itaconic acid-methacrylamide-bentonite monomers [31].

In this research, chitosan was grafted with IA-MAm monomer mixture in the presence of ethylene glycol dimethacrylate (EGDMA) and ACV. After characterization by various methods, it was used as an adsorbent of gallic acid (GA) (as a model of a cyclic compound) and Pb<sup>2+</sup>.

#### 2. Experimental

# 2.1. Materials

Medium molecular weight CTS (Sigma-Aldrich) and all of chemicals such as IA and MAm as monomers, EGDMA as cross-linker, ACV as radical initiator, GA, Lead acetate and the required solvents were supplied by Merck and were analytical grade and highest purity. Also doubly distilled water has been used in all experiments.

# 2.2. Instruments

The structure of the composite and the amount of adsorption were investigated by Fourier-transform infrared spectroscopy (FTIR; Bruker Equinox Spectrophotometer with KBr disks), X-ray diffraction (XRD; PW 1730), scanning electron microscopy (SEM; Philips CM120), thermogravimetric analysis (TGA; Church Stretton), atomic absorption spectroscopy (AAS; GBC 932 AA) and UV-Spectrophotometer (Shimadzu 1208).

# 2.3. Synthesis of composite

0.2 g of chitosan was dissolved in 35 mL of 5 wt.% acetic acid. Then, ACV was dissolved in 5 mL of distilled water and added to the solution. Two monomers of itaconic acid (IA), methacrylamide (MAm) are also added. Finally, ethylene glycol dimethacrylate (EGDMA) is added to the solution as a cross-linker. The contents were placed in a water bath present at 90°C. Finally, the purification was performed with Whatman Filter Paper and washed with fresh methanol and dried at 50°C. By changing the percentage of monomers, initiator concentration, cross-linker concentration and time, optimal conditions were obtained by gravimetrically [32,33].

# 2.4. Adsorption procedure

Investigation of the adsorption of 50 mg/L of adsorbates with 0.01 g of adsorbent in the total volume of 20 mL, inside Erlenmeyer flask (250 mL) with constant shaker speed (150 rpm) and at a constant temperature of 298 K as a function pH, adsorbate concentration, the amount of adsorbent and time were accomplished. The solutions were then filtered and centrifuged at 3,000 rpm. Residual solution concentration in GA was measured using a UV-Spectrophotometer at  $\lambda_{max} = 265$  nm using a calibration curve and the residual concentration of Pb<sup>2+</sup> ions in solution was measured using an AAS instrument at  $\lambda_{max} = 217$  nm with the help of a calibration curve. Calibration curves for GA and Pb<sup>2+</sup> were plotted in the concentration range of 5–50 mg/L.

Finally, the adsorbent efficiency was calculated using the calibration curve and the following equation:

$$q = \frac{\left(C_0 - C_e\right)}{m} V \tag{1}$$

where  $C_0$  and  $C_e$  are the initial and equilibrium concentrations of adsorbate (mg/L), *V* total volume of solution (L), *m* absorbent weight (g) and *q* is the amount of the adsorbed at the specified time (mg/g) [34].

#### 3. Results and discussion

#### 3.1. Adsorbent synthesis and characterization

According to the preparation method, in a constant amount of monomers (0.1 g) and a change in the weight percentage of two monomers, the best conditions were observed in 0.09 g MAm–0.01 g IA.

The amount of initiator was performed from 0.005 to 0.075 g and the maximum grafting yield was observed in 0.01 g of initiator.

By changing the amount of cross-linker in the range of 1–4 mL, the best grafting yield was obtained in 2 mL. The synthesis time varied from 30 to 120 min and the optimum condition was observed in 45 min. Under optimal conditions of synthesis, the characterization of the composite was studied.

Comparison of chitosan and grafted chitosan FTIR spectra showed the following spectral changes: The main characteristic peaks of pure chitosan are at 3,436 cm<sup>-1</sup> (O–H stretch), 2,879 cm<sup>-1</sup> (C–H stretch), 1,605 cm<sup>-1</sup> (N–H bond), 1,379 cm<sup>-1</sup> (C–N stretch), 1,089 cm<sup>-1</sup> (C–O stretch) and 1,155 cm<sup>-1</sup> (bridge O stretch). In addition to chitosan peaks, 2 new peaks at 3,106 cm<sup>-1</sup> (N–H stretch of amide) and 1,725 cm<sup>-1</sup> (C=O stretch of acid) confirm the correlation of 2 monomers IA, MAm to chitosan (Figs. 1 and 2).

TGA results show that after grafting, degradation begins at a lower temperature. As shown in Figs. 3 and 4, chitosan



Fig. 1. FTIR spectrum of chitosan.



Fig. 2. FTIR spectrum of IA/MAm grafted onto chitosan.



Fig. 3. TGA of chitosan.



Fig. 4. TGA of IA/MAm mixture grafted onto chitosan.

degradation temperature starts from 300°C and grafted chitosan degradation temperature starts from about 220°C. This is due to the reduction in crystallinity.

The XRD pattern in chitosan shows 2 peaks at  $2\theta = 10^{\circ}$  and  $20^{\circ}$  indicates high crystallinity. This result is consistent with previous reports [35,36]. The corresponding peaks disappeared after grafting and a weak peak appeared at  $2\theta = 15^{\circ}$  which indicates a decrease in crystallinity and an increase in amorphous structure (Figs. 5 and 6).

SEM image show that the chitosan surface is smooth and dense. But the surface of grafted chitosan with IA and MAm is rough and porous. This image confirms the grafting (Figs. 7 and 8).

# 3.2. Effect of pH

Under constant conditions of adsorption time (60 min), adsorbent (0.01 g), shaker speed (150 rpm), adsorbate concentration (50 mg/L), total volume (20 mL), temperature 298 K, pH changes from 2.0–10.0, were investigated. The best pH for adsorption of Pb<sup>2+</sup> ions was recorded at pH = 4.0 and for GA at pH = 3.0. At pH above 6.0, lead precipitates in the form of hydroxide [37].

In the case of GA, due to the presence of functional groups in chitosan and monomers grafted to it, high adsorption was observed in various pH. On the other hand, at alkaline pH, gallic acid is formed as galleate, but at acidic pH, it is possible to bond with the amide group. The best adsorption was recorded at pH = 3.0 (Figs. 9 and 10).

#### 3.3. Effect of time

The effect of contact time on adsorption was studied from 30 to 120 min at the fixation condition of other variables (pH = 4.0 for Pb<sup>2+</sup> and pH = 3.0 for GA, solution volume = 20 mL, T = 298 K, adsorbent = 0.01 g, adsorbate concentration = 50 mg/L and shaking rate = 150 rpm). As shown in Figs. 11 and 12, the best adsorption time for GA, 60 min and for Pb<sup>2+</sup>, 75 min were recorded.

# 3.4. Effect of adsorbate concentration and adsorbent amount

By changing the concentration of adsorbates, from 50 to 400 mg/L, under optimized conditions, experiments were performed. The maximum adsorption at a concentration of 300 mg/L for GA and 150 mg/L for Pb<sup>2+</sup> were observed. By increasing the amount of adsorbent, the highest adsorption was recorded with 0.02 g for gallic acid and 0.05 g for Pb<sup>2+</sup>. In the mentioned values, the removal percentage in both cases reached 98%. Figs. 13–16.



Fig. 5. XRD pattern of chitosan.



Fig. 6. XRD pattern of IA/MAm mixture grafted onto chitosan.



Fig. 7. SEM micrograph of chitosan.

The value of  $q_{\rm max}$  for GA was 576 and 276 mg/g for Pb<sup>2+</sup>. Due to the grafting of monomers and the presence of functional groups, the adsorption capacity has increased significantly.

# 3.5. Adsorption isotherm study

The experimental data were determined according to the commonly used adsorption isotherms, like Langmuir and Freundlich. Langmuir isotherm is based on the hypothesis



Fig. 8. SEM micrograph of IA/MAm mixture grafted onto chitosan.

that point of valence exists on the surface of the adsorbent and that each of these sites is authoritative of one molecule. The Langmuir equation is given as follows [38].

$$\frac{C_e}{q_e} = \frac{1}{q_m b} + \frac{C_e}{q_m}$$
(2)

where  $C_e$  is the equilibrium concentration (mg/L),  $q_e$  is the amount adsorbed at equilibrium (mg/g).  $q_m$  and b are



Fig. 9. Effect of pH on the adsorption of GA.



Fig. 10. Effect of pH on the adsorption of Pb2+.



Fig. 11. Effect of contact time on the adsorption of GA.



Fig. 12. Effect of contact time on the adsorption of Pb<sup>2+</sup>.

Langmuir constants depend to adsorption efficiency and energy of adsorption. The linear plots of  $C_e/q_e$  vs.  $C_e$  identify the applicability of the Langmuir isotherms. The values of  $q_m$  and b can be determined from slope and intercepts of the curve.

The Freundlich isotherm is given by the following equation [39].

$$\ln q_e = \ln K_f + \frac{1}{n} \ln C_e \tag{3}$$



Fig. 13. Effect of GA concentration (mg/L).



C(mg/L)

Fig. 14. Effect of Pb<sup>2+</sup> concentration (mg/L).



Fig. 15. Effect of adsorbent amount on the adsorption of GA.



Fig. 16. Effect of adsorbent amount on the adsorption of Pb2+.

where  $K_f$  and n are constants the factors affecting the adsorption capacity and strength adsorption, respectively. As can be seen from Figs. 17–20 and Table 2, the best isotherm to justify the adsorption of both adsorbates is the Langmuir isotherm. The correlation coefficient in both cases is higher than  $R^2 = 0.99$ . This indicates the homogeneity of the adsorption process on the adsorbent surface [32].

# 3.6. Adsorption kinetics

Two well-known and important pseudo-first-order, pseudo-second-order kinetic models were used to investigate

Table 1 Percentage of weight loss in the range of 100°C to 700°C

Temperature (°C)	Chitosan	Composite
100	5.11	0.2
200	5.04	1.06
300	17.54	8.67
400	54.58	51.53
500	61.79	94.01
600	64.78	94.38
700	66.77	94.59



Fig. 17. Langmuir isotherm for GA (T = 298 K).



Fig. 18. Freundlich isotherm for GA (T = 298 K).



Fig. 19. Langmuir isotherm for  $Pb^{2+}$  (*T* = 298 K).

the adsorption kinetics. Fixed conditions in experiments: 0.01 g of adsorbent, total volume 20 mL, adsorbate concentration 50 mg/L, zero shaker speed, optimal pH 3.0 and 4.0 for lead ions and gallic acid was performed from 5 to 45 min.

Pseudo-first-order: 
$$\ln(q_e - q_t) = \ln q_e - K_1 t$$
 (4)

Pseudo-second-order: 
$$\frac{t}{q_t} = \frac{1}{K_2 q_e^2} + \frac{t}{q_e}$$
 (5)



Fig. 20. Freundlich isotherm for  $Pb^{2+}$  (*T* = 298 K).

Table 2 Adsorption isotherm constants

Langmuir constants	$K_a$ (L/mg)	$q_m (\mathrm{mg/g})$	$R^2$
Pb <sup>2+</sup> (298 K)	0.26	285.7	0.9979
Freundlich constants	$K_f(mg/g)$	<i>n</i> (L/mg)	$R^2$
Pb <sup>2+</sup> (298 K)	99.753	4.24	0.5634
Langmuir constants	$K_a$ (L/mg)	$q_m (\mathrm{mg/g})$	$R^2$
GA (298 K)	0. 168	625.0	0.9918
Freundlich constants	$K_f(mg/g)$	<i>n</i> (L/mg)	$R^2$
GA (298 K)	123.53	2.55	0.7023

where  $q_e$  and  $q_t$  denote the amounts of metal adsorbed (mg/g) at equilibrium and at time *t* (min).  $K_1$  (min<sup>-1</sup>) and  $K_2$  (g/mg·min) are the rate constants of pseudo-first-order and pseudo-second-order models, respectively [40–42].

As shown in Figs. 21–24 and Table 3, in both cases, due to the compatibility of the results with the pseudo-second-order kinetic model, in addition to physical adsorption, chemical adsorption has also been performed [32].

# 3.7. Thermodynamic parameters

Thermodynamic parameters including Gibbs free energy change ( $\Delta G^{\circ}$ ), enthalpy change ( $\Delta H^{\circ}$ ) and entropy change ( $\Delta S^{\circ}$ ) were evaluated to determine the nature of the adsorption reaction. These parameters were evaluated by the following equations:

$$\Delta G^{\circ} = \Delta H^{\circ} - T \Delta S^{\circ} \tag{6}$$

$$\Delta G^{\circ} = -RT \ln K_d \tag{7}$$

$$\ln K_d = \frac{\Delta S^\circ}{R} - \frac{\Delta H^\circ}{RT} \tag{8}$$

$$K_d = \frac{q_e}{C_e} \tag{9}$$

where  $K_d$  is the equilibrium constant, T is the temperature (K), R is the gas constant.  $\Delta H^\circ$  and  $\Delta S^\circ$  are the change of enthalpy and entropy and by the help of Van't Hoff plots were determined. Results presented in Figs. 25 and 26 and Tables 4 and 5.

According to the obtained results, the adsorption process is endothermic and spontaneous and due to the positive



Fig. 21. Pseudo-first-order kinetics for GA.



Fig. 22. Pseudo-second-order kinetics for GA.



Fig. 23. Pseudo-first-order kinetics for Pb<sup>2+</sup>.



Fig. 24. Pseudo-second-order kinetics for Pb2+.

Table 3
Adsorption kinetic parameters



Fig. 25. Van't Hoff plot for thermodynamic parameters of GA.



Fig. 26. Van't Hoff plot for thermodynamic parameters of  $Pb^{2+}$ .

Table 4 Thermodynamic parameters at different temperatures for  $Pb^{\scriptscriptstyle 2+}$ 

T (K)	$\Delta H^{\circ}$ (kJ/mol)	$\Delta S^{\circ}$ (kJ/mol·K)	$\Delta G^{\circ}$ (kJ/mol)
298			-5.750
303	18.542	71.487	-6.896
308			-7.800
313			-8.995



Thermodynamic parameters at different temperatures for GA

T (K)	$\Delta H^{\circ}$ (kJ/mol)	$\Delta S^{\circ}$ (kJ/mol·K)	$\Delta G^{\circ}$ (kJ/mol)
298			-5.202
303	( (10	27 (02	-5.652
308	6.619	27.603	-6.633
313			-7.371

Pseudo-first-order kinetic model (Pb <sup>2+</sup> )		Pseud	Pseudo-second-order kinetic model (Pb <sup>2+</sup> )			
$q_e (\mathrm{mg/g})$	$K_1 (\min^{-1})$	$R^2$	$q_e (\mathrm{mg/g})$	$K_2$ (g/mg·min)	$R^2$	
76.30	0.046	0.9649	103.09	0.00085	0.9726	
Pseudo-first-order kinetic model (GA)			Pseud	Pseudo-second-order kinetic model (GA)		
$q_e (\mathrm{mg/g})$	$K_1 (\min^{-1})$	$R^2$	$q_e (\mathrm{mg/g})$	$K_2$ (g/mg·min)	$R^2$	
41.96	0.053	0.9953	100	0.0026	0.9978	

Application of carbonized datestone for the removal of Pb2+ in groundwater samples

Water samples	Before adsorption	After adsorption	% removal	RSD (%)
	Pb <sup>2+</sup> in mg/L	Pb²⁺ in mg/L	of Pb <sup>2+</sup>	
Tap water	7.33	1.61	78.03	2.2
Groundwater	9.33	2.11	77.37	2.3
Industrial water from Charmshar	20.78	4.80	76.90	2.4
Quds Industrial Town	21.25	4.95	76.70	2.4

Table 7

Comparison of this study with some of the methods reported in the literature

Adsorbent	Adsorbate	Adsorption capacity	References
		(mg/g)	
PEO/chitosan	Pb <sup>2+</sup>	237.2	[45]
Chitosan/HAp	$Pb^{2+}$	296.7	[46]
Fe <sub>2</sub> O <sub>3</sub> /Al <sub>2</sub> O <sub>3</sub>	$Pb^{2+}$	23.75	[47]
PEI/PES	$Pb^{2+}$	94.34	[48]
Chitosan/GO	$Pb^{2+}$	461.3	[49]
Chitosan-vanilin	$Pb^{2+}$	66.23	[50]
Hydrothermal	$Pb^{2+}$	133.2	[51]
biochar			
Lactic acid bacteria	$Pb^{2+}$	10.51	[52]
Magnetic adsorbent	$Pb^{2+}$	70.4	[53]
Citrus grandis	$Pb^{2+}$	207.2	[54]
Magnetic chitosan	$Pb^{2+}$	175.4	[55]
PET fibers	GA	38.2	[56]
Magnetic resin	GA	30.75	[57]
This study	Pb <sup>2+</sup> , GA	276,576	

entropy, confirms the appropriate interaction of adsorbent and adsorbate.

# 3.8. Desorption and regeneration study

Desorption was performed by 0.1 M  $\text{HNO}_3$ . Adsorption and desorption were performed 5 times and the adsorbent efficiency reached approximately 90% [33,43,44].

# 3.9. Proposed adsorbent application in wastewater samples

At the best condition of adsorbent (time = 75 min, pH = 4.0, adsorbent = 0.05 g), the adsorbent efficiency in the removal of the  $Pb^{2+}$  was investigated. Percentage of removal of  $Pb^{2+}$  ions in samples of well water, Charmshahr industrial water and other water samples were about 77% (Table 6). The presence of other ions in real water samples slightly interfered with the adsorption of  $Pb^{2+}$  ions under optimal conditions. Hence, the percentage of  $Pb^{2+}$  ion removal in groundwater samples was less than the value obtained using Box–Behnken statistical analysis.

#### 3.10. Comparison of this method with previous methods

Comparison of this method with other methods shows that the adsorption speed and capacity are high and the adsorption time is shorter. Therefore, it is an advantage over the previous methods. Some results can be seen in Table 7.

# 4. Conclusion

In this study, a new adsorbent based on chitosan grafted with a mixture of IA-MAm monomers was prepared and characterized. Its efficiency in Pb<sup>2+</sup> and GA adsorption was evaluated according to the batch technique. Effective parameters on adsorption were optimized. In acidic pH, high adsorption efficiency was observed. Consistency with the Langmuir isotherm model and the pseudo-second-order kinetic model showed that the adsorption process of both adsorbates is homogeneous, monolayer and chemical. On the other hand, the results of adsorption–desorption tests showed that the adsorbent has good efficiency and reproducibility. The adsorbent efficiency in the removal of lead ions in industrial effluents was appropriate and acceptable.

#### **Conflict of interest**

The authors declare that they have no conflict of interest. The datasets generated during and/or analysed during the current study are available from the corresponding author on reasonable request.

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