

Adsorption of aluminum from aqueous solutions utilizing waste polyethylene terephthalate derived activated carbon

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Received 7 May 2023; Accepted 27 August 2023

ABSTRACT

This study aims to adsorb aluminum (Al^{3+}) from aqueous solutions through batch experiments utilizing activated carbon (AC). Waste polyethylene terephthalate was carbonized in a N_2 environment and then chemically activated with 1 M KOH to produce AC. Fourier-transform infrared spectroscopy, surface area analysis, as well as scanning electron microscopy were used to analyze the AC. The pertinent adsorption behaviors of AC for Al^{3+} have been studied, including adsorption isotherms, adsorption kinetics, the effect of initial Al^{3+} concentration, adsorbent dosage, contact time, as well as pH, etc. According to the results, AC has the maximum ability for adsorbing Al^{3+} from aqueous solution (>95%). The findings show that the pseudo-second-order equation can accurately predict the adsorption kinetics of AC for Al^{3+} , and that the experimental data and Langmuir isotherm model match well, demonstrating homogeneous monolayer adsorption. According to the research findings, AC may be used as a sort of very effective adsorbent in the treatment of wastewater, which would enable a simple approach for the removal of Al^{3+} .

Keywords: Activated carbon; Aqueous solution; Adsorptive removal; Aluminum

1. Introduction

After oxygen and silicon, aluminium (Al^{3+}) is the third most common element in the Earth's crust. It is a poisonous metal that is not required for human life [1]. Al^{3+} is also contained in the deposited particulate matter [2]. The excessive use of aluminum salts (alum) in chemical treatment processes in water treatment plants as a coagulant to remove organic matter, microorganisms, color, as well as turbidity [3,4]. Al^{3+} can cause neurological disorders, including Parkinson's Dementia, amyotrophic lateral

sclerosis, and Alzheimer's disease (AD) [5,6]. The World Health Organization estimates that the number of people infected with the AD disease will reach 9 million by 2030 [1].

Much effort has been done on wastewater treatment during the last few years [7]. Diverse methods are created to cater the elimination of toxic elements as of wastewater, which incorporate membrane inlet mass spectrometry method [8], aerobic granule in continuous-flow [9] reduction, chemical precipitation, chemical reduction/oxidation, electrochemical treatment, photocatalytic reduction, membrane separation, ion exchange, coagulation or flocculation,

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chemical precipitation, complexation, filtrations, electrochemical precipitation, solid phase adsorption, biosorption, adsorption [10–24], biological denitrification treatment [25], photoinduced reaction [26], direct hydration reaction [27]. The majority of these approaches, however, are hindered by issues like expensive startup and maintenance costs and unclear working mechanisms [28].

Adsorption is an important wastewater treatment technique that is widely used to remove toxic metals from wastewater [29–37]. Assortment of adsorbents utilized for this reason incorporates chemically treated biosorbents [38,39], novel coagulant [40], Fe-doped and terephthalaldehyde-modified carbon nitride [41], RuPt/AC bimetallic catalysts [42], polyacrylonitrile/Na-Y-zeolite composite [43], biochar/bentonite/waste polyethylene terephthalate as well as biochar/bentonite/waste polystyrene [29], *Phoenix dactylifera* coir wastes [44], magnetic chitosan grafted with Schiff's base polymer [45], L-cystein-modified montmorillonite-immobilized alginate nanocomposite [46], microalgae [47], synthetic chelating resin [48], poly(methyl methacrylate)-grafted alginate/Fe₃O₄ nanocomposite [49], various other adsorbents [50], biosorbents [51], montmorillonite clay [52], ligand embedded nano-conjugate adsorbent [53], green synthesized nanoengineered materials [54], bamboo charcoal [55], nanocomposites [56], non-thermal plasma catalysis [57], green synthesis of silver nanoparticles [58], ligand based composite material [59], metal/mineral-incorporating materials [60], iron oxide-impregnated dextrin nanocomposite [61], ether based mesoporous adsorbent [62], magnetic glycine-modified chitosan [63], papaya peel carbon [64], ligand functionalized organic-inorganic based novel composite [65], thiourea-formaldehyde resin and its magnetic derivative [66], thin-film composite nanofiltration membrane [67], phthalide-derived analogues [68], dimeric salophen platinum(II) complexes [69], chitosan grafted polyaniline [70], modified conjugate material [71], CuO and Cu(OH)₂ embedded chitosan [72], waste rubber tires [73], ligand doped conjugate adsorbent [74], magnetic glycidyl methacrylate resin [75], iron based metal organic framework [76], nano zero-valent iron [77], natural wood-derived charcoal embedded with bimetallic iron/cobalt sites [78], element sulfur-based autotrophic denitrification constructed wetland [79], novel optical adsorbent [80], ligand based efficient conjugate anomaterials [81], novel facial composite adsorbent [82,83], nanoadsorbent [84], MoS₂/Fe₃O₄/aminosilane/glycidyl methacrylate/melamine dendrimer grafted polystyrene/poly(N-vinylcaprolactam) nanocomposite [85], 3D polymer nanomagnetic (3D/GO/Fe₃O₄) [86], and graphene oxide modified with 4-aminodiphenylamine [87].

Over the past 50 y, plastic usage has dramatically increased over the world. The growing manufacturing and use of plastic is threatening the environment because of its improper dumping and waste management. A variety of methods are offered for dealing with expanding plastic solid waste, including land filling, incineration, recycling, and its conversion into liquid or gaseous fuels [29]. After being evaluated critically, each of these methods has been found to have certain drawbacks. However, using plastic wastes in carbonized form as adsorbents to remove various hazardous chemicals from wastewater is a more feasible and environmentally beneficial method of reducing water

pollution. In this study, activated carbon (AC) has been successfully prepared by a simple and low-cost method. The structural characteristics of AC and the adsorption state of Al³⁺ were studied. Through batch studies, the effectiveness of AC adsorption was evaluated and its adsorption behavior on Al³⁺ was carefully examined, including the effects of initial Al³⁺ concentration, temperature, contact time, pH, adsorbent dosage, adsorption isotherm, and adsorption kinetics. This study makes a significant contribution to the removal of Al³⁺ from wastewater and establishes a new benchmark for reducing environmental pollution by reducing solid waste. This in turn enhances the aesthetic value of the local environment.

2. Material and methods

2.1. Sample collection and preparation

Samples of waste polyethylene terephthalate (wPET) were collected from the municipal solid wastes. After being cleansed with distilled water, the wPET samples were allowed to dry at room temperature before being shredded into tiny flakes. The water that was used in this study was deionized, and all chemicals and reagents were of analytical grade and used without additional purification. Stock solutions of Al³⁺ (1,000 mg/L) were prepared by dissolving its salt aluminium nitrate (Al(NO₃)₃) in 1,000 mL of deionized water, and the experimental solutions of the desired concentration were prepared by diluting the stock solution with deionized water.

2.2. Synthesis and characterization of activated carbon

Samples of wPET were collected from the street waste bins. The wPET samples were washed, dried, and shredded to small pieces. The waste plastic samples were converted into AC by carbonization in a specially designed steel reactor under inert atmosphere at high temperature. The steel reactor containing 5 g of wPET was placed in the tube furnace, linked to the nitrogen supply, and heated for 1 h at 600°C as indicated in Fig. 1. The AC was removed from the reactor and chemically activated by treating it alternately with 1 M KOH in accordance with the method described in the literature [83,88]. The solid carbon was finally ground to fine powder and passed through a sieve with pore size 250 μm and then stored in glass vials for further study.

Scanning electron microscopy (SEM), Fourier-transform infrared spectroscopy (FTIR), and surface area were used to characterize the AC that was synthesized in the laboratory. Using a scanning electron microscope made by JEOL (Tokyo, Japan), the surface morphology of the AC was investigated. A Nicolet iS10 FTIR spectrometer (Cary 630, Agilent Technologies, USA) with a scanning range of 4,000–400 cm⁻¹ was used to measure the FTIR spectra. The average pore size, total pore volume, and specific surface area were calculated using a Quantachrome, USA, model NOVA 2200e porosimetry system and surface area.

2.3. Al³⁺ adsorption measurements of AC

Adsorption of Al³⁺ from aqueous solution over AC was studied in batch mode adsorption experiments. About

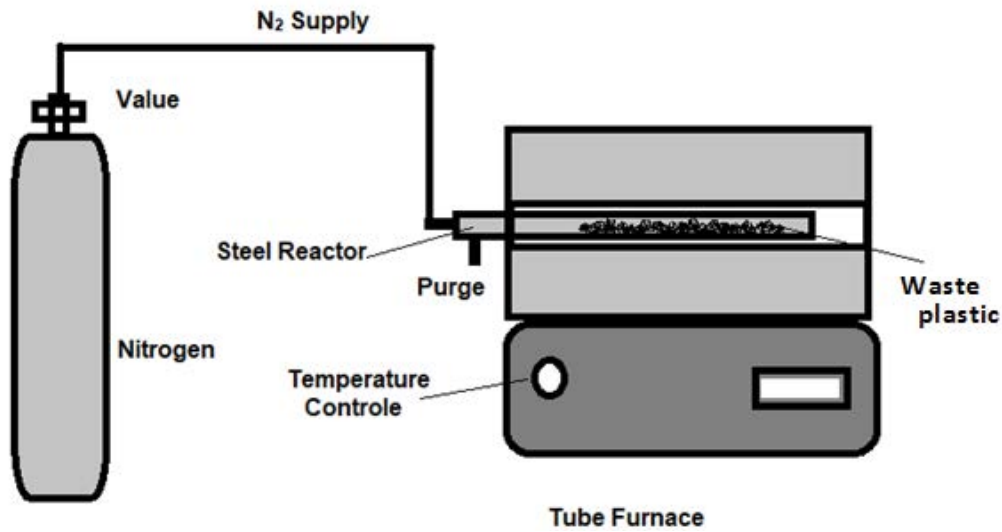


Fig. 1. Diagrammatic presentation of activated carbon in the lab.

100 mL of aqueous solution containing Al^{3+} was taken in 500 mL Erlenmeyer flasks at $24^\circ\text{C} \pm 2^\circ\text{C}$, a definite amount of AC adsorbent were added to it and stirred on a magnetic stirrer with 150 rpm speed until equilibrium was reached. Several variables were measured in the study, including dosage of AC (0.05, 0.10, 0.15, and 0.20 g), contact time (15, 30, 45, 60, and 90 min), Al^{3+} concentrations (10, 20, and 30 mg/L) and pH of solutions (1–7), pH was adjusted by adding 0.1 M NaOH to the solution as required.

Atomic absorption spectrophotometer (AAAnalyst 700, PerkinElmer, USA), operating in flame mode, was used to quantify the amount of Al^{3+} [89]. Calculations of the corresponding removal percentage (%) and adsorption capacity (q_e) of AC (mg/g) were made using Eqs. (1) and (2), respectively [29,89].

$$\%MR = \frac{(C_o - C_i)}{C_o} \times 100 \quad (1)$$

where C_i and C_o represent the final and initial concentration (mg/L).

$$q_e = \frac{[(C_o - C_i) \times V]}{m} \quad (2)$$

where V is the volume (L) of the stock solution and m is the adsorbent weight (g). For all tests carried out, the mean values only of triplicate were presented.

To assess the fitness of the isotherm, as well as the kinetic equations to the experimental data, the root mean square error (RMSE) was utilized to determine the isotherm along with kinetic constants. RMSE can be defined as:

$$RMSE = \sqrt{\frac{1}{N-2} \sum_{i=1}^N (q_{e,exp} - q_{e,calc})^2} \quad (3)$$

where N is the number of observations and the subscripts “calc” and “exp” show the calculated and experimental

values in the experimental data. The smaller the RMSE value, the better the curve fitting [90].

Adsorption kinetics was investigated by applying different rate equations [91–93]. Construction of linear plots was studied by applying pseudo-first and pseudo-second-order kinetic models. The pseudo-first-order rate constant was calculated using the following Eq. (4) [29,89]:

$$\ln(q_e - q_t) = \ln q_e + k_1 t \quad (4)$$

The pseudo-second-order rate constant was determined using following equation [29,89]:

$$\frac{t}{q_t} = \frac{t}{q_e} + \frac{1}{K_2 q_e^2} \quad (5)$$

Different adsorption isotherms were also applied to investigate the nature of adsorption. The Langmuir isotherm equation was used in the following form [29,89]:

$$\frac{C_e}{q_e} = \frac{1}{KQ_{max}} + \frac{C_e}{Q_{max}} \quad (6)$$

The Freundlich adsorption isotherm was applied as the following equation [29,89]:

$$\ln q_e = \ln K_f + \frac{1}{n} \ln C_e \quad (7)$$

where q_e represents the concentration adsorbed on the composite (mg/g), C_e is the equilibrium concentration of Al^{3+} (mg/L), while the constant and the adsorption coefficient are represented by n and K_f for the model, respectively.

3. Results and discussion

3.1. Characterization of AC

Specific surface area analysis, SEM, and FTIR were used to depict AC. SEM observations of the AC are shown

in Fig. 2a and b, where it is clear that the surface of AC is relatively rough and porous, and it is made up of KOH particles, many of which have micropores that could serve as additional Al^{3+} adsorption sites (Fig. 2a). After adsorbing Al^{3+} , the surface of AC appeared glossy, smooth, and had a closed pore structure, as shown in Fig. 2b. This change may be the result of physicochemical interactions between the functional groups on the surface of AC and Al^{3+} .

Fig. 3a and b show FTIR spectra of wPET and AC. The characteristic infrared absorption bands of wPET (Fig. 3a) at 691 cm^{-1} (C–H, stretching), 722 cm^{-1} (C–H, stretch), $1,009\text{ cm}^{-1}$ (C–O–C, stretching), $1,100\text{ cm}^{-1}$ (C–O, stretching vibration), $1,240\text{ cm}^{-1}$ (C=C, stretch), $1,409\text{ cm}^{-1}$ (CH_2 stretching), $1,594\text{ cm}^{-1}$ (C=C, stretching), and $1,696\text{ cm}^{-1}$ (C=O, stretch) are clearly observed [34]. The FTIR spectrum of AC (Fig. 3b) showed that the AC had functional groups of wPET. The noticeable absorption peaks at 722 cm^{-1} (C–H, stretch), 779 cm^{-1} (C–H, stretch), 872 cm^{-1} (C–H, stretch), $1,017\text{ cm}^{-1}$ (C–O–C, stretching), $1,100\text{ cm}^{-1}$ (C–O, stretching vibration), $1,250\text{ cm}^{-1}$ (C=C, stretch), $1,406\text{ cm}^{-1}$ (CH_2 stretching), and $1,681\text{ cm}^{-1}$ (C=O, stretch) [94]. These results suggest that wPET-derived AC contains various oxy-functional groups that may be due to surface oxidation of AC. The point of zero charge (PZC) was found to be 4.25 (Fig. 4), and its surface will be negatively and positively charged at pH values above and below the PZC.

Surface properties such as total pore volume (V_t), mean pore radius (R_p), Brunauer–Emmett–Teller (BET) surface area (S_{BET}), and Barrett–Joyner–Halenda (BJH) surface area (S_{BJH}) were calculated for AC. Results show that wPET-AC exhibits S_{BET} of $65.35\text{ m}^2/\text{g}$, S_{BJH} of $11.02\text{ m}^2/\text{g}$, V_t of $0.009\text{ cm}^3/\text{g}$ and an R_p of 15.02 \AA . After adsorption, the surface properties are S_{BET} $2.40\text{ m}^2/\text{g}$, S_{BJH} $1.95\text{ m}^2/\text{g}$, V_t $0.0001\text{ cm}^3/\text{g}$ and R_p 2.90 \AA . The surface properties decreased after adsorption, confirming the adsorption of Al^{3+} .

3.2. Effect of process parameters

We evaluated the adsorption under various conditions of pH, Al^{3+} concentration, adsorbent dosage, and contact time in order to determine the optimal conditions necessary for maximum Al^{3+} adsorption.

3.2.1. Effect of contact time

The tests were conducted at various adsorption times, that is, 15, 30, 45, 60, and 90 min, to determine the effect of adsorption time on the percentage adsorption of Al^{3+} on the AC. The results are displayed in Fig. 5, which demonstrates that the adsorption of Al^{3+} increases linearly with an increase in adsorption duration and reaches its maximum within 90 min. However, an increase in adsorption

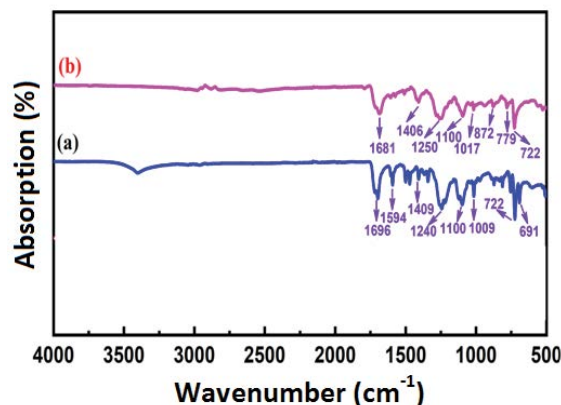


Fig. 3. Fourier-transform infrared spectra of (a) waste polyethylene terephthalate and (b) activated carbon.

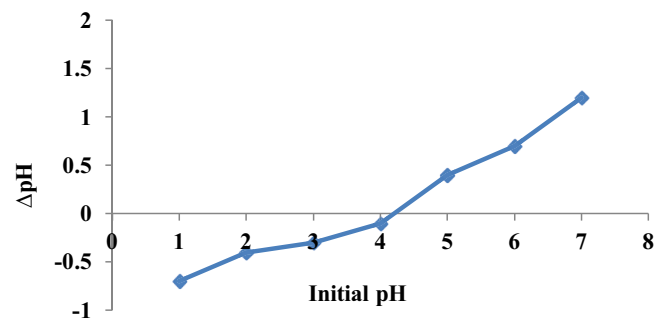


Fig. 4. Determination of point of zero charge of activated carbon.

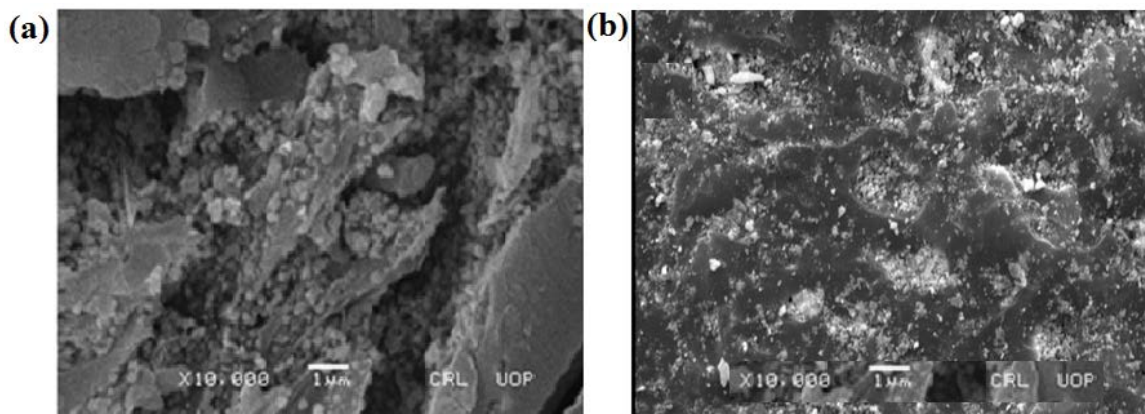


Fig. 2. Scanning electron microscopy images of activated carbon before Al^{3+} adsorption (a), and activated carbon after Al^{3+} adsorption (b).

beyond this point has no further effect on adsorption. These results suggest that the adsorption rate initially increases with the number of accessible adsorption sites on the AC surface as the adsorption time increases. The voids were filled within 90 min, and thereafter, the adsorption became steady [29,88,89,94]. These findings led to the decision that 90 min was the ideal adsorption time.

3.2.2. Effect of initial Al³⁺ concentration

At various initial Al³⁺ concentrations, the adsorption of Al³⁺ by AC was studied. Fig. 6 displays the data. This implies that the rate of Al³⁺ adsorption on AC increases as the initial Al³⁺ concentration rises from 10 to 30 mg/L. This results from Al³⁺ adhering to the surface-active sites of AC when the initial Al³⁺ concentration is increased. As Al³⁺ concentration rises, the existence of a large concentration gradient produces a more grounded primary impetus that uses active sites to enhance the amount of AC that can be absorbed per unit mass while overcoming mass exchange resistance [29]. Three removal mechanisms, chelation, ion exchange, and electrostatic attraction mechanism existed simultaneously when the AC was used [28,94–96].

3.2.3. Effect of AC adsorbent dosage

The amount of adsorbent used also has a big impact on how much toxic metal is absorbed. The ability of the adsorbent to be used efficiently while maintaining high adsorption effectiveness is typically necessary for the investigation of the optimal adsorbent dosage. The

necessary adsorbent dosage was adjusted between 0.05 and 0.20 g/L in order to find the ideal value for Al³⁺ adsorption on AC. The adsorption effect of AC adsorbent dose from an aqueous medium is depicted in Fig. 7. As can be shown in Fig. 7, the adsorption capacity of the adsorbent for Al³⁺ decreases as the adsorption rate rises. This can be attributed to Al³⁺ competitive adsorption tendency to disperse across the wide surface of the adsorbent, which lowers the adsorption capacity per unit mass of AC [98]. The removal rate is almost high when the adsorbent dosage is 0.20 g/L. Accordingly, 0.20 g/L of AC would be utilised as the ideal adsorbent dosage in the upcoming studies based on the aforementioned experimental findings in order to thoroughly examine its adsorption ability.

3.2.4. Effect of pH

Investigating the ideal pH for Al³⁺ adsorption is crucial because the pH of the investigated solution system is crucial for both the state of the adsorbate and the surface chemical characteristics of the adsorbent. It has been studied how well AC adsorbs Al³⁺ in the pH range of 2–6. The related adsorption tests were carried out at 25°C and a pH range of 1–7 in order to assess the impact of pH on the adsorption of Al³⁺ on AC and to optimize a specific pH for maximum adsorption efficiency. Fig. 8 displays the final result. Fig. 8 clearly illustrates how the variation trends are similar in nature and how the adsorption capacity quickly rises with increasing pH. The results showed that the adsorption process was highly dependent on pH, and

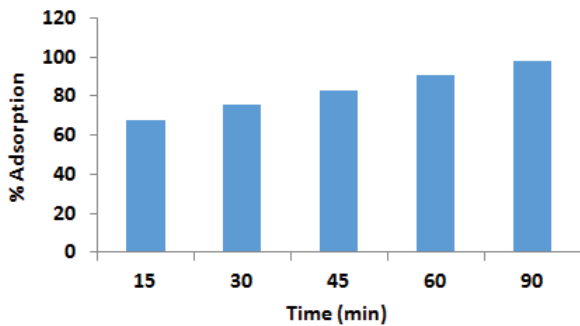


Fig. 5. Effect of contact time (15, 30, 45, 60, and 90 min) on % adsorption of Al³⁺ over activated carbon.

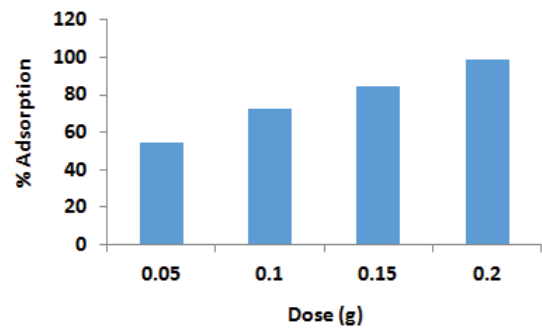


Fig. 7. Effect of dose (0.05, 0.10, 0.15, and 0.20 g) on % adsorption of Al³⁺ over activated carbon.

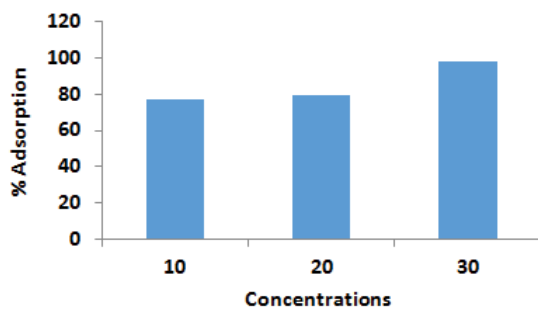


Fig. 6. Effect of concentrations (10, 20, and 30 mg/L) on % adsorption of Al³⁺ over activated carbon.

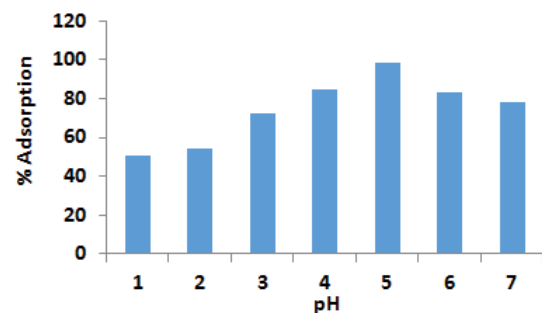


Fig. 8. Effect of pH (2, 3, 4, 5, 6, and 7) on % adsorption of Al³⁺ over activated carbon.

increasing the pH from 1 to 5 increased the adsorption of Al^{3+} . However, adsorption decreased if the pH value was increased, so pH 5 was chosen for further study. The PZC value of AC is 4.25, indicating that its surface is positively and negatively charged at pH values below and above the PZC, respectively. The data showed that there was no significant difference in the adsorption of Al^{3+} in the pH range of 1–4, which may be because there are too many H^+ ions competing with Al^{3+} for the same adsorption sites [29,94]. Therefore, the subsequent Al^{3+} adsorption experiments are conducted in the solution condition of pH = 5.0.

3.3. Adsorption kinetics

The kinetic data are crucial for choosing the best conditions while developing the AC adsorption system. Two well-known kinetic models, pseudo-first-order and pseudo-second-order equations, were applied to the data for this purpose. According to Table 1, the obtained coefficient of the pseudo-second-order model was high for the AC, indicating that this model had greater conformance than the pseudo-first-order model at describing the adsorption behavior of AC for Al^{3+} . The adsorption kinetics are thus better described by the pseudo-second-order model, which also suggests that chemisorption mostly governs the adsorption of Al^{3+} over AC.

3.4. Isotherm study

We can determine how the adsorbates are distributed between the solution and the adsorbent when the adsorption process is at an equilibrium state by studying the adsorption isotherms [99], which is essential and required to comprehend the interactions between the adsorbate and adsorbent and to establish an appropriate correlation of adsorption equilibria in order to optimize the adsorption

Table 1
Kinetic parameters of Al^{3+} adsorption over activated carbon

Pseudo-first-order equation parameters		Pseudo-second-order equation parameters	
K_1 (min^{-1})	0.039	K_2 ($\text{mg/g}\cdot\text{min}$)	0.001
q_e (mg/g)	2.254	q_e (mg/g)	0.192
R^2	0.960	R^2	0.986
RMSE	0.460	RMSE	0.090

Table 2
 Al^{3+} adsorption isotherm parameters with their corresponding correlation coefficients

	Langmuir isotherm parameters	Freundlich isotherm parameters	
	Al^{3+}		Al^{3+}
q_m (mg/g)	0.141	N	1.832
K_1	0.673	K_f	2.639
R^2	0.989	R^2	0.976
RMSE	0.053	RMSE	0.250

system. The fitting of the suitable model to the isotherm data analysis can be employed in the design process. Studies on the relevant adsorption isotherms have been done in order to examine the link between equilibrium adsorption capacity and equilibrium concentration for the AC. Table 2 provides a summary of the derived isotherm constants and correlation coefficients (R^2) for Freundlich and Langmuir isotherms as well as their corresponding RMSE values. The Langmuir isotherm model is predicated on the idea that a surface monolayer will only be able to adsorb homogenous substances at a few number of adsorption sites.

3.5. Industrial wastewater treatment

Due to the widespread industrialization and rapid population growth around the world, the management of industrial and municipal wastewater has become an inevitable challenge in today's world [29]. Excessive urbanization has caused serious environmental pollution, which is related to the high consumption of products containing trace metals and heavy metals in different concentrations produced by factories [88,90]. A variety of harmful metals enters water bodies and enters the food chain, causing serious health risks to humans. Different toxic metals have been reported to cause different physiological toxicity, but in general, trace metals and heavy metals are known to be associated with damage to nerves, kidneys, liver, bones, and block enzyme functional groups [29]. Since toxic metal ions persist in wastewater, removal of toxic metal ions is critical. Toxic metals can be removed and recovered from our environment and wastewater using a variety of physical and chemical techniques. The removal of Al^{3+} from wastewater by AC adsorbent was studied by batch adsorption test. The experimental results show that under the optimal conditions (concentration 30 mg/L, adsorbent dosage 0.20 g, contact time 90 min, temperature 60°C, pH 5), the maximum adsorption rate of AC to Al^{3+} is 98.44%.

3.6. Comparison of adsorption potential of different adsorbents

The comparative adsorption efficiencies of the adsorbents, specifically AC and different adsorption media, are compiled in Table 3 for Al^{3+} adsorption. The statistics demonstrate that AC has higher adsorption efficiency than any other adsorbents. The greater surface area of AC, which is what led to the higher adsorption capacity employed in our investigation, seems to be responsible for the efficiency of AC.

3.7. Desorption experiment

Desorption/regeneration of adsorbents is one of the key components of metal removal processes since it regulates the cost of water treatment technology [88,94]. In various experiments, adsorbents were recovered and effectively regenerated using salts (such as NH_4NO_3 , $\text{CaCl}_2\cdot 2\text{H}_2\text{O}$, KNO_3 , $(\text{NH}_4)_2\text{SO}_4$, $\text{C}_6\text{H}_5\text{Na}_3\text{O}_7\cdot 2\text{H}_2\text{O}$, KNO_3 , NaCl and KCl), buffer solutions (such as bicarbonate and phosphate), acids (such as H_2SO_4 , HCOOH , HCl , CH_3COOH , and HNO_3), chelating agents as well as deionized water [88,89,94]. In this study, deionized water and 0.1 M HCl were used for

Table 3
Comparison of different adsorbents used for Al³⁺ adsorption

Adsorbent	Adsorption potential	References
	(mg/g)	
	Al ³⁺	
Activated carbon	137.70	This article
PAN beads as-prepared	0.25	[100]
BDH activated carbon	0.02	[101]
Date pit	0.31	[101]
AXAD-16 resin	24.30	[102]
201 × 8 anion resin	5.60	[103]
XAD-4	4.40	[104]
IIP-PEI/SiO ₂	53.50	[105]

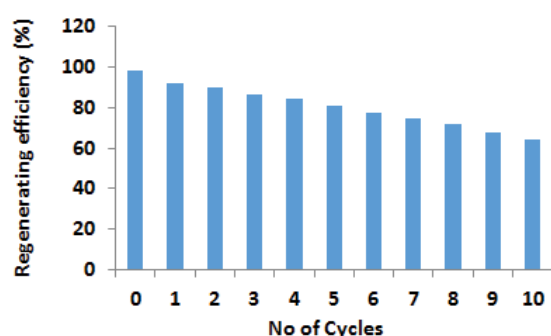


Fig. 9. Regenerating efficiency of activated carbon.

desorption experiments. The spent AC was added to 0.1 M HCl and stirred for 30 min, then the AC was recovered by filtration and subsequently dried in an oven. The dry, clean AC is then used to absorb a new batch of wastewater (Fig. 9).

4. Conclusion

Herein, we have provided a simple method for preparing AC from wPET. The developed AC can be served as an efficient adsorbent to remove Al³⁺ from aqueous solution as well as wastewater (>95%). The effective synthesis and microstructure of the AC were confirmed by a number of characterization techniques. AC possess a porous structure and large specific surface area, which provides sufficient active adsorption sites for Al³⁺. Moreover, the external factors affecting the adsorption performance were fully investigated. The outcomes demonstrated the AC superior reusability and high removal efficiency. The adsorption process is well modelled by the pseudo-second-order ($R^2 = 0.986$) and Langmuir models ($R^2 = 0.989$). The findings of this work suggest that AC is a potential adsorbent with the potential to effectively remove Al³⁺ from aqueous solutions. For future work, we plan to examine the adsorption efficiency of AC for organic and inorganic pollutants.

Research funding

The authors acknowledge the funding offered by University of Peshawar, Khyber Pakhtunkhwa, Pakistan.

Conflict of interest

Authors state no conflict of interest.

Informed consent

Informed consent is not applicable.

Ethical approval

The conducted research is not related to either human or animal use.

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