



Optimization of COD, apparent color, and turbidity reductions of landfill leachate by Fenton reagent

Amir Talebi, Norli Ismail*, Tjoon Tow Teng, Abbas F.M. Alkarkhi

*School of Industrial Technology, Universiti Sains Malaysia, Penang 11800, Malaysia
Tel. +604 6532824; Fax: +604 6573678; email: norlii@usm.my*

Received 18 June 2012; Accepted 13 March 2013

ABSTRACT

The efficiency of Fenton reagent ($\text{Fe}^{2+} + \text{H}_2\text{O}_2 + \text{H}^+$) for the leachate generated from Pulau Burung sanitary landfill site, located in Malaysia, was evaluated with the objectives of chemical oxygen demand (COD), apparent color, and turbidity reductions. Response surface methodology was used to study the effect of hydrogen peroxide (as oxidant) and ferrous sulfate (as catalyst) on COD, apparent colour, and turbidity reductions of sanitary landfill leachate. A central composite design in two factors [hydrogen peroxide dosage (mg/L) and ferrous sulfate dosage (g/L)] was employed to build a model as well as to perform optimization of the Fenton process. The optimum results for COD, apparent color, and turbidity reductions were 68, 81, and 86%, respectively, setting $[\text{H}_2\text{O}_2]$ at 747.8 mg/L and $[\text{FeSO}_4]$ at 12.8 g/L.

Keywords: Central composite design; Fenton; Landfill leachate; Response surface methodology

1. Introduction

Landfill leachate in municipal solid waste landfills is a contaminated liquid that contains a number of dissolved or suspended materials [1]. Over 200 individual organic compounds have been identified in leachate including cyclic and bicyclic compounds, aromatic hydrocarbons, and many others within a wide concentration range from less than 1 $\mu\text{g}/\text{L}$ to several hundred $\mu\text{g}/\text{L}$ [2]. Although landfill leachate has been proved to be toxic and recalcitrant, landfilling still remains one of the main methods for municipal and industrial solid waste disposal [3].

Many factors affect the municipal landfill leachate characterization such as landfill age, local climate and weather, disposal methods etc. Based on leachate characterization, several studies have been investigated on

leachate treatment including coagulation–flocculation, chemical precipitation, ammonium stripping, membrane filtration, biological treatment methods, etc. Although biological treatments are the most common and cost-effective methods of waste water treatment, it should be considered that high amount of chemical oxygen demand (COD), biochemical oxygen demand (BOD_5), and low biodegradability (in terms of BOD_5/COD ratio) in leachate characterization makes it as a refractory pollutant and classifies it as a resistant wastewater for any treatment (include biological treatment methods). In this case, to achieve better yield in order to enhance the leachate biodegradability, many researchers have worked on various treatment methods [4–17].

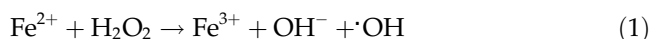
One of the most used pretreatment methods is advanced oxidation processes (AOP). Traditionally, AOP involve the generation of the hydroxyl radicals

*Corresponding author.

(·OH) which has a very high oxidation potential and is able to oxidize almost all organic pollutants [18]. Hydroxyl radicals are highly reactive species which are extremely strong oxidants resulting in several reactions including synergistic action of two oxidants (O_3 , H_2O_2), a catalyst and an oxidant (Fe^{2+} , H_2O_2), a photo catalyst and an oxidant (TiO_2 , H_2O_2) or irradiation plus oxidation (UV, O_3 , H_2O_2) [18].

One common feature of such system is the high demand of electrical energy for devices such as ozonizers, UV lamps, etc. and as a consequence, rather high treatment costs [3]. Among the mentioned AOP systems, Fenton process is cheap and effective enough to have preference over similar treatment methods. Fenton reagent is a method in which a mixture of [$FeSO_4/H_2O_2$] under acidic condition produces (·OH) radicals and oxidizes and decomposes organic compounds. Considering other advantages of the Fenton reagent (Fe^{2+} , H_2O_2 , H^+), it could be mentioned that there is no mass transfer limitation due to its homogeneous catalytic nature, there is no form of energy involved as catalyst, and the process is technologically simple [3].

Fenton's reagent is defined as the catalytic generation of hydroxyl radicals (·OH) resulting from the chain reactions between ferrous ion and hydrogen peroxide. Oxidation of organic compounds (RH) by Fenton's reagent can proceed by the following chain reactions [19]:



Many researchers have worked on Fenton process within a wide range of [Fe^{2+}]/[H_2O_2] ratios. In a study on a landfill leachate with 1,500 mg/L of initial COD, 200 mg/L H_2O_2 and 300 mg/L Fe^{2+} (Fe^{2+}/H_2O_2 : 1.5) have been used and 70% COD reduction is reported [20]. On the other hand, in another research on raw leachate with 10,540 mg/L initial COD concentration, 60% COD reduction is reported by using 10,000 mg/L H_2O_2 and 830 mg/L Fe^{2+} (Fe^{2+}/H_2O_2 : 0.083) [3]. Electro-Fenton technique and response surface methodology (RSM) were applied in a research in order to determine the optimal experimental conditions for E-Fenton process and the maximum COD and color removals achieved were 94.04 and 95.83% [21].

The present work was carried out to evaluate the effect of Fenton process on landfill leachate generated from Pulau Burong landfill, located in Penang, Malaysia. The objective of this study was to optimize the oxidant [H_2O_2] and catalyst [$FeSO_4$] in order to achieve highest COD reduction, apparent color removal, and turbidity reduction by implementing RSM through central composite design (CCD).

2. Materials and methods

Samples were taken from Pulau Burong landfill site, located in Penang. Pulau Burong landfill site is classified as a semi-aerobic landfill site in which the generated and collected leachate is aerated by the air pumps installed at the bottom of the collecting ponds and hence, generates mature and stabilized leachate type. Leachate sample was placed in high density polyethylene containers (volume of 40 L), transported to the laboratory, and was immediately stored in the refrigerator at 4°C as stated in the American Public Health Association (APHA, 1992) preservation techniques to minimize any changes in properties and characteristics of the samples [22].

Analytical grade chemical reagents were used in the study. Ferrous sulfate ($FeSO_4 \cdot 7 H_2O$) and sodium hydroxide (NaOH) supplied by System, sulfuric acid (H_2SO_4 , 97%), and hydrogen peroxide (H_2O_2 , 30%), disodiumhydrogen phosphate heptahydrate ($Na_2HPO_4 \cdot 7H_2O$, >98), di-kaliumhydrogen phosphate, (K_2HPO_4 , reagent grade) supplied by Merck, magnesium sulfate heptahydrate ($Na_2HPO_4 \cdot 7H_2O$, 99%) supplied by Bio-Lab, calcium chloride anhydrous ($CaCl_2$, ~97%) supplied by Sigma, Iron (III) chloride anhydrous ($FeCl_3 \cdot 6H_2O$ ~97%) supplied by Fluka, potassium dichromate ($K_2Cr_2O_7$, MW:294.19 g/mol) by R&M and silver sulfate (Ag_2SO_4 MW: 311.79) supplied by QReC.

The apparent color and COD were measured by using HACH DR/120 portable data-logging spectrophotometer (reported in units Pt/Co and mg/L, respectively). The apparent color unit calibrated based on APHA recommended standard of one colour unit being equal to 1 mg/L platinum as chloroplatinate ion. COD was measured using dichromate reactor design colorimetric, United State Environmental Protection Agency (USEPA) approved. BOD₅ was measured according to the standard method for the examination of water and wastewater using YSI 5100 dissolved oxygen meter. pH was measured by using HACH Senion pH meter. Turbidity was measured by using HACH 2100 turbidity meter (reported in N.T.U).

Leachate characterization was carried out instantly within two days for pH, COD, apparent colour, tur-

bidity, and BOD₅. All analysis procedures were performed according to standard methods and HACH method (modification of APHA standard method as approved by USEPA. Table 1 shows the characterization of Pulau Burung landfill leachate and based on leachate characterization, it can be considered a mature leachate with a BOD₅/COD ratio of 0.063. Hydrogen peroxide (H₂O₂, 30%) supplied by Merck OHG was used as an oxidant and FeSO₄ supplied by SYSTERM chemicals was used as a catalyst. Primarily, based on the literature review and according to the studies on the same samples [13 and 21], initial pH was fixed at 3.0, then predetermined dosage of FeSO₄ was added to samples and were stirred for three minutes in rapid mixing phase (200 rpm), then related H₂O₂ dosage were added to the samples and stirred for another three minutes in the same rapid mixing stage. Thereafter, the samples were stirred for one hour in slow mixing (50 rpm) and finally, one hour settling time was set to allow the generated sludge and particles to settle down.

Based on jar-test procedure, first, the various dosages of FeSO₄ (3, 6, 9, 12, 15, and 18 g/L) were examined at a fixed arbitrary H₂O₂ dosage (500 mg/L). After selecting the best FeSO₄ dosage – the dosage that gives the highest COD reduction– various dosages of H₂O₂ (200, 400, 600, 800, 1,000, 1,200, 1,400, 1,600, 1,800, 2,000, 2,200, and 2,400 mg/L) were examined. Based on results, the highest COD reduction percentage (60%) in related hydrogen peroxide (800 mg/L) and ferrous sulfate dosages (15 g/L) were selected for optimization.

3. Experimental design

RSM is a collection of mathematical and statistical techniques for empirical model building. The objective of RSM is to optimize a response (output variables) which is influenced by several independent variables (input variables). This method is used for a careful design of experiments, which is a series of tests called runs in which changes are made in the input variables

Table 1
Landfill leachate characterization

Parameters	Average
pH	8.3
COD (mg/L)	3,511
BOD ₅ (mg/L)	206
BOD ₅ /COD	0.06
Apparent colour (Pt.Co.)	6,019
Turbidity (N.T.U)	113

in order to identify the reasons for changes in the output response. The most extensive applications of RSM are in industrial fields where several input variables potentially influence some performance measure or quality characteristic of the product or process. This performance measure or quality characteristic is called the response [23]. RSM also specifies the relationships between one or more measured responses and the essential controllable input factors [24]. When all independent variables are measurable, controllable, and continuous in the experiments, with negligible error, the response surface can be expressed by

$$y = f(x_1, x_2, \dots, x_k) \quad (5)$$

where k is the number of independent variables. To optimize the response “ y ”, it is necessary to find an appropriate approximation for the true functional relationship between the independent variables and the response surface. Usually, a second-order polynomial Eq. (6) is used in RSM.

$$y = \beta_0 + \sum_{i=1}^2 \beta_i x_i + \sum_{i=1}^2 \beta_{ii} x_i^2 + \sum_{i < j} \beta_{ij} x_i x_j \quad (6)$$

where β_0 , β_i , β_{ii} , and β_{ij} are regression coefficients, and x_i and x_j are the coded variables.

A CCD with two independent variables was used for optimization and 13 runs were required to cover all possible combinations of factors levels. Data were collected from 2² factorial augmented by five (5) center points and four extra points (axial points or star points) as shown in Table 2. Experiments were run in random order to minimize the effects of unexpected variability in the observed responses.

4. Result and discussion

The experimental range for each independent variable was based on the preliminary trials, H₂O₂ dosage levels were selected in the range of (200–800) mg/L and FeSO₄ dosage levels were (6–15) g/L.

The results of 13-run CCD in the two variables, H₂O₂ and FeSO₄ dosage and three responses, COD, apparent color, and turbidity reductions are given in Table 2. The percentage of COD, apparent color, and turbidity reductions were studied in their specified levels, H₂O₂ dosage (200–800) mg/L and FeSO₄ dosage (6–15) g/L.

RSM was used to evaluate the effect of H₂O₂ and FeSO₄ dosage on the COD, apparent color, and turbidity reductions, then a model was built in order to describe the behavior of each response in order to

Table 2
CCD in natural and coded variables with the experimental data

Natural variable		Coded variable		Responses (%)			Predicted value (%)		
ξ_1	ξ_2	H ₂ O ₂	FeSO ₄	COD	Apparent colour	Turbidity	COD	Apparent colour	Turbidity
200	6	-1	-1	45.00	53.80	-41.70	36.36	47.84	-78.98
800	6	1	-1	44.80	59.70	-15.40	46.04	63.01	-12.59
200	15	-1	1	56.60	70.20	47.10	55.16	66.00	19.94
800	15	1	1	62.70	75.90	81.60	64.84	81.16	86.33
75.74	10.5	-1.41	0	35.40	47.60	-79.40	41.83	62.03	-39.72
924.26	10.5	1.41	0	58.60	82.30	65.40	55.52	83.48	54.17
500	4.14	0	-1.41	34.70	41.70	-88.30	39.24	43.41	-69.82
500	16.86	0	1.41	67.00	70.00	60.10	65.81	69.09	70.07
500	10.5	0	0	68.00	77.40	60.50	66.38	72.76	56.48
500	10.5	0	0	65.80	75.00	52.60	66.38	72.76	56.48
500	10.5	0	0	66.10	74.40	53.00	66.38	72.76	56.48
500	10.5	0	0	66.10	76.60	59.60	66.38	72.76	56.48
500	10.5	0	0	65.90	75.20	56.70	66.38	72.76	56.48

optimize the process by finding the best setting of H₂O₂ and FeSO₄ dosage that maximize COD, apparent color, and turbidity reductions. Second-order model for COD reduction, apparent color removal and turbidity reduction in terms of coded variable are given by Eqs. (7)–(9), respectively:

$$y_1 = +66.38 + 4.84x_1 + 9.40x_2 - 8.85x_1^2 - 6.93x_2^2 \quad (7)$$

$$y_2 = +72.76 + 7.58x_1 + 9.08x_2 - 8.25x_2^2 \quad (8)$$

$$y_3 = +56.48 + 33.20x_1 + 49.46x_2 - 24.63x_1^2 - 28.18x_2^2 \quad (9)$$

where y_1 , y_2 , and y_3 represent the percentage of COD, apparent color, and turbidity reduction, respectively, and x_1 and x_2 represent [H₂O₂] and [FeSO₄], respectively.

The second-order regression models obtained for COD, apparent color, and turbidity reductions are satisfied since the values of the coefficient of determination (R^2) are high and close to 1. The values of R^2 for COD, apparent color and turbidity reductions models are 0.913, 0.818, and 0.895, respectively. When the value of R^2 for COD is 0.913, it means that 0.913 of total variation is explained by the model and 0.087 of total variation is unexplained. Similarly, when the value of R^2 for apparent color is 0.818 of total variation is explained by the model, 0.117 of total variation is unexplained. This indicates that 0.82–0.91 of the total variation is explained by the model and only (0.09–0.18) of the total variation is unexplained.

The 3-D response surface for COD, apparent color, and turbidity reductions are given in Figs. 1–3 showing the effect of FeSO₄ and H₂O₂ dosage on COD, apparent color, and turbidity reductions, respectively. A positive sign for the regression coefficient in the fitted model indicates the ability of the factor to increase the response while the negative sign indicates the ability of a factor to decrease the response. It can be seen that all responses exhibited a clear peak suggesting that the optimum condition for maximum COD, apparent color, and turbidity reductions are well inside the design boundary.

Figs. 1–3 show a decrease in responses reduction efficiency. This could be due to the fact that during

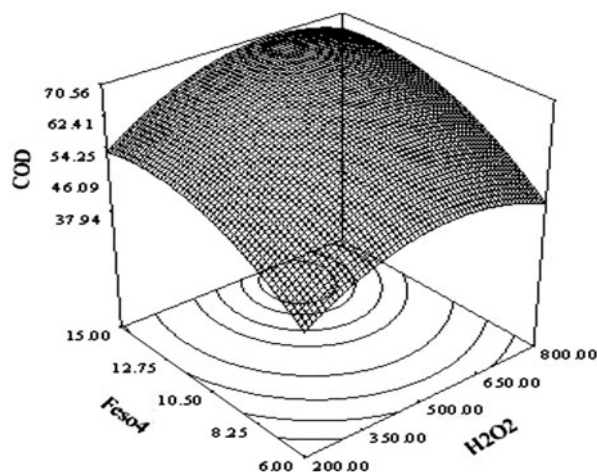


Fig. 1. 3D response surface for COD reduction as a function of H₂O₂ and FeSO₄.

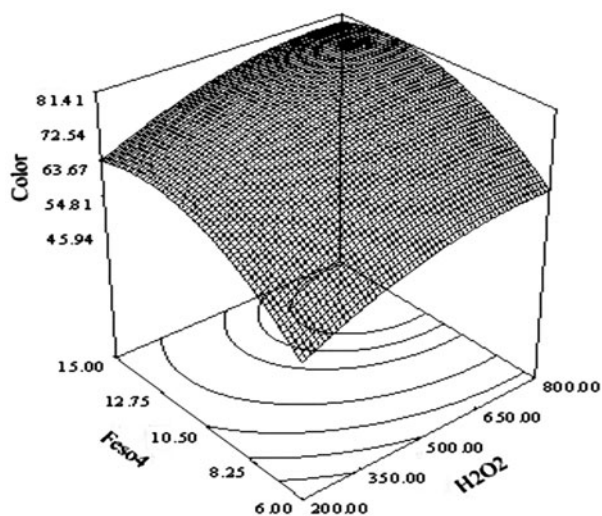


Fig. 2. 3D response surface for apparent color removal as a function of H_2O_2 and FeSO_4 .

oxidation process, these organic compounds are less stabilized and more ready for degradation through oxidation. These compounds are oxidized first and the remaining compounds are refractory complexes and resistant to oxidation. The results of the analysis of variance (ANOVA) for the quadratic models summarize the analysis of each response and show significant model terms. Table 3 shows the ANOVA for COD, apparent color, and turbidity reductions.

The analysis of variance revealed that a second-order model adequately fitted the experimental data for all responses. For COD, apparent color and turbid-

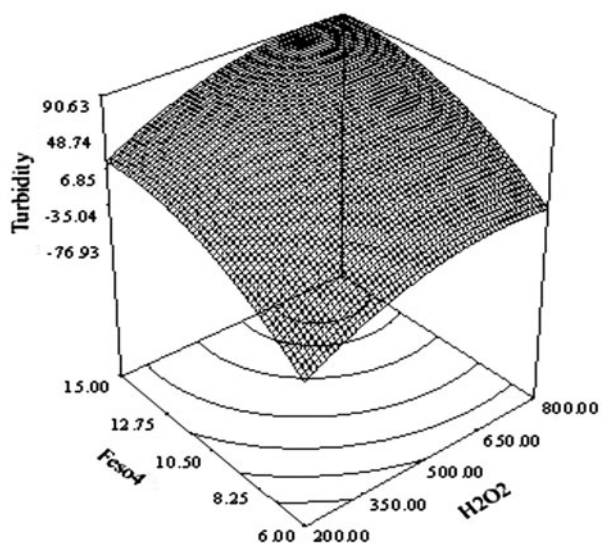


Fig. 3. 3D response surface for turbidity reduction as a function of H_2O_2 and FeSO_4 .

ity reductions, it could be seen that linear and quadratic effect of $[\text{H}_2\text{O}_2]$ (x_1) and $[\text{FeSO}_4]$ (x_2) are significant, except the quadratic effect of $[\text{H}_2\text{O}_2]$ for apparent color (all related P -values shown in Table 3 are less than 0.05). The interaction between $[\text{H}_2\text{O}_2]$ and $[\text{FeSO}_4]$ (\times) shows insignificant effect for all responses (related P -values > 0.05).

Results reveal that higher concentration of $[\text{FeSO}_4]$ yields higher efficiency on responses reductions until maximum reductions called optimum conditions are achieved and then a decrease in efficiency is observed. Same behavior for $[\text{H}_2\text{O}_2]$ was detected meaning that as far as $[\text{H}_2\text{O}_2]$ concentration was increased, COD, apparent color and turbidity reductions increased as well and reached up to maximum reduction, optimum condition, and then beyond that, responses reductions started to slow down (decrease). However, when $[\text{H}_2\text{O}_2]/[\text{Fe}^{2+}] \geq 2.0$, sludge settling was hindered by the oxygen bubbling produced from the decomposition of excess hydrogen peroxide which leads to the side reactions (10) and (11) [19,25, 26].

One of the deterrent factors in Fenton process, which causes the observed descent in process efficiency is due to hydroxyl scavenging role of iron ion against the produced $\cdot\text{OH}$ (by H_2O_2) during reaction 10, which describes that by increasing the FeSO_4 dosage, the available additional iron ions try to oxidize themselves through the reaction below



Incidentally, another hydroxyl radicals scavenging is done by hydrogen peroxide itself as shown in reaction 11.



5. Optimization of the experiment

Based on the above discussion, it is better to run an optimization study to find out the optimal operating conditions for H_2O_2 concentration and FeSO_4 . In fact, once the model was developed and checked for adequacy, the optimization criteria was set to find out the optimum operating conditions. Maximum reduction for all responses was achieved at 12.8 g/L of FeSO_4 and 747.8 mg/L of H_2O_2 concentration. To validate the optimum combination of the process variables, confirmatory experiments were carried out. The selected combinations of the two variables resulted in 68% COD reduction, 81% apparent color removal, and 86% turbidity reduction.

Table 3
ANOVA for COD, apparent color and turbidity reductions

Source	Sum of squares	DF	Mean Square	F-Value	P-Value
<i>COD</i>					
Model	1,674.79	4	418.70	21.07	0.0003
H ₂ O ₂	187.31	1	187.31	9.43	0.0153
FeSO ₄	706.49	1	706.49	35.56	0.0003
H ₂ O ₂ × H ₂ O ₂	545.16	1	545.16	27.44	0.0008
FeSO ₄ × FeSO ₄	333.85	1	333.85	16.80	0.0034
Residual	158.95	8	19.87		
Total	1,833.75	12			
<i>Apparent color</i>					
Model	1601.58	3	533.86	13.52	0.0011
H ₂ O ₂	460.15	1	460.15	11.66	0.0077
FeSO ₄	659.25	1	659.25	16.70	0.0027
FeSO ₄ × FeSO ₄	482.18	1	482.18	12.21	0.0068
Residual	355.30	9	39.48		
Total	1,956.88	12			
<i>Turbidity</i>					
Model	37,015.75	4	9,253.94	17.01	0.0006
H ₂ O ₂	8,816.47	1	8,816.47	16.21	0.0038
FeSO ₄	19,569.27	1	19,569.27	35.97	0.0003
H ₂ O ₂ × H ₂ O ₂	4,219.23	1	4,219.23	7.76	0.0237
FeSO ₄ × FeSO ₄	5,523.28	1	5,523.28	10.15	0.0129
Residual	4,352.35	8	544.04		
Total	41,368.10	12			

6. Conclusion

The effects of [H₂O₂] and [FeSO₄] on the Fenton process using RSM, were evaluated in a batch reactor. It was proved that a second-order model adequately fitted the experimental data for all responses. For COD, apparent color, and turbidity reductions, it could be seen that linear and quadratic effect of [H₂O₂] and [FeSO₄] are significant, except the quadratic effect of [H₂O₂] for apparent color. It could be concluded that optimal condition for Fenton process is at 747.8 mg/L of [H₂O₂] and 12.8 g/L of [FeSO₄] for COD, apparent color, and turbidity reductions. The combination of two factors (H₂O₂ and FeSO₄ dosage) resulted in 68, 81, and 86% reductions for COD, apparent color, and turbidity, respectively. The responses reduction efficiency decreases beyond the optimal conditions.

Acknowledgment

The study was funded through USM short term grant (grant number 304/PTEKIND/638090. The authors acknowledge the USM for providing all facilities.

References

- [1] X. Qian, R.M. Koerner, D.H. Gray, Geotechnical Aspects of Landfill Design and Construction, Prentice-Hall, Upper River, NJ, 2002.
- [2] N. Paxeus, Organic Compounds in Municipal Landfill Leachate, J. Water Sci. Technol. 42 (2000) 323–333.
- [3] A. Lopez, M. Pagano, A. Volpe, A.C. Di Pinto, Fenton's pre-treatment of mature landfill leachate, Chemosphere 54 (2004) 1005–1010.
- [4] F. Ilhan, U. Kurt, O. Apaydin, M.T. Gonullu, Treatment of leachate by electrocoagulation using aluminum and iron electrodes, J. Hazard. Mater. 154 (2008) 381–389.
- [5] J.J. Schoeman, Evaluation of electro dialysis for the treatment of a hazardous leachate, Desalination 224 (2008) 178–182.
- [6] K.Y. Foo, B.H. Hameed, An overview of landfill leachate treatment via activated carbon adsorption process, J. Hazard. Mater. 171 (2009) 54–60.
- [7] M.K. Mehmood, E. Adetutu, D.B. Nedwell, A.S. Ball, In situ microbial treatment of landfill leachate using aerated lagoons, Bioresource. Technol. 100 (2009) 2741–2744.
- [8] H.S. Li, S. Zhou, Y.B. Sun, P. Feng, J.D. Li, Advanced treatment of landfill leachate by a new combination process in a full-scale plant, J. Hazard. Mater. 172 (2009) 408–415.
- [9] Y.N. Lima, M.G. Shaabana, C.Y. Yinb, Treatment of landfill leachate using palm shell-activated carbon column: Axial dispersion modeling and treatment profile, Chem. Eng. J. 146 (2009) 86–89.
- [10] J.S. Guo, A.A. Abbas, Y.P. Chen, Z.H. Liu, F. Fang, P. Chen, Treatment of landfill leachate using a combined stripping, Fenton, SBR, and coagulation process, J. Hazard. Mater. 178 (2010) 699–705.
- [11] H. Li, Y. Gu, Y. Zhao, Z. Wen, Leachate treatment using a demonstration aged refuse biofilter, J. Environ. Sci. 22 (2010) 1116–1122.
- [12] S.Q. Aziz, H.A. Aziz, M.S. Yusoff, M.J. Bashir, Landfill leachate treatment using powdered activated carbon augmented sequencing batch reactor (SBR) process: Optimization by response surface methodology, J. Hazard. Mater. 189 (2011) 404–413.

- [13] M.J. Bashir, H.A. Aziz, M.S. Yusoff, New sequential treatment for mature landfill leachate by cationic/anionic and anionic/cationic processes: Optimization and comparative study, *J. Hazard. Mater.* 186 (2011) 92–102.
- [14] E.M.R. Rocha, V.J.P. Vilar, A. Fonseca, I. Saraiva, R.A.R. Boaventura, Landfill leachate treatment by solar-driven AOPs, *Sol. Energy* 85 (2011) 46–56.
- [15] W.G. Shim, J.M. Abdul, T. Mohammad, S. Vigneswaran, H.H. Ngo, Biofilter in leachate treatment processes, *Desalin. Water Treat.* 41 (2012) 249–257.
- [16] T. Ismail, M. Toshihiko, Bio-treatment of landfill leachate having low Carbon–Nitrogen ratio in a bio-film reactor packed with granular activated carbon under control of oxygen gas concentration, *Desalin. Water Treat.* 37 (2012) 55–61.
- [17] C.G. Litas, I.A. Zouboulis, D. Zaboulis, P. Samaras, The use of a submerged membrane batch reactor (S.M.B.R) for co-treatment of landfill leachates and domestic wastewater, *Desalin. Water Treat.* 39 (2012) 284–290.
- [18] J.J. Morais, P.P. Zamora, Use of AOP to improve the biodegradability of mature landfill leachate, *J. Hazard. Mater.* 123 (2005) 181–186.
- [19] H. Zhang, H.J. Choi, C.P. Huang, Optimization of Fenton process for the treatment of landfill leachate, *J. Hazard. Mater.* 125 (2005) 166–174.
- [20] I.W.C. Lau, P. Wang, H.H.P. Fang, Organic removal of anaerobically treated leachate by Fenton coagulation, *J. Environ. Eng.* 127 (2001) 666–669.
- [21] S. Mohajeri, H.A. Aziz, M.H. Isa, M.A. Zahed, M.N. Adlan, Statistical optimization of process parameters for landfill leachate treatment using electro-Fenton technique, *J. Hazard. Mater.* 176 (2010) 749–758.
- [22] APHA, Standard Methods for the Examination of Water and Wastewater, eighteenth ed., American Public Health Association, Washington, DC, 1992.
- [23] R.H. Meyers, D.C. Montgomery, Response Surface Methodology, second ed., Wiley-Interscience publication, New York, NY, 2002.
- [24] A.I. Khuri, J.A. Cornell, Response Surfaces Design and Analysis, second ed., Marcel Dekker, New York, NY, 1996.
- [25] H. Tekin, O. Bilkay, S.S. Ataberk, T.H. Balta, I.H. Ceribasi, F.D. Sanin, F.B. Dilek, U. Yetis, Use of Fenton oxidation to improve the biodegradability of pharmaceutical wastewater, *J. Hazard. Mater.* 136 (2006) 258–265.
- [26] D. Hermosilla, M. Cortijo, C.P. Huang, Optimizing the treatment of landfill leachate by conventional Fenton and photo-Fenton processes, *Sci. Total Environ.* 407 (2009) 3473–3481.