A review on interaction of phenolic pollutant with other pollutants in the binary adsorption system

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ABSTRACT

Due to the drastic industrial development, it has become evident that various harmful chemicals are released into the environment. Industrial effluent comprises pollutants of varied nature such as organic pollutants, heavy metals, dyes, and other toxic chemicals. Phenolic and its derivative pollutants are such groups of organic compounds which are widely detected in domestic wastewater. The pollutant is also associated with many different industrial effluents, making it a heterogenous mixture. So, to abate this toxic waste many treatment methods have emerged. Adsorption is one such treatment methods which has become the best available methods to treat an effluent of this nature. However, during the process of adsorption mixture of contaminants involves complex interactions such as synergism, antagonism, non-interaction, and competition between the solute molecules for the adsorbent. In this review paper, we have tried to identify interactions in binary adsorption factors present in 3 multicomponent isotherms and other factors which are mostly associated with binary system to arrive on to suitable interaction.

Keywords: Adsorption; Competitive; Interaction; Isotherm models; Multicomponent adsorption

1. Introduction

Effluent from the industry is the source of many diversified pollutants. The effluent is mainly composed of a heterogeneous mix of contaminants such as heavy metals, organic and inorganic pollutants and is accompanied by other parameters such as total dissolved solids, colour [1–3]. Phenolic compounds are the most commonly monitored pollutants in industrial effluents, agricultural and domestic discharges [4–6]. Recently, phenol and its derivatives, such as chlorophenol, 2,4-dichlorophenol (2,4-DCP) etc., have attracted significant attention. It could be owing to their comprehensive resources, such as effluents discharged from many industries. They are coke ovens in steel plants, petroleum refineries, petrochemical, phenolic resin, pesticide and fertiliser, pharmaceutical chemicals, dye industries, coal processing, plastics and wood products, paint and paper industries, explosives and leather treatments [7–9]. Though these compounds are present in small concentrations in frequently used items like plants, microorganisms, foods, cosmetics etc. their large-scale presence in the environment is attributed to the various human activities as these compounds are extensively used in various industries [10–13].

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There are strict regulations worldwide when the phenolic compound is considered because of its extreme toxicity. Due to their toxic effects, effluents containing phenolic compounds cannot be discharged without proper treatment. Also, the regulations on the discharge of effluents containing phenol and associated compounds, such as 2,4-DCP, chlorophenol etc., have been imposed by various environmental protection agencies and boards. Some of the important properties of phenol and 2,4-DCP are given in Table 1.

Methods such as adsorption, biodegradation, photocatalysis, etc., have been implemented to date for the remediation of such pollutants. Among them, adsorption using various adsorbents such as granular activated carbon [11], coconut husk activated carbon [18], biologically activated carbon

Table 1

Physical and chemical properties of phenol and 2,4-dichlorophenol

Parameters	Phenol	2,4-DCP	References
Formula	C ₆ H ₆ O	C ₆ H ₄ Cl ₂ O	
Molecular weight	94.11 g/mol	163.1 g/mol	
Solubility in water at 25°C	8.3 g/100 mL	5 g/100 mL	
Density	1.07 g/cc	1.38 g/cc	
Melting point	40.91°C	42°C-43°C	
Boiling point	181.75°C	210°C	
λ _{max} (nm)	270	283	[13–17]
Dipole moment	1.22 D	1.60 D	
рКа	9.89	7.9	
Structural formula	0 ^{,H}	0 ^{.H}	

[17], rice husk ash [19], iron nanocomposites [20] etc, are investigated.

Several studies have been carried out on the adsorption of single-component pollutants [21–23]. This has happened for many decades because of the ease in conducting experiments. It is evident from the literature that, in recent two decades, studies on binary and ternary component adsorption have been advancing [24–27]. Due to the heterogeneity associated with phenol, 2-DCP and many such pollutants, it becomes challenging for an environmental engineer to abate phenolic pollutants at the lab scale using adsorption techniques [28,29]. Adding to this, difficulty is also due to complexity of interaction factors when more than one pollutant is considered for remediation using adsorption technique.

Moreover, the interaction phenomena such as synergistic (SYI), antagonistic (ANI) and non-interaction (NI) are observed depending on the adsorption capacity of the pollutants. This could be due to some inherent effects such as hydrophobic effect, π – π interactions, electrostatic attraction, surface complexation, co-precipitation etc [16–19]. The interactions between adsorbents, adsorbate and between contaminants could be due to charge on pollutant, surface area, porosity of adsorbent etc. [34]. So, it is essential to understand the potential interactions occurring in the system.

A survey was made using the Scopus survey tool for the keywords "Simultaneous", "Adsorption", and "Interaction". It was found that a around 2643 (from 1930–2022) different research articles have investigated the interaction factor in their respective domains. Fields such as Chemistry, Chemical Engineering, and Environmental Engineering contribute to more than 50% of the total research. From Fig. 1 it can also be concluded that there is ample scope to explore this field. Furthermore, it becomes imperative for current researchers to contribute more towards this study and investigate the interaction factors participating in adsorption process. These interaction factors will help to understand the multi-component adsorption mechanism



Fig. 1. Scope of the study.

better. A considerable amount of literature has been published on investigating interactions in binary adsorption system by means of adsorption capacity only. To some extent affinity, selectivity and inhibition was also discussed in some investigations. However, degree of interactions of SYI, ANI and NI are least discussed. This review looks into the simultaneous adsorption of organic components (phenols/2,4-dichlorophenol) with other components in a binary system. It is focused on (i) understanding the different interactions in binary adsorption setup by means of different factors as mentioned in section 2. (ii) Further, different possible mechanisms responsible for the interaction are discussed. (iii) Besides, an attempt to generalise the factors is carried out to arrive at specific interaction.

2. Types of interaction in binary system

Studies on competitive adsorption of two pollutants involve three possible mechanisms. The basic principle governing these mechanisms is by considering the adsorption capacity of the contaminants. They are adsorption capacity (i) in the binary solution $(Q_{i,o})_b$ and the adsorption capacity adsorbate (ii) in a single-component system $(Q_{i,o})_s$.

2.1. Based on adsorption capacity

2.1.1. Synergistic interaction

It is defined as a type of interaction observed in bisolute adsorption system where, presence of one solute enhances the adsorption capacity of other [35,36]. This interaction can also be depicted in terms $(Q_m)_b$ and $(Q_m)_s$ which represents adsorption capacity in binary and single adsorption systems, respectively. In this case, the ratio of $(Q_m)_b$ to $(Q_m)_s$ is always greater than 1. Mathematically, it can be represented as below:

$$\frac{\left(Q_{m}\right)_{b}}{\left(Q_{m}\right)_{s}} > 1 \tag{1}$$

The synergistic interaction is exemplified in the work undertaken by Nguyen et al. [35] and Dhandole et al. [37] wherein the former study involves simultaneous adsorption of phenol and Ni(II) and latter involves investigation on synergistic effect of heavy metals on to nanorods/ nanotubes.

2.1.2. Antagonistic interaction

It is defined as a type of interaction observed in binary adsorption system, in which there is decline in adsorption capacity of a pollutant when associated with another pollutant [19,33] and the ratio of $(Q_m)_b$ to $(Q_m)_s$ is less than 1. Mathematically it can be written as below:

$$\frac{\left(Q_{m}\right)_{b}}{\left(Q_{m}\right)_{s}} < 1 \tag{2}$$

Few examples of what is meant by antagonistic interaction can be found in research carried out by Thakur et al. [19] and Rathore et al. [38]. The schematic representation of various interactions in bisolute adsorption system is shown in Fig. 2.

2.1.3. Non-interaction

It is defined as a "type of interaction in which there is no significant change in adsorption capacity of a pollutant when associated with another pollutant" [39,40]. Mathematically, non-interaction can be represented as follows:

$$\frac{(Q_m)_b}{(Q_m)_s} \approx 1 \tag{3}$$

The above interaction has been reported in an investigation carried out by Jin et al. [41], where adsorption capacity of copper ions and 2,4-DCP did not alter in binary adsorption system. Similar interaction was communicated by Kumar et al. [42].

2.2. Based on multi-component isotherm model and other factors

Significant literature is available for binary adsorption systems. These studies can be modelled and validated using many modified and non-modified competitive isotherms/ models. But not all of these models efficiently identify the interaction between the adsorbate molecules of a binary system. Also, in most of the cases the parameters calculated from the single-component adsorption isotherm model do not delineate the interaction between the individual components. Some models, for example, modified competitive Langmuir isotherm, modified competitive Redlich–Peterson isotherm and Sheindorf–Rebuhn–Sheintuch (SRS) equation have incorporated parameters to identify an interaction.

2.2.1. Modified competitive Langmuir isotherm

In a binary system, an interaction term has to be introduced to describe the mechanism of the adsorption process. This interaction term shows the competitive effect of adsorbates in the solution. It is a characteristic parameter of a component, and it depends on the concentration of the other components present in the solution [11,18]. It is represented mathematically as below:

$$\left(Q_{i,e}\right)_{b} = \frac{\left(Q_{i,o}\right)_{b} \times b_{i} \times \frac{\left(C_{i,e}\right)_{b}}{\eta_{i}}}{1 + \sum_{j=1}^{N} b_{j} \times \frac{\left(C_{j,e}\right)_{b}}{\eta_{i}}}$$
(4)

where $Q_{i,e}$ equilibrium adsorption capacity for component *i*, mg/g; $(C_{i,e})_b$ concentration, mg/L; $(Q_{i,o})_b$ monolayer adsorption capacity for component *i*, mg/g; b_i Langmuir constant for component *i*, L/mg.

The value of *N* for the binary system is 2. The notations η_i and η_j are the interaction factors for components *i* and *j*. The value of η_i is a function of characteristics of



Fig. 2. Schematic representation for different interactions in binary system. (a) Single system – before adsorption, (b) single system – after adsorption, (c) binary system – before adsorption, (d) binary system – after adsorption (synergistic interaction), (e) binary system – after adsorption (non-interaction), (e) binary system – after adsorption (non-interaction).

adsorbent, concentration of component *j*. A higher value of η_i implies a lower affinity of component *i* to the adsorbent and vice versa.

2.2.2. Sheindorf-Rebuhn-Sheintuch equation

The interaction factor out of SRS equation has potential to investigate the inhibition of adsorption for a component in a binary solution by another component [36–38]. The equation is given as below:

$$\left(Q_{i,e}\right)_{b} = k_{i} \times \left(C_{i,e}\right)_{b} \times \left(\sum_{j=1}^{N} \left(\theta_{i,j}\left(C_{j,e}\right)_{b}\right)\right)^{\left(\frac{1}{n_{i}}\right)^{-1}}$$
(5)

where $Q_{i,e}$ – equilibrium adsorption capacity for component i, mg/g; $C_{e,i}$ – equilibrium concentration of an *i*th component, mg/L; k_i – Freundlich constant; n_i – adsorption intensity; θ_{ii} – competition coefficient.

 θ_{ij} indicates inhibition of adsorption of the *i*th component by adsorption intensity (N = 2) for a binary solution. The interaction component θ_{ij} shows inhibition of *i*th component on adsorption of the *j*th component. The higher value of θ_{ij} than θ_{ji} signifies that, in binary solution, the *j*th component has more potent inhibition. It also infers that component *i*th component has weaker inhibition on the adsorption of component *j* [25].

2.2.3. Modified competitive Redlich-Peterson isotherm

The special term $(\eta_{PR,i})$ added in modified competitive Redlich–Peterson isotherm (MCRPI) shows the interaction among the components in a multi-component solution The model can be represented as:

$$\left(Q_{i,e}\right) = \frac{\left(k_{\mathrm{PR},i}\left(C_{e,i} / \eta_{\mathrm{PR},i}\right)\right)}{1 + \sum_{j=1}^{N} \left(a_{\mathrm{PR},j}\left(C_{e,j} / \eta_{\mathrm{PR},j}\right)^{\beta,j}\right)}$$
(6)

where $Q_{i,e}$ – equilibrium adsorption capacity for component *i*, mg/g; $C_{e,i}$ – concentration, mg/L; $\eta_{PR,i}$ and $\eta_{PR,j}$ – model constant, L/mg; β and a_{PR} – model constant, (L/mg)^{β}.

The value of N for the binary system is 2. The notations $\eta_{\text{PR},i}$ and $\eta_{\text{PR},i}$ are the interaction factors for component *i* and *j* and the value of $\eta_{PR,i}$ is dependent on the concentration of component *j*. A higher value of $\eta_{PR,i}$ implies lower affinity of that component towards adsorbent and vice versa. The studies including these factors are calculated in Table 2 and the significance of each factor is explained. Further, it is important to study the interaction behaviour of the pollutant in a binary solution. It is required to investigate on how the adsorbent-components interact with each other and with adsorbate. It is moreover prominent to note that, how one adsorbent-pollutant dominates over other or gets inhibited by other pollutant in solution. Furthermore, it is also significant to understand affinity of a pollutant towards the adsorbent and how the adsorbents are attracted towards the solid surface. In the previous studies, these discussion on various interactions were investigated comparing the adsorption capacity as factor. They can also be explained based on the interaction factors namely selectivity ratio, sorption affinity, P-factor, and inhibition factor.

2.2.4. Selectivity ratio $(S_{(i/j)})$

It is the ratio of adsorption capacity of *i*th solute to *j*th in the single/binary adsorption system. The ratio gives an idea of the affinity of an adsorbent towards a component in a binary solution. It also aids in the investigation of selective preference of the desired adsorbent for a particular pollutant. It is due to surface structure, morphology, and pore distribution [23–26]. The selectivity ratio can be represented in two forms, one each for single component and binary component adsorption system as show below:

$$S_{(i/j)} = \frac{\left(Q_{i,o}\right)_b}{\left(Q_{j,o}\right)_b}$$
(7)

$$S_{(i/j)} = \frac{\left(Q_{i,o}\right)_s}{\left(Q_{j,o}\right)_s} \tag{8}$$

where $(Q_{i,o})_b$ — monolayer adsorption capacity for component *i* in binary system, mg/g; $(Q_{j,o})_b$ — monolayer adsorption capacity for component *j* in binary system, mg/g; $(Q_{i,o})_s$ — adsorption capacity for component *i* single system, mg/g; $(Q_{j,o})_s$ — adsorption capacity for component *j* in single system, mg/g.

If the ratio $S_{(i/j)}$ is more than $S_{(j/i)'}$ it implies that the adsorbent has a stronger affinity towards component *i* than component *j* and vice versa.

2.2.5. Sorption affinity (k_s^*)

It refers to an attraction of solvent to a solute molecule present in a binary solution. The factor helps compare the adsorption capabilities of different adsorbents in a binary system. It is represented by the equation as given below:

$$k_s = \frac{\left(C_{i,o}\right)}{\left(C_{i,s}\right)_h} \tag{9}$$

where $(C_{i,o})$ — initial concentration of a component, mg/L; $(C_{i,o})_b$ — concentration of solute in a solid phase in binary solution, mg/L.

The greater value of k_s^* of component *i* and *j*, indicates a weaker affinity of the adsorbent to the solute molecule in a binary system [28]. The factor can also be used to distinguish the interaction in bisolute adsorption setup. However, the formula can only be used to identify interaction when the initial concentrations of component *i* and *j* are same (say 100 mg/L).

2.2.6. P-factor (P_{c})

It is defined as the ratio of adsorption capacity of pollutant in mono-component to adsorption capacity in binary solution. This factor compares data about single-component equilibrium and the binary component. P-factor helps to understand the competitive adsorption of the components in a binary solution [23–28]. The factor also explains how another component in a binary solution inhibits the adsorption of a component. It is given by the expression:

$$P_{\rm fi} = \frac{\left(Q_{i,o}\right)_s}{\left(Q_{i,o}\right)_h} \tag{10}$$

where $(Q_{i,o})_s$ — adsorption capacity for component *i* single system, mg/g; $(Q_{i,o})_b$ — monolayer adsorption capacity for component *i* in binary system, mg/g

The value of $P_{\rm fi}$ determines the inhibition of adsorption of a component by another component in a binary solution [23]. $P_{\rm fi}$ greater than one, implies the adsorption of component *i* is inhibited in presence of component *j* also the component *i* exhibits antagonistic interaction in the presence component *j*. The value of $P_{\rm fi}$ lower than unity, implies the adsorption of component *i* is enhanced in presence of component *j* in the binary solution and the component *i* exhibits synergistic interaction in presence component *j*. For the value of $P_{\rm fi}$ equal to unity suggests that the adsorption of both pollutants are unaltered by other component and both components follow non-interaction.

2.2.7. Inhibitory effect (ΔIE)

The adsorption of an adsorbate may be inhibited by another component present in the binary system which can be described by the inhibitory effect (Δ IE). Mathematically it is represented as below:

$$\Delta IE_{i} = \frac{(Q_{i,o})_{s} - (Q_{i,o})_{b}}{(Q_{i,o})_{s}}$$
(11)

where $(Q_{i,o})_s$ — adsorption capacity for component *i* single system, mg/g; $(Q_{i,o})_b$ — monolayer adsorption capacity for component *i* in binary system, mg/g.

	Refer-	ences	[62]	[1]
	Inference		Phenol and CN adsorp- tion process enhanced at a pH 9. The type of inter- action observed was syn- ergistic. The governing mechanisms were found to be hydrogen bonding between phenol and cya- nide and ion exchange.	Synergism in cyanide ions is not significant which can be observed from the value of inhibitory factor nearing to 0. This inter- action between CN and phenol could be attributed to hydrogen bonding and intraparticle diffusion.
	Interaction	based on factors	Synergistic (phenol, CN) Synergistic (phenol, CN) Synergistic (phenol, CN) CN) CN	Synergistic (phenol, CN) Synergistic (phenol, CN) CN) CN) CN) CN) CN)
	Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 1 Pollutant 1 has less affinity w.r.t pollut- ant 2 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed Adsorption of pollutant 2 is suppressed by	pollutant 1 Adsorbent has a stron- ger affinity towards pollutant 1 Pollutant 1 has less affinity w.r.t pollut- ant 2 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed Adsorption of pollutant 2 is suppressed by pollutant 1
	Factors		Selectivity ratio: 91.26, 13.79 Sorptive affinity: 1.11, 1.17 P-factor: 0.037, 0.910 0.037, 0.910 Inhibitory effect: -26.3, -0.09	Selectivity ratio: 138, 7.3 Sorptive affinity: 1.25, 1.068 P-factor: 0.049, 0.92 ffect: -19.73, -0.09
	g/g)	Pollut- ant 2 (Binary)	3.21	1.95
	apacity (m	Pollut- ant 1 (Binary)	292.95	269.7
	sorption c	Pollut- ant 2 (Single)	2.92	1.79
	Ad	Pollut- ant 1 (Single)	40.29	13.01
		Speed (rpm)	125	125
	litions	Hq	6	œ
	Experimental con	Initial concen- tration (g/L)	0.2/0.02	0.2/0.02
		Temp. (K)	303	303
		Adsor- bent dosage (g/L)	10	06
	ollutant	(2)	CN (cyanide)	S
	Ŀ	(1)	Phenol	Phenol
	Adsorbent _		Copper impreg- nated GAC (Cu-GAC)	Granular activated carbon (GAC)
	s.	No.		ci

Refer-	ences	[18]	[63]
Inference		Synergistic interaction of phenol and CN was observed which can be clearly understood by negative values of inhibi- tory effect. Weak van der Waal forces were found to influence the adsorption process.	The values of <i>P</i> -factor and inhibitory effect depict strong synergistic effect and a weak antagonistic effect in phenol and CN, respectively. This is due the fact that adsorbent has stronger affinity towards phenol as observed from the values of selectivity ratio.
Interaction	based on factors	Synergistic (phenol, CN) Synergistic (phenol, CN) Synergistic (phenol, CN)	Synergistic (phenol) Antag- onistic (cyanide) Synergistic (phenol) Antag- onistic (phenol) Antagonis- tic (CN) Antagonis- tic (CN) tic (CN)
Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 1 Pollutant 1 has less affinity wr.t pollutant 1 has ant 2 Adsorption of pollutant 2 is suppressed Adsorption of pollutant 2 is suppressed by pollutant 1	Adsorbent has a stron- ger affinity towards pollutant 1 Pollutant 1 has less affinity w.r.t pollut- ant 2 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed Adsorption of pollutant 2 is suppressed by pollutant 1
Factors		Selectivity ratio: 45.3, 6.12 Sorptive affinity: 1.39, 1.152 P-factor: 0.028, 0.206 Inhibitory effect: -34.96, -3.86	Selectivity ratio: 160.5, 7.45 Sorptive affinity: 1.4, 1.13 1.13 P-factor: 0.052, 1.12 0.052, 1.12 finhibitory effect: -18.237, 0.107
ıg/g)	Pollut- ant 2 (Binary)	5.30	2.18
apacity (m	Pollut- ant 1 (Binary)	239.85	349.86
lsorption c	Pollut- ant 2 (Single)	1.09	2.44
AG	Pollut- ant 1 (Single)	6.67	18.19
	Speed (rpm)	120	120
ditions	Hq	∞	м
nental con	Initial concen- tration (g/L)	0.3/0.03	0.5/0.05
Experin	Temp. (K)	303	303
	Adsor- bent dosage (g/L)	4	6 4
Pollutant	(1) (2)	Phenol CN	Phenol CN
Adsorbent		Coconut shell-ac- titvated carbon	<i>Pseudomo-</i> <i>mas putida</i> on coconut shell-ac- tivated carbon
s.	No.	с,	4

Refer-	ences	[58]	[64]
Inference		Filling of pores by solute molecules can observed as one of the adsorption mechanisms as found from FE-SEM images. Higher degree of Synergism in phenol was observed as compared to CN. From the investigation, potential route of adsorption was found to be chemisorption.	In the process of binary adsorption, antagonistic interaction was depicted by phenol and CN. The interaction was least sig- nificant. However, CHAC showed a higher affinity towards phenol.
Interaction	based on factors	Synergistic (phenol, CN) Synergistic (phenol, CN) CN) Synergistic (phenol, CN)	Antagonis- tic (phenol, CN) Antagonis- tic (phenol, CN) CN) CN) CN) CN) CN) CN) CN)
Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 1 has less affinity w.r.t pollutant 1 has ant 2 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed Adsorption of pollutant 2 is	Adsorbent has a stron- ger affinity towards pollutant 1 has less affinity w.r.t pollut- ant 2 Adsorption of pollutant 2 is enhanced and pollutant 1 is suppressed pollutant 1 is suppressed by pollutant 2
Factors		Selectivity Ratio: 17.26, 5.253 Sorptive affinity: 1.49, 1.26 P-factor: 0.048, 0.159 Inhibitory effect: -19.62, -5.3	Selectivity ratio: 10.16, 10.8 Sorptive affinity: 1.11, 1.07 P-factor: 1.334, 1.259 Inhibitory effect: 0.25, 0.205
g/g)	Pollut- ant 2 (Binary)	5.46	0.58
apacity (m	Pollut- ant 1 (Binary)	94.25	5.89
sorption ca	Pollut- ant 2 (Single)	0.87	0.73
Ads	Pollut- ant 1 (Single)	4.57	7.86
	Speed (rpm)	120	120
itions	Hq	Ν	∞
nental cond	Initial concen- tration (g/L)	0.3/0.03	0.3/0.03
Experin	Temp. (K)	303	303
	Adsor- bent dosage (g/L)	64	30
Pollutant	(2)	CN CN	E CN
Ţ	(1)	Phenc	Phenc
Adsorbent		Sugarcane bagasse	Coconut husk activated carbon (CHAC)
s.	No.	ы.	٠

Refer-	ences	[17]	[35]
Inference		Both the pollutants depicted antagonis- tic interaction in binary adsorption setup. The reason could be the affin- ity of pollutant towards adsorbent, relative concen- tration, molecular size of each pollutant, and their mutual interaction.	The adsorbent exhibited stronger affinity towards phenol as observed from the values of sorption affinity. The positively charged ions get attracted towards the adsorbent due to negative charge induced on the surface of the adsorbent. Chemical bonding and co-ordination also support in adsorption process.
Interaction	based on factors	Antagonis- tic (phenol, 2,4-DCP) tic (phenol, 2,4-DCP) Antagonis- tic (phenol, 2,4-DCP) 2,4-DCP) 2,4-DCP) 2,4-DCP)	Synergistic (phenol, Ni) Synergistic (phenol, Ni) Ni) Ni) Synergistic (phenol, Ni) Ni)
Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 2 Pollutant 2 has less affinity w.r.t pollut- ant 1 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed Adsorption of pollutant 2 is suppressed by pollutant 1	Adsorbent has a stron- ger affinity towards pollutant 1 Pollutant 2 has less affinity w.r.t pollut- ant 1 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed pollutant 2 is suppressed by pollutant 1
Factors		Selectivity ratio: 0.85, 0.55 Sorptive affinity: 2.65,4.51 P-factor: 1.3, 1.9 Inhibitory effect: 0.21, 0.5	Selectivity ratio: 1.13, 1.08 Sorptive affinity: 2.36, 2.65 P-fac- tor:0.52, 0.543 0.543 0.543 Inhibitory effect: -0.923,
g/g)	Pollut- ant 2 (Binary)	22.18	8. 8.
ıpacity (m	Pollut- ant 1 (Binary)	18.95	21.2
orption ce	Pollut- ant 2 (Single)	43.75	10.2
Ads	Pollut- ant 1 (Single)	24.04	1
	Speed (rpm)	120	500
ditions	Hq	Ν	4
nental cond	Initial concen- tration (g/L)	0.116/0.1	0.05
Experim	Temp. (K)	303	303
	Adsor- bent dosage (g/L)	1.87	р
llutant	(2)	2,4-dichlo- rophenol (2,4-DCP)	Ni(II)
Po	(1)	Phenol	Phenol
Adsorbent		Biolog- ically activated carbon	Chitosan beads (chemically and bio- logically modified)
s.	No.	ĸ	ά

Refer-	ences	an take atake atio sug- sug- nuc- nner- $\pi - \pi$ nnsi- nnsi- istic vas ved vved vved	1.7). t of the the tore: and [42] and fon. be for s for
Inference		In the presence of organic compound, up of $Cr(VI)$ increased. values of selectivity and sorption affinity gest the same. Also, it due to the coupling re- tion of $Cr(IV)$ with ϵ gy-yielding reactions. significant forces such van der Waals forces, interactions were respo ble for adsorption pro- ble for adsorption pro- the degree of synerginteraction in phenol less as compared to c mium ion as obser- from the values of inl	tory factor (-0.0185, -5 In a binary mixture phenol and resorcinol, former adsorbed m This is because the soli ity of phenol is less, r ing it hydrophobic favourable for adsorp Non-interaction can explained from the e values of all the factor. 2 solutes.
Interaction	based on factors	Synergistic (phenol, Cr) Synergistic (phenol, Cr) Synergistic (phenol, Cr) Cr) Cr) Cr) Cr)	No interac- tion (phe- nol and resorcinol) No interac- tion (phe- nol and resorcinol) No interac- tion (phe- nol and resorcinol) resorcinol) resorcinol) resorcinol)
Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 2 Pollutant 1 has less affinity wr.t pollut- ant 2 Adsorption of pollutant 1 is suppressed pollutant 1 is suppressed by	Adsorbent has a equal affinity towards pol- lutant 1 and 2 Pollutant 1 and pollutant 2 have same affinity Adsorption of pollutant 1 and 2 was nei- ther enhanced nor inhibited Adsorption of pollutant 1 and 2 did not alter
Factors		Selectivity ratio: 0.01, 0.54 Sorptive affinity: -1.17, 1.02 P-factor: 0.982, 0.019 Inhibitory effect: -0.0185,	Selectivity ratio: 1.05, 1.05 Sorptive affinity: 1.03, 1.02 P-factor: 1.0, 1.0 Inhibitory effect: 0, 0
(g/gu	Pollut- ant 2 (Binary)	741.389	24.25
capacity (n	Pollut- ant 1 (Binary)	7.7608	25.41
lsorption (Pollut- ant 2 (Single)	14.075	24.25
Ac	Pollut- ant 1 (Single)	7.62	25.41
	Speed (rpm)	120	150
ditions	Hq	ى ب	2.2
nental con	Initial concen- tration (g/L)	0.25-0.5	0.1-1
Experir	Temp. (K)	310	303
	Adsor- bent dosage (g/L)	ដ	4 -5
lutant	(2)	Cr(VI)	Resorcinol
Poll	(1)	Phenol	Phenol
Adsorbent		<i>Bacillus</i> sp. onto tea waste biomass	GAC
s.	No.	6	10.

Pollut- Pollut	I Speed Poll(rpm) ant(Sin	l pH an-	Initia. conce	Tomm Later
ingle) (E		(Single) (S	(Single) (Si	teuty. цицан р.т. эрееа гоцит- Га (K) concen- (rpm) ant 1 ar tration (Single) (S (g/L)
	6.2	150 9.03 6.2	.0.00-1 0 150 0.03 0.27	303 0.050-1 6 150 9.03 6.2
c				
	4.97	120 9.97 4.97	1.025 - 120 9.97 4.97	76.4 76.6 071 - 67.0 50.6

Pollı	utant		Experim	nental conc	ditions		A	dsorption	capacity (me/e)	Factors	Findings from	Interaction	Inference	Refer-
(2)	1	Adsor- bent dosage (g/L)	Temp. (K)	Initial concen- tration (g/L)	Hq	Speed (rpm)	Pollut- ant 1 (Single)	Pollut- ant 2 (Single)	Pollut- ant 1 (Binary	Pollut- ant 2 (Binary)		factor	based on factors		ences
ol blue (M	B B)		303	0.025	1	120	26.6	2.88	8.92	6.51	Selectivity ratio: 1.37, 1.265 Sorptive affinity: -1.23, 1.23 P-factor: 1.12, 1.2 1.12, 1.2 Inhibitory effect: 0.105, 0.174	Adsorbent has a stron- ger affinity towards pollutant 1 Pollutant 2 has less affinity w.r.t pollut ant 1 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed Adsorption of pollutant 2 is suppressed by pollutant 1	Antagonis- tic (phenol, MB) Antagonis- tic (phenol, MB) MB) MB) Antagonis- tic (phenol, MB)	The adsorption of phe- nol and MB was signifi- cantly less from the binary adsorption system than single-contaminant solu- tions, suggesting com- petition between the two for adsorption sites, both being organic molecules. The affinity of both pollut- ants was also most same as observed from the values of sorption affinity. Fur- ther the value of inhibitory effect suggests minimal degree of antagonistic interaction. The adsorp- tion was due to electro-do- nor-acceptor mechanism and ion-exchange.	[99]
ol Cr(VI)		20	298	0. D	-1	150	244.80	302.9	103.2	198.4	Selectivity ratio: 0.52, 0.808 Sorptive affinity: 3.0, 2.36 2.4, 1.53 2.4, 1.53 2.4, 1.53 2.4, 1.53 0.35 0.35	Adsorbent has a stron- ger affinity towards pollutant 2 Pollutant 1 has less affinity w.r.t pollut- ant 2 Adsorption of pollutant 2 is enhanced and pollutant 1 is suppressed Adsorption of pollutant 1 is suppressed by pollutant 2	Antagonis- tic (phenol, Cr) Antagonis- tic (phenol, Cr) Cr) Antagonis- tic (phenol, Cr) Cr)	The value of P-factor (> 1) for both the solute sug- gests antagonistic inter- action. The reason for antagonistic interaction during the multi-compo- nent adsorption was com- petition for similar binding sites onto the adsorbent of cells and/or the screening effect by the second-com- ponent. Competition on similar binding sites and screening effect was also observed.	[40]

Refer-	ences	[54]	[16]
Inference		Adsorbent depicted stron- ger affinity towards cad- mium ion. Significant antagonistic behaviour of phenol was observed. Adsorption of Cd(II) and phenol onto composites is assigned due to strong complexation, and for phe- nol, it is due to π - π cova- lent interaction between phenol and adsorbent.	GAC showed more affinity towards phenol in com- parison with 2,4-DCP. The adsorption and interaction was also influenced by Weak hydrophobic inter- action.
Interaction	based on factors	Antagonis- tic (phenol) Synergistic (Cd) Antagonis- tic (phenol) Synergistic (Cd) Synergistic	(Cd) Synergistic (phenol, 2,4-DCP) 2,4-DCP) Synergistic (phenol, 2,4-DCP) Synergistic (phenol, 2,4-DCP) 2,4-DCP)
Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 2 Pollutant 1 has less affinity wr.t pollutant 2 ant 2 Adsorption of pollutant 2 is enhanced and pollutant 1 is suppressed by pollutant 1 is	pollutant 2 Adsorbent has a stron- ger affinity towards pollutant 1 Pollutant 2 has less affinity wr.t pollut- ant 1 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed pollutant 2 is suppressed by pollutant 1
Factors		Selectivity ratio: 2.39, 3.148 Sorptive affinity: P-factor: 1.3, 0.98 Inhibitory effect: 0.223, -0.02214	Selectivity ratio: 1.97, 1.32 Sorptive affinity: 1.2, 1.23 1.23 P-factor: 0.57, 0.85 0.57, 0.85 effect: -0.747, -0.17
g/g)	Pollut- ant 2 (Binary)	87.7	22.18
apacity (n	Pollut- ant 1 (Binary)	209.9	43.75
sorption c	Pollut- ant 2 (Single)	82 82	18.96
A	Pollut- ant 1 (Single)	270.09	25.04
s	Speed (rpm)	1	120
Experimental conditions	Hd -u u		0.1 7
	 p. Initia conce tratio (g/L) 	0.05	0.116,
	e (K)	293	4 393
	Adsoi bent dosag (g/L)	0 5	0.3/0.1
lutant	(2)	Cd(II)	2,4-DCP
Pol	(1)	Phenol	Phenol
Adsorbent		Magnetic graphene oxide composites (polyani- line func- tionalized)	Granular activated carbon (GAC)
s.	No.	15.	.16.

Refer-	ences	[67]	[20]
Inference		Solubility of solutes has an important role while simultaneous adsorption of organic pollutants is investigated. Due to weak solubility of chlorophenol weak interaction with sol- vent is induced. The syn- ergistic and antagonistic behaviour of phenol and chlorophenol, respectively is very clearly portrayed by values of P-factor and inhibitory effect.	Synergistic interaction of phenol and MO was demonstrated in the study. This can also be deduced based on observation of values of adsorption capacities. Electrostatic and non-electrostatic inter- actions were among the governing forces of the synergistic interaction.
Interaction	based on factors	Synergistic (phenol) Antagonis- tic (2-CP) Synergistic (phenol) Antagonis- tic (2-CP) Synergistic (phenol) Antagonis- tic (2-CP) Synergistic (phenol) Antagonis- tic (2-CP) Synergistic (phenol) Antagonis- tic (2-CP) MO)	Synergistic (phenol, MO) Synergistic (phenol, MO) Synergistic (phenol, MO)
Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 2 Pollutant 1 has less affinity wr.t pollut- ant 2 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed pollutant 2 is suppressed by pollutant 1 Adsorbent has a stron- our affinity.	ger aurury towards pollutant 1 Pollutant 2 has less affinity w.r.t pollut- ant 1 Adsorption of pollutant 2 is enhanced and pollutant 1 is suppressed Adsorption of pollutant 1 is suppressed by pollutant 2
Factors		Selectivity ratio: 0.85, 0.759 Sorptive affinity: 1.123, 1.00 P-factor: 0.962, 1.076 0.962, 1.076 Inhibitory effect: -0.039, 0.07 Selectivity ratio: 1.02, 2.73	Sorptive affin- ity:1.11, 6.25 P-factor: 0.583, 0.268 Inhibitory effect: -0.72, -2.7
g/g)	Pollut- ant 2 (Binary)	251	71.02
apacity (n	Pollut- ant 1 (Binary)	213	72.68
sorption c	Pollut- ant 2 (Single)	270	19.03
Ρd	Pollut- ant 1 (Single)	505	42.34
	Speed (rpm)	250	0 200
onditions	Hq -n	4. Ú	4-1(
mental c	. Initial conce tration (g/L)	0.0	0.050
Experi	Temp (K)	58	298
	Adsor- bent dosage (g/L)	0.5	0.5-4
ollutant	(2)	2-Chlo- rophenol (2-CP)	Methyl orange (MO)
Pc	(1)	Phenol	Phenol
Adsorbent		Spherical activated carbon	FeNP/ Carbon nanocom- posites
S.	No.	13.	18.

Refer-	ences	[10]	[68]
Inference		The drop in adsorp- tion capacity of phenol and cadmium in binary adsorption system was significant. The positive value of inhibitory effect supports the antagonistic interaction. The interaction could be competition of the pollutants for adsorbent site. The governing mech- anism could be electron sharing and ion exchange, respectively.	In binary system, the affinity of phenol is more enhanced than aniline. This can be understood by the value of selectivity ratio and P-factor. Fur- ther, the negative value of inhibitory effect implies synergistic interaction. The driving force could be intraparticle diffusion. However, the solubility of solutes did not play sig- nificant role in bisolute system.
Interaction	based on factors	Antagonis- tic (phenol, Cd) Antagonis- tic (phenol, Cd) Antagonis- tic (phenol, Cd) Antagonis- tic (phenol, Cd)	
Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 1 Pollutant 2 has less affinity wr.t pollut- ant 1 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed pollutant 2 is suppressed by pollutant 1	per affinity best affinity towards pollutant 1 Pollutant 2 has less affinity w.r.t pollut- ant 1 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed pollutant 2 is suppressed by pollutant 1
Factors		Selectivity ratio: 2.53, 2.33 Sorptive affinity: 1.88, 3.57 P-factor: 1.09, 1.19 Inhibitory effect: 0.08, 0.16 Selectivity	ratio: 1.14, 1.109 Sorptive affinity: – P-factor: 0.97, 0.99 Inhibitory effect: –0.034,
g/g)	Pollut- ant 2 (Binary)	11.95	18.01
apacity (m	Pollut- ant 1 (Binary)	30.33	20.60
sorption ca	Pollut- ant 2 (Single)	17 2.	17.97
Ads	Pollut- ant 1 (Single)	33.11	19.92
	Speed (rpm)	200	600
nditions	Hd .	<u>ິ</u> ນ 21	٢
nental coi	Initial concen- tration (g/L)	0.1	0.050
Experin	Temp. (K)	298	294
	Adsor- bent dosage (g/L)	n	0 .5
utant	(2)	Cd(II)	Aniline
Pollı	(1)	Phenol	Phenol
Adsorbent	I	AC (municipal sludge)	GAC
s.	No.	19.	20.

Refer-	ences	33	[69]
Inference		Adsorption quantities of 2-nitrophenol are higher than those of phenol. This experimental evidence suggested an antagonistic effect between adsorbates, generating inhibitive adsorption.	Interaction of Ni(II) in the presence of phenols was antagonistic at pH 1. This could be due to compe- tition for different sites/ screening effect.
Interaction	based on factors	Antagonis- tic (phenol, 2-nitrophe- nol) Antagonis- tic (phenol, 2-nitrophe- nol) Antagonis- tic (phenol, 2-nitrophe- nol) Antagonis- tic (phenol, 2-nitrophe- nol)	Nij - - Antagonis- tic (phenol, Ni) Antagonis- tic (phenol, Ni)
Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 2 Pollutant 1 has less affinity w.r.t pollut- ant 2 Adsorption of pollutant 2 is enhanced and pollutant 1 is suppressed Adsorption of pollutant 2 is	has a stron- ger affinity towards pollutant 1 Pollutant 2 has less affinity wr.t pollut- ant 1 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed Adsorption of pollutant 2 is suppressed by pollutant 1
Factors		Selectivity ratio: 0.19, 0.893 0.893 Sorptive affinity: -1.57, 1.25 P-factor: 4.76, 1.04 hnhibitory effect: 0.79, 0.035 Selectivity	ratio: 1.05, 0.912 Sorptive affinity: – P-factor: 1.482, 1.71 1.482, 1.71 Inhibitory effect: 0.33, 0.414
g/g)	Pollut- ant 2 (Binary)	175.74	32
apacity (m	Pollut- ant 1 (Binary)	34.18	33.6
lsorption c	Pollut- ant 2 (Single)	182.17	54.6
Ac	Pollut- ant 1 (Single)	162.6	49.8
	Speed (rpm)	500	150
nditions	Hd -	ιņ	-
iental co	Initial concen tration (g/L)	0.6	0.1
Experin	Temp. (K)	238	298
	Adsor- bent dosage (g/L)	0.05	C Lý
lutant	(2)	2-Nitro- phenol	Ni(II)
Pol	(1)	Phenol	Phenol
Adsorbent		Biopoly- mer based biochar	Dry activated sludge
s.	No.	21.	5

Refer-	ences	[69]	[4]							
Inference		The presence of phenol, the adsorption was decreased. Competition for different sites/screening effect might have influenced the process.	The investigation showed no significant difference in adsorption of pheno- lic compounds in single and binary solute sys- tems which resulted in no interaction. However, the values of P-factor and inhibitory effect suggest negligible synergistic interaction. The governing force in the binary adsorp- tion process was π – π stak- ing interactions.							
Interaction	based on factors	Antagonis- tic (phenol, Ni) Antagonis- tic (phenol, Ni) Ni) Ni) Antagonis- tic (phenol, Ni) Ni)	No inter- action (2,4-DCP, BPA) No inter- action (2,4-DCP, BPA) Synergistic (2,4-DCP, BPA) Synergistic (2,4-DCP, BPA) BPA)							
Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 2 has less affinity w.r.t pollut- ant 1 Adsorption of pollutant 2 is enhanced and pollutant 1 is suppressed pollutant 1 is suppressed by pollutant 2	Adsorbent has a stron- ger affinity towards pollutant 2 Pollutant 1 and pollutant 2 have same affinity Adsorption of pollutant 1 is suppressed Adsorption of pollutant 1 is suppressed by pollutant 2							
Factors		Selectivity ratio: 0.56, 0.62 Sorptive affinity: -1.1, 1.21 P-factor: 1.533, 1.34 Inhibitory effect: 0.35, 0.282	Selectivity ratio: 0.64, 0.644 Sorptive affinity: -1.3, 1.32 P-factor: 0.989, 0.978 Inhibitory effect: -0.011,							
(g/g)	Pollut- ant 2 (Binary)	51.4	58.09							
apacity (m	Pollut- ant 1 (Binary)		36.99							
lsorption c	Pollut- ant 2 (Single)	71.6	56.80							
Ad	Pollut- ant 1 (Single)	44.3	36.6							
	Speed (rpm)	150	I							
nditions	Hď	4. Ú	0							
nental co	Initial concen tration (g/L)	0.1	I							
Experin	Temp. (K)	298	238							
	Adsor- bent dosage (g/L)	о Ю	0.005							
ollutant	(2)	Ni(II)	Bisphenol A (BPA)							
Pc	(1)	Phenol	2,4- DCP							
Adsorbent		Dry activated sludge	Flower-like covalent organic framework							
s.	No.	3	24.							

Refer-	ences	[41]	[02]
Inference		The value selectivity ratio and P-factor show that there is no interaction in binary system when humic acid modified bentonite was used. Ion exchange depict the adsorption of copper ion and hydro- phobic interaction caused adsorption of 2,4-DCP.	In the binary system, no competition between 2,4- DCP and lead ions was observed. This was due to sorption of pollutant occurred due to two inde- pendent mechanisms namely ion-exchange and solute partitioning.
Interaction	based on factors	Non-in- teraction (2,4-DCP, Cu) Non-in- teraction (2,4-DCP, Cu) Antag- onistic (2,4- DCP) Synergistic (Cu) Antag- onistic (2,4- DCP) Synergistic (Cu)	Non-un- teraction (2,4-DCP, Pb) Non-in- teraction (2,4-DCP, Pb) No inter- action (2,4- DCP) No inter- action (2,4- DCP) No inter- action (2,4- tric (Pb) No inter- action (2,4- tric (Pb)) Antagonis- tric (Pb)
Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 2 Pollutant 2 has less affinity wr.t pollutant 2 has ant 1 Adsorption of pollutant 1 is suppressed pollutant 1 is suppressed by pollutant 2	Adsorbent has a stron- ger affinity towards pollutant 2 Pollutant 2 has less affinity wr.t pollutant 2 has ant 1 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed Adsorption of pollutant 2 is suppressed by pollutant 1
Factors		Selectivity ratio: 0.7, 0.721 Sorptive affinity: -17.7, 25 P-factor: 1.022, 0.997 Inhibitory effect: 0.0211, -0.0035	Selectivity ratio: 0.78 Sorptive affinity: -1.05, 1.12 P-factor: 1.0, 1.009 Inhibitory effect: 00, 0.0091
pacity (mg/g)	Pollut- ant 2 (Binary)	23.07	08.
	Pollut- ant 1 (Binary)	16.22	69.2
sorption c	Pollut- ant 2 (Single)	22.99	68. 68
ΡЧ	Pollut- ant 1 (Single)	16.57	69.7
	Speed (rpm)	250	0 250
onditions	Hd u		7-10
mental c	. Initial conce tratio: (g/L)	0.1	0.1
Experi	Temp (K)	303	308
	Adsor- bent dosage (g/L)	4	10
llutant	(2)	Cu(II)	Pb(II)
Pc	(1)	2,4- DCP	2,4- DCP
Adsorbent		Bentonite modi- fied with humic acid (HAB)	Organo- clay
s.	No.	25.	26.

Refer-	ences	[12]	[2]
Inference		Adsorption of both the pollutants occurred on different sites. The adsorp- tion of Cr(VI) in presence of phenol was suppressed and the governing forces are weak van der Waals force and Steric Hindrance	The inhibition effect of phenol on catechol is observed. The negative value of inhibitory effect for phenol points out syn- ergistic interaction and the positive value for cat- echol indicates antagonis- tic interaction. The factors causing the interaction could be dipole moment and solubility.
Interaction	based on factors	Antagonis- tic (phenol, Cr) Antagonis- tic (phenol, Cr) Antagonis- tic (phenol, Cr) Cr)	opinctguance (phenol) Antag- onistic (catechol) Antag- onistic (phenol) Antag- onistic (phenol) Synergistic (phenol) Synergistic (phenol) Antag- onistic (catechol) Synergistic (phenol) (catechol) (c
Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 2 Adsorbent has a equal affinity towards pol- lutant 1 and 2 Adsorption of pollutant 2 is suppressed Adsorption of pollutant 2 is suppressed by pollutant 1	has a stron- ger affinity towards pollutant 1 Pollutant 2 has less affinity wr.t pollut- ant 1 Adsorption of pollutant 2 is suppressed Adsorption of pollutant 2 is suppressed by pollutant 1
Factors		Selectivity ratio: 0.56, 0.53 Sorptive affinity: 1.05, 1.05 P-factor: 1.001, 1.050 Inhibitory effect: 0.001, 0.047	ratio: 2.13, 1.324 Sorptive affinity: 3.7, 4 P-factor: 0.68, 1.09 Inhibitory effect: -0.48, 0.080
g/g)	Pollut- ant 2 (Binary)	50.58	6.27
apacity (m	Pollut- ant 1 (Binary)	28.15	13.36
sorption c	Pollut- ant 2 (Single)	53.13	6.82
Ad	Pollut- ant 1 (Single)	28.18	9.03
	Speed (rpm)	200	150
ditions	Hq	Γ	ц С
iental con	Initial concen- tration (g/L)	0.05/0.1	0.05-0.5
Experim	Temp. (K)	293	303
	Adsor- bent dosage (g/L)	Q	0.02
utant	(2)	Cr(VI)	Catechol
Pol	(1)	Phenol	Phenol
Adsorbent		Activated Carbon	Rice husk ash (RHA)
s.	No.	27.	

Refer-	ences										101	[0]								
Inference								In binary state, both pol-	lutants exhibited weaker	interaction with the adsor-	bent due to Intramolecular	attractions. Also, the value	of P-factor depicts the	same antagonistic interac-	tion.					
Interaction based on factors			Antagonis-	tic (phenol,	hydroqui-	none)		Antagonis-	tic (phenol,	hydroqui-	none)	Antagonis-	tic (Phenol,	hydroqui-	none)		Antagonis-	tic (phenol,	hydroqui-	none)
Findings from	factor		Adsorbent	has a stron-	ger affinity	towards	pollutant 2	Pollutant 2 has	less affinity	w.r.t pollut-	ant 1	Adsorption of	pollutant 2 is	enhanced and	pollutant 1 is	suppressed	Adsorption of	pollutant 1 is	suppressed by	pollutant 2
Factors			Selectivity	ratio: 0.74,	0.765			Sorptive	affinity:	1.26, 1.10		P-factor:	1.42, 1.32				Inhibitory	effect: 0.297,	0.271	
g/g)	Pollut- ant 2	(Binary)									17 010	10.012								
pacity (mg	Pollut- ant 1	(Binary)									1 1 1 0	77.701								
Adsorption ca	Pollut- ant 2	(Single)										67.667								
	Pollut- ant 1	(Single)										17.477								
	Speed (rpm)	4										700								
ditions	Hq											D								
ental cor	Initial concen-	tration (g/L)									10 10	C'0-T'0								
Experim	Temp. (K)	~									000	067								
	Adsor- bent	dosage (g/L)										cn.0								
utant	(2)			Hydro- Juinone																
Pollu	(1)										Diama I									
Adsorbent _										Grapnene : Jo		magnenc	nanocom-	pusue						
S.	No.										Ċ	-7.								

Refer-	ences	[55]	[23]					
Inference		The presence of Pb(II) enhanced the adsorption of phenol and synergistic effect was observed. The adsorption mechanism of adsorbents for lead and phenol was governed by electrostatic interactions, hydrogen binding and sur- face complexation.	Adsorption of phenol and phosphate ions increased in binary system. Some other factors causing increased adsorption were temperature and proper- ties of adsorbent.					
Interaction	based on factors	Synergistic (phenol) Antagonis- tic (Pb(II)) Synergistic (phenol) Antagonis- tic (Pb(II)) Antagonis- tic (Pb(II)) Antagonis- tic (Pb(II))	tic (Pb(II)) Synergistic (phenol, phosphate) - -					
Findings from	factor	Adsorbent has a stron- ger affinity towards pollutant 2 Pollutant 2 has less affinity w.r.t pollutant 2 has ant 1 Adsorption of pollutant 1 is enhanced and pollutant 2 is suppressed Adsorption of pollutant 2 is suppressed by	pollutant 1 Adsorbent has a stron- ger affinity towards pollutant 2 -					
Factors		Selectivity ratio: 0.981, 0.57 Sorptive affinity: 1.72, 1.93 P-factor: 0.697, 1.168 Inhibitory effect: -0.472,	0.144 Selectivity ratio: 0.48 Sorptive affinity: – P-factor: – Inhibitory effect: –					
g/g)	Pollut- ant 2 (Binary)	123.3	38 .5					
ipacity (m	Pollut- ant 1 (Binary)	121	18.82					
orption ce	Pollut- ant 2 (Single)	144	I					
Ads	Pollut- ant 1 (Single)	82.2	I					
	Speed (rpm)	150	150					
ditions	Hd		Ν					
ental cor	Initial concen- tration (g/L)	0.2	0.047					
Experim	Temp. (K)	318	293-					
	Adsor- bent dosage (g/L)	0.05	0.4					
lutant	(2)	Pb(II)	Phos- phate(V)					
Pol	(1)	Phenol	Phenol					
Adsorbent _		Zwitteri- onic adsor- bents	HDTMA bentonite					
s.	No.	30.	31.					

The higher or positive value of ΔIE_i suggests the adsorption capacity of component *i* is suppressed in the presence of *j* and the antagonistic interaction is observed in component *i*. The lower or negative value of ΔIE_i suggests the adsorption capacity of component *i* is enhanced in the presence of *j* and the synergistic interaction is observed in component *i*.

3. Mechanisms in adsorption system

The prominent mechanism of adsorption can be either follow physical or chemical processes. The interaction mechanism for the adsorption of the organic molecules could be due to hydrophobic effect, π – π interactions, filling of pores and electron–donor reactions. Further in case of inorganic molecules electrostatic attraction, ion-exchange, surface complexation, co-precipitation etc. may be responsible for adsorption process. Some of the prominent mechanisms are shown in Fig. 3.

3.1. Ion-exchange

The primary principle involved in this process is exchange of protons and ionised cations with dissolved salts on the surface of the adsorbent. A study conducted by Ru et al., on removal of organic molecules (phenol and aniline) reports that, the sorption of molecules is mainly attributed to combination of ion exchange, hydrophobic interaction [45]. Similar findings were reported by Dong et al. [46], while investigating adsorption of phenol and chromium on biochar.

3.2. Surface complexation

In this adsorption mechanism, the multi-atom formation interacts with metal to create a metal complex. For example, research conducted by Zong et al. [54] on competitive removal of cadmium and phenol attributes surface complexation as interaction factor responsible for adsorption of cadmium(II) ions. Similar results were also described by Wang et al. [55].

3.3. Precipitation and electrostatic interaction

It is one of the main mechanisms which is usually observed in mixture of inorganic pollutant. The mechanism involves formation of precipitate onto substrate surface. Electrostatic interaction involves attraction of organic/ inorganic compounds which are ionisable to charged surface of adsorbent by means of electrostatic interaction. For example an investigation carried out by Huang et al. [56], on competitive sorption of phenol and chromium found that adsorption of former followed partition phenomena and latter followed electrostatic interaction.

3.4. Partition onto uncarbonized area

Diffusion of adsorbate material into the porous uncarbonized part of adsorbent results in partition of uncarbonized area and this area is freely available to organic pollutant for interaction leads to adsorption [50]. According Zhang et al. [57], organic carbon (OC) components of biochar exhibit high sorbate partitioning of the atrazine contaminant.



Fig. 3. Different mechanisms in adsorption. (a) Ion-exchange [47], (b) surface complexation [48], (c) precipitation [49], (d) partition onto uncarbonized area [50], (e) filling of pores [51], (f) surface sorption [52], (g) electron donor and acceptor [46], and (h) hydrophobic effect [53].

3.5. Filling of pores

In this mechanism the organic contaminants, based on the polarity and nature and variant of adsorbent gets attached to the surface of porous adsorbent having mesopore (2–50 nm) and micropores (<2 nm). For example a study investigated by Singh et al. [58] found that, the adsorption of phenol and cyanide was attributed to filling of pores. The same was further proved from by analysing surface morphology by means of FESEM. According to Kasozi et al., the micropore filling process for the sorption of catechol was more effective than other adsorption methods [59].

3.6. Surface sorption

Sorption is defined combination of physical and chemical process by means of which a pollutant molecule accumulates within another phase or on the phase boundary of two phases. The process also means accumulation of a substance onto the substrate and absorption of substrate molecule within the bulk of solid surface. The physical process which involves establishment of chemical bond as result of metal ion diffusion into porous adsorbent. The results from the study investigated by Tahermansouri et al., confirm that both the external surface sorption and intraparticle diffusion dominated in the adsorption of phenol by carbon nanotubes [60].

3.7. Electron donor and acceptor (EDA) interaction

The EDA interaction mechanism is mostly applied to the adsorption of aromatic compounds on biochar. For example a study carried out by Dong et al. [46] of competitive removal of phenol and chromium proposes π – π interaction and electron donor–acceptor complex for phenol removal. Similar results were also reported by Al-Malack and Dauda [10], who investigated removal of phenol using activated carbon.

3.8. Hydrophobic effect

When organic compounds of hydrophobic and neutral nature are available, the mechanism of adsorption takes place by virtue of partitioning and hydrophobic interaction processes. In terms of energy requirement, hydrophobic interaction process works on lesser energy as compared to partition. For example, a study carried out by Rubin et al. [61], on sorption of phenolic pollutants using macro algae attributes hydrophobic effect as main interaction factor responsible for binding. Similar conclusions were drawn by Zhou et al. [8], who examined the removal of phenol using magnetic nanohybrids.

To summarise, in case of organic pollutants, the attraction of pollutant onto adsorbent was mainly dependent on the aromatic and/or unique functional groups and on the polarity of both adsorbent and the organic pollutants. The characteristic of intermolecular gravitation and electrostatic force between the adsorbent and the pollutant was the determining factor for the physical adsorption, while the chemical adsorption mainly depends on the chemical interactions between the adsorbent and the organic pollutant, which can be π - π bonds, hydrogen bonds and coordination bonds. However, for inorganic contaminants the combined effects of several interactions, such as electrostatic interaction, surface sorption, ion exchange, complexation, and precipitation, were often included in the possible adsorption process.

The mechanisms discussed above leads to furthermore interactions (SYI, ANI, NI). In Table 2 most investigations on simultaneous adsorption are identified and incorporated to discuss on interaction factors as mentioned in section 2.

Various studies on the interaction between the solute molecules in a binary solution have been published in the literature. The adsorption of molecules onto the adsorbent, the adsorption capacity of the adsorbate in single and binary solution mode, and the kind of interaction are all highlighted in this study. It also analyses the likely cause of the components' interaction. As a result, the review looks into the simultaneous adsorption of organic components (phenols/2,4-dichlorophenol) with other components in a binary system. The interaction effect among the components in the above investigation is determined from available standard relationships. The components involved in simultaneous adsorption are the adsorption capacity for each component, and the interaction mechanism are discussed below. Some of the significant findings from the literature reported in Tables 2 and 3 are discussed below.

In an investigation on multi-component adsorption, Nguyen et al. [35] used chemical and biologically modified chitosan beads (crosslinked) to remove phenol and nickel. The adsorption in binary systems increased in the second-component compared to their adsorption in single systems synergistic biosorption between phenol and Ni(II) was found. The positive value of the P-factor also imply synergism. The isotherm modelling showed that the modified Langmuir model could acceptably describe the simultaneous biosorption of phenol and Ni(II) from binary solutions, reflecting the nature of chemical adsorption. Chemical binding and coordination were found to be support mechanism for adsorption and synergism.

A study on competitive adsorption of phenol and resorcinol was carried out by Thakur et al., [19]. Lower value of η_i of resorcinol depicts a stronger affinity towards the adsorbent. The substitution of the hydroxyl group takes place with resorcinol. Thus, the π - π interaction leads to the adsorption of resorcinol with the adsorbent [19].

Aksu and Akpinar studied the performance of dried anaerobic activated sludge (DAAC) to adsorb phenol and chromium(VI) ions in a single and binary system. It was observed that the equilibrium uptakes of both the pollutants were changed due to the presence of another component. Adsorption isotherms were developed for single and binary component systems. The study also found that adsorption of phenol was inhibited in the presence of a higher concentration of Cr(VI). This could also be understood by the value of the interactions factor of modified competitive Langmuir isotherm, that is, η_{nh} and η_{Cr} , which is found to be 1.07 and 1.00, respectively. Here η_{ph} is greater than η_{Cr} which implies Cr(IV) has a greater affinity towards sludge than phenol [76]. Positive value of inhibitory effect depicts antagonistic interaction and factors such as competition on similar binding sites and screening effect could be responsible for such interaction. Some of the other reasons could be specific surface properties and some physicochemical properties of solution [40].

Simultaneous adsorption of phenol and cyanide on copper-infused activated carbon was carried out by Agarwal et al. [62]. It was investigated that the adsorption of phenol was increased nearly 27-fold (from 10.74 to 292.95 mg/g) in the presence of cyanide. The ratio of $(Q_{i,o})_b/(Q_{i,o})_s$ for phenol and cyanide was calculated as 27.27 and 1.10, respectively. This relation hence establishes the fact that phenol and cyanide show synergism in the process of adsorption. The degree of synergistic effect was not prominent in cyanide; however, it can be explained with the help of a inhibitory value as calculate in Table 2. Further, this implies the adsorbent had a greater affinity towards phenol. The process can be attributed to hydrogen bonding between hydrogen atoms from phenol and a lone pair of electrons of the nitrogen atom of cyanide [62].

Agarwal and Balomajumder conducted a similar adsorption study on phenol and cyanide using granulated activated carbon (GAC) as an adsorbent. It was investigated that the adsorption of phenol was comparatively very high in the presence of cyanide. The reason could be high surface area of GAC. Further, this can be demonstrated by the value of selectivity ratio which was found to be greater than one. This implies that phenol molecule had greater affinity towards adsorbent. The interaction between cyanide and phenol could be attributed to hydrogen bonding and intraparticle diffusion [11].

Singh and Balomajumder worked on simultaneous adsorption of phenol and cyanide using coconut shells as a source of adsorbent. The adsorption of phenol was enhanced in the presence of cyanide. The adsorption of phenol in the binary state was more than its single state which is supported by negative value of inhibitory effect. Thus, indicating the interaction to be the synergistic type of interaction. The phenomena governing both solutes' adsorption were chemisorption and weak van der waal's force of attraction [18].

A similar study was investigated by Singh and Balomajumder, where *Pseudomonas putida* was immobilised on coconut shell activated carbon. The study found an increase in the adsorption of phenol from 18.19 to 349.86 mg/g in the presence of cyanide, indicating synergism. The values also suggest that phenol had a greater affinity towards the adsorbent. In addition, this can be further explained by selectivity ratio 349.86/18.19=160.49, which is greater than one indicating a greater affinity of phenol than cyanide. Another reason for the higher adsorption of phenol could be initial concentration [63].

Singh and Balomajumder used sugarcane bagasse as a source of adsorbent to remove phenol and cyanide. The adsorption capacities of adsorbent were predicted and were 94.25 and 5.46 mg/g for phenol and cyanide, respectively. The greater affinity of phenol could be due to rise in temperature. This can be explained using sorption affinity $S_{(ij)}$ factor. Here the calculated value was found greater than one [76], that is, 1.497 (phenol) and 1.26 (CN). The study shows a synergistic interaction mechanism involved in the binary adsorption process [58].

Further, a study conducted by Ha and Vinitnantharat [17] used biological activated carbon by considering phenol and 2,4-dichlorophenol (DCP) in a single and binary system. The adsorption capacity of activated carbon for both phenol and DCP decreased, indicating antagonistic behaviour. This finding can be supported by positive value of inhibitory effect. Further, the reason for such interaction could be mutual interaction and molecular size. Components in binary system were inhibited mutually. It was found that degree of inhibition was profound in DCP, which was also the reason for the fall in the adsorption capacity of DCP.

Istratie and Stoia, in their research, adsorption of methyl orange (MO) and phenol used magnetic iron oxide-based nanocomposites (MNP3). The adsorption of MO was higher compared to phenol but both components followed synergistic interaction. Electrostatic and non-electrostatic interactions were among the governing forces of the synergistic interaction. However, both electrostatic and non-electrostatic forces governed methyl orange adsorption, and phenol adsorption was due to non-electrostatic forces. Though the interaction behaviour was synergistic for both, it was found that MO had a greater affinity towards the adsorbent in the presence of phenol. This can be explained by the selectivity ratio and P-factor [20].

In work reported by Wang et al. on the adsorption of methyl red and methylene blue onto biochar, the adsorption of the molecules was governed by size and ionisation of molecules. Variation in the above factors at equilibrium and different isotherm were used to determine the interaction between the molecules. A significant rise in adsorption capacity of 2,4-DCP and Cr(VI) was observed. The synergistic interaction was due to the bridge interactions which created additional new sites for both pollutants [34].

In some of the studies, it was possible to explain the interaction behaviour of pollutants using competitive models.

In the present study, rice husk ash (RHA) was used as an adsorbent for individual and simultaneous adsorption of phenol and resorcinol. Langmuir, Freundlich and Redlich–Peterson showed a similar fit with regard to the mono-component adsorption data and Extended Langmuir best depicted the binary component experimental data. The affinity of RHA was found to be greater for phenol than that for resorcinol in both single component and binary solutions [19].

Aksu and Akpinar used dried activated sludge to abate phenol and chromium from synthetic wastewater and modelled using modified competitive Langmuir isotherm (MCLI). The model predicted the value η_i for phenol and was greater than that of chromium, which implies higher affinity of adsorbent towards chromium. Further, it indicates that chromium adsorption was greater than phenol in the binary state. Antagonistic pollutant behaviour during competitive removal can be proven using a modified competitive Langmuir isotherm model [40].

Research carried out by Singh and Agarwal used coconut husk-activated carbon (CHAC) as an adsorbent. It was found that phenol adsorption in the binary system was more than cyanide, implying antagonistic behaviour. It may be due to higher surface area of CHAC with active sites

Table 3 Interaction based on multicomponent Adsorption isotherm

Sl.	Adsorbent	Pollutant <i>i</i>	Pollut-	Interaction parameter			Inference from	Interaction	Refer-
No			ant j	θ _{ij} (SRS)	η _i (MLI)	η_{PRi} (MRP)	factors	effect	ences
1.	Dried anaerobic activated sludge	Phenol	Cr(VI)	-NA-	Phenol – 1.07 Cr(VI) – 1.00	-NA-	Lower affinity of pollutant 1 to the adsorbent	Antagonistic (phenol, Cr)	[40]
2.	Copper impreg- nated GAC (Cu-GAC)	Phenol	CN	-NA-	Phenol – 0.71 CN – 0.76	-NA-	Lower affinity of pollutant 2 to the adsorbent	Synergistic (phenol, CN)	[62]
3.	Coconut shell AC	Phenol	CN	-NA-	Phenol – 3.06 CN – 0.52	Phenol – 0.13 CN – 0.03	Lower affinity of pollutant 1 to the adsorbent	Synergistic (phenol, CN)	[18]
4.	Sugarcane bagasse	Phenol	CN	-NA-	Phenol – 0.37 CN – 0.34	Phenol – 1.29 CN – 1.37	Lower affinity of pollutant 1 to the adsorbent, lower affinity of pollutant 2 to the adsorbent	Synergistic (phenol, CN)	[58]
5.	Coconut husk AC	Phenol	CN	-NA-	Phenol – 0.36 CN – 0.70	Phenol – 0.20 CN – 0.23	Lower affinity of pollutant 2 to the adsorbent	Antagonistic (phenol, CN)	[64]
6.	Chitosan beads (chemical and bio- logical modified)	Phenol	Ni(II)	-NA-	Phenol – 0.11 Ni(II) – 0.27	-NA-	Lower affinity of pollutant 2 to the adsorbent	Synergistic (phenol, Ni)	[35]
7.	<i>Bacillus</i> sp. immo- bilis onto tea waste biomass	Phenol	Cr(VI)	-NA-	Phenol – 14.14 Cr(VI) – 41.9	Phenol – 0.05 Cr(VI) – 0.82	Lower affinity of pollutant 2 to the adsorbent	Synergistic	[65]
8.	Tea waste biomass	Phenol	Cr(VI)	-NA-	Phenol – 6.03 Cr(VI) – 0.25	-NA-	Lower affinity of pollutant 1 to the adsorbent	Synergistic	[2]
9.	GAC	Phenol	Resor- cinol	-NA-	Phenol – 0.44 Resorcinol – 0.5	-NA-	Lower affinity of pollutant 2 to the adsorbent	No interac- tion	[42]
10.	RHA	Phenol	Resor- cinol	Phenol – 1 Resor- cinol – 2.11	Phenol – 1.097 Resorcinol – 1.0768	Phenol – 0.39 Resorcinol – 0.77	Pollutant 2 has weaker inhi- bition on the adsorption of pollutant 1, lower affinity of pollutant 1 to the adsorbent, lower affinity of pollutant 2 to the adsorbent	Synergistic (phenol, resorcinol)	[19]
11.	Chitosan	Phenol	CN	-NA-	Phenol – 0.67 CN – 1	-NA-	Lower affinity of pollutant 2 to the adsorbent	Synergistic	[74]
12.	Peanut shell	Phenol	Cd(II)	-NA-	Phenol – 0.63 Cd(II) – 0.68	Phenol – 1.58 Cd(II) – 1.60	Lower affinity of pollutant 2 to the adsorbent	Synergistic	[75]

for phenol. The adsorption study was also modelled using a modified competitive Langmuir isotherm and modified Redlich–Peterson isotherm (MRPI). Both exemplified the interaction phenomena as antagonistic. The reason could be interaction parameter for phenol was lesser than cyanide which implies phenol showing greater affinity [64].

An investigation on the removal of chromium and phenol by using *Bacillus* sp. immobilised tea waste biomass was carried out by Gupta and Balomajumder. The study was modelled for competitive isotherms, i.e.; MCLI, extended Langmuir, extended Freundlich, non-modified competitive Redlich–Peterson etc. The extended Langmuir best fitted for phenol and non-modified competitive Redlich–Peterson best fitted for Cr(VI) sorption. It was found that chromium is synergistic in the presence of phenol, and out of all the models, MCLI can determine the interaction [65].

Tea waste activated carbon was used as an adsorbent to remove phenol and chromium from a binary solution. It was observed that the adsorption of phenol was less than chromium. The possible reason may be functional groups, the structure of the adsorbent, size, shape and molecular nature of solutes present. It was also explained that phenol gets attached to the adsorbent surface through weak van der Walls forces and p–p interaction, and chromium gets attracted to the functional groups present on the adsorbent. The synergic type of behaviour was observed in the binary system. Though the binary system experienced synergetic behaviour, phenol inhibition was higher in the presence of chromium ions. This could be explained using the interaction factor of MCLI [2].

Kumar et al. [42] investigated co-adsorption of phenol and resorcinol on GAC. Three equilibrium isotherm models, non-modified Langmuir model, modified Langmuir model and extended Freundlich, model, were considered for multi-component adsorption. The effect of low and high initial concentration was studied for different operating conditions. It was found that there was the least interaction between the pollutant in binary system. This phenomena can be understood by interaction factor n_i are n_j of MCLI, which are approximately equal to each other.

Agarwal and Sengupta, in their study, analysed the potential of modified chitosan for removal of phenol and cyanide in binary system. Modified and non-modified isotherms models were fitted to the experimental data of the binary adsorption system. It was that extended Freundlich best depicted the adsorption in a binary system. It was found that adsorption of phenol was enhanced in the presence of chromium [74].

The present study investigates the potential of peanut shells as adsorbent for the removal of phenol and cadmium(II) from a bisolute wastewater system. Langmuir–Freundlich model best fitted with regard to the mono-component adsorption data and modified Langmuir best depicted the binary component experimental data. The adsorption capacity of phenol and cadmium was almost the same in single and binary states, indicating no interaction. This may be because phenol adsorbed rapidly, making it non-interactive. And the non-interaction could also be explained from the interaction parameters present in MCLI and MRPI [75]. In one of the studies, rice husk incorporated with iron particles was used as adsorbent. Multi-component models modified, and non-modified competitive isotherms were applied to investigate the adsorption of phenol and Cr(VI) in a single and binary state. The adsorption capacity for phenol and Cr(VI) was enhanced in binary systems. This was because phenol acts as an electron donor, enhancing the Cr(VI) adsorption. Extended Freundlich for both phenol and Cr(VI) for the multi-component system was agreed better with the experimental results. The results show that, adsorption of Cr(VI) is enhanced from 7.55 mg/g in single system to 36.38 mg/g in binary system implying synergistic interaction [77].

4. Discussion and conclusions

This review papers gives a brief idea of possible interaction of phenolic compounds with other pollutants which can be either organic or inorganic nature. From the studies tabulated in Table 2, it was evident that, phenol exhibited synergistic behaviour in presence of pollutants such as resorcinol, cyanide, 2,4-DCP, methyl orange, catechol, aniline, nickel, chromium, lead and cadmium. Further, in some of the studies, antagonistic interaction of phenol was observed in presence of cyanide, methylene blue, 2,4-DCP, chromium, cadmium, nitrophenol, nickel and hydroquinone. In addition, few cases of non-interaction of phenolic pollutant was observed in presence of resorcinol and mercury. The reason for such interaction could be charge on the adsorbing molecule and pH of the sample, possible overlap of surface area and many other mechanisms as discussed in section 3.

An initial objective of the review was to identify a relation between interaction and the factors. It was possible to correlate selectivity ratio, P-factor and inhibitory effect to adsorption behaviour and the interactions in simultaneous adsorption studies considered in the review but same was not convincing with sorption affinity. In case of sorption affinity, the concept of relating the interaction and factor did not hold good. However, it holds good only for same initial concentration of both the solutes.

Based on the values of selectivity ratio and sorption affinity for a binary system, a brief idea of the degree of affinity of pollutants towards the adsorbent can be understood. However, by considering P-factor and inhibitory effect for a single and binary adsorption system would aid in identifying adsorption of the prominent pollutant (inhibition/enhancement) in the binary system. Further, by analysing the values of P-factor and inhibitory effect, it was possible to generalise and distinctly arrive at interaction (synergistic, antagonistic, non-interaction) of a solute in binary adsorption system.

The relation between P-factor, inhibitory effect, and interactions (SI, AN, NI) were identified by following convention as mentioned in previous section 2. Further, it was also possible to find the interactions in the binary system using modified competitive isotherms such MCLI, MCRPI and SRS equations. The interaction factors present in MCLI and MCRP, that is, η_i and η_{RP_i} were used to depict the nature of the affinity of pollutant, and in the SRS equation the

factor θ_{ij} was used to determine the inhibition of pollutant i on j.

Prior to this review, adsorption capacity of adsorbent in binary system was only the sole way to recognize interaction. In this review, a new understanding on interaction factors would help researchers, to improve predictions of the interaction mechanism in binary adsorption system. The findings of this review would also provide the following insights for future research. More studies on competitive adsorption of phenol and 2,4-DCP would help us to establish a greater degree of accuracy on this matter. Further research could also be conducted to determine the effectiveness of linking interaction mechanism to the factors as discussed in previous sections and this would be a fruitful area for further work. Studies involving adsorption of ternary pollutants can be investigated and modelled. Further, the parameters related to interaction can be investigated. No studies were found, which can relate Gibb's free energy, enthalpy and entropy to interaction between molecules.

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240

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