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Influence of microbial floatation on membrane fouling due to particles and organic matters in submerged microfiltration

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ABSTRACT

This study investigated the effect of microbubble floatation on the fouling of microfiltration membranes. Synthetic feed solutions containing either kaolin or alginic acid were used for the membrane filtration tests. A dissolve air flotation system using a pump-type microbubble generator was adopted as a pretreatment for microfiltration system. A multi-array sub-merged membrane filtration system was used to monitor the changes in transmembrane pressure with time under various operating conditions. Turbidity removal efficiency, particle counts, and fouling rate for the microfiltration membrane were measured after the microfiltration floatation. The effect of coagulant dose on the treated water quality and membrane performance was also investigated. Results showed that microbubble floatation without coagulant was effective to reduce membrane fouling by kaolin but it was not very effective to control fouling by alginate. With the aid of coagulant, microbubble floatation could control fouling by alginate. This is attributed to the removal mechanisms of foulants by microbubble floatation: suspended particles can be separated by the microbubbles without coagulant but dissolved organics can be only removed by the combined effect of coagulation and floatation.

Keywords: Microbubble; Flotation; Microfiltration; Fouling; Coagulation; Pretreatment

1. Introduction

Dissolved air flotation (DAF) is one of the advanced water treatment techniques, aiming at the

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removal of pollutants for the clarification of water and wastewater [1–3]. This can be done by pressurizing air into water and releasing it in a flotation tank or basin, which forms small air bubbles [4]. Pollutants such as suspended particles, oils, and greases can be adsorbed on the surface of the bubble and separated as a form

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of scum on the surface of water. This is widely accepted for the treatment of municipal and industrial wastewaters [2,5,6].

The efficiency of DAF can be improved by adjusting the properties of bubbles used for pollutant removal [7]. One of such approaches is microbubble floatation [8]. Microbubbles are defined as bubbles with diameters on the order of $10-50 \,\mu\text{m}$ and have unique properties due to their small size [8,9]. For instance, microbubbles have high-specific surface area (the ratio of surface area to the volume), allowing the high efficiency of pollutant adhesion [7,10]. Due to its high effectiveness of pollutant removal, microbubble floatation has drawn attention for not only wastewater treatment [11–13] but also water purification [14,15] and pretreatment for seawater desalination [16–18].

One of the possibilities of using microbubble floatation is the pretreatment of feedwater to membrane process. Microfiltration (MF) and ultrafiltration (UF) are increasingly used although fouling of these membranes is still a serious issue [19-21]. Currently, coagulation is applied for the pretreatment of feedwater to MF/UF process [22-28]. However, it is difficult to remove particles with low density and organic matters by this technique. Accordingly, the application of microbubble floatation prior to MF or UF may have potential to control fouling by low-density particles, organic matters, and algae. Nevertheless, there are relatively few fundamental research works on the application of microbubble floatation as a pretreatment for MF or UF membranes. In addition, optimum conditions for microbubble floatation process are not fully revealed yet. Accordingly, this study focused on the investigation of the effect of microbubble floatation on the fouling of microfiltration membranes. Using synthetic feedwaters containing model foulants, the fundamental properties of microbubble floatation were experimentally examined. Moreover, the effect of coagulant dose on the pretreatment efficiency by microbubble floatation was also explored.

2. Materials and methods

2.1. Feed solutions

Two different synthetic feed solutions were prepared for the experiments of microbubble floatation and microfiltration. The first synthetic solution contained kaolin of 150 mg/L and the second one contained sodium salt of alginic acid of 150 mg/L. Prior to the experiments, either kaolin or alginate was added to the deionized water of 40 L. All the solutions were mixed by an electromagnetic stirrer for 1 h. After the stirring, each solution was transferred into the flotation device or into the MF system. All experiments were carried out at room temperature (25° C). Table 1 shows the water quality parameters of the synthetic feedwaters.

2.2. Experimental setup

Fig. 1 shows the schematic diagram for the microbubble floatation device. This system consists of a feed tank, a microbubble generator, and a floatation tank. The feed and floatation tanks are made of acrylic resin. The microbubble generator used a three phase induction motor (KTM15N1D042 M-000, Toshiba, Japan). The water flow rate was 8 L/min and the gas flow rate was 0.64 L/min. Microbubbles were produced using a gas–water circulation type generator, and supplied to the bottom of the tank under the pressure of 4 bar. According the manufacturer, the average size of the microbubbles was 40μ m. The floatation was operated in a continuous flow mode for 5 min. Water samples were taken at 0, 2.5, and 5 min of floatation. Therefore, three samples were obtained for each test.

This microbubble floatation system was operated without and with adding coagulant. Poly aluminum chloride (PACl) was used as the coagulant. The dose of PACl was determined to be 2 mg/L based on the Jar test results. To examine the effect of coagulation on the efficiency of microbubble floatation, PACl was directly added to the feed tank of the microbubble floatation system.

After the microbubble floatation, microfiltration experiments were carried out using a multi-array membrane filtration system. The detailed information on this system is available elsewhere [29]. The experiments were conducted using hollow fiber membranes made of polyvinylidene fluoride. Prior to the filtration test, the membrane fibers were immerged into ethanol solution for 1 h and then rinsed with deionized water for pore wetting. Then, the membranes were vertically submerged into a cylindrical feed tank. In each feed tank, a magnetic stirrer was placed for continuous mixing of solution. The pure water flux was measured using deionized water and then the permeate flux was measured using the feedwater pretreated bv microbubble floatation. A multi-channel cartridge

Table 1 Water quality parameters of the synthetic feedwaters

	Turbidity (NTU)	pН
Kaolin solution	110–125	5.5
Alginate solution	50-70	5



Fig. 1. Schematic diagram for the microbubble floatation device.

peristaltic pump (model 7535-08, Cole Palmer, USA) was used, which allows up to 15 filtration tests at the same time. A pressure transducer monitored the transmembrane pressure through the membrane under constant flux mode. The filtration tests were done at three different flux conditions: 30, 60, and 90 L/m² h. All the data were collected by a data logger and recorded by a data acquisition software. The test conditions for the membrane filtration are summarized in Table 2.

2.3. Calculation of fouling rate

Using the results on transmembrane pressure, fouling rate was calculated using the following method. To begin, a simple filtration model to estimate the fouling rates in dead-end filtration system was applied. Assuming that fouling occurs through cake formation mechanisms, the transmembrane pressure is given by [30,31]:

Table 2 Conditions for microfiltration experiments using hollow fiber membranes

Parameters	Conditions
Membrane material	PVDF (polyvinylidene difluoride)
Pore size	$0.1 \ \mu m$
Fiber diameter	O.D $1.15/ID \ 0.7 \ mm$
Length	240 mm
Flow configuration	Outside-In
Type of treatment	Dead-end filtration
Flux	30, 60, 90 L/m ² h for 2 h at each flux
Temperature	25°C



Fig. 2. Effect of microbubble flotation on turbidity removal with and without using PACI: (a) kaolin solution and (b) alginate solution.





Fig. 3. Dependence of fouling rate on floatation time at different flux conditions for kaolin solution. (a) Microbubble floatation without PACl and (b) Microbubble floatation with PACl (2 mg/L).

$$\Delta P = \mu (R_{\rm m} + R_{\rm c}) J \tag{1}$$

where *J* is the permeate flux, ΔP is the transmembrane pressure, μ is the absolute viscosity of water, $R_{\rm m}$ is the intrinsic membrane resistance, and $R_{\rm c}$ is the cake resistance. The cake resistance is a function of the specific cake resistance (α), the mass of the cake deposited on the membrane ($m_{\rm c}$), and the membrane area ($A_{\rm m}$) [32]:

$$R_{\rm c} = \frac{\alpha m_{\rm c}}{A_{\rm m}} \tag{2}$$

Since the dead-end filtration was assumed, the cake mass m_c is given by:

Fig. 4. Dependence of fouling rate on floatation time at different flux conditions for alginate solution: (a) Microbubble floatation without PACl and (b) Microbubble floatation with PACl (2 mg/L).

$$m_{\rm c} = J A_{\rm m} c t \tag{3}$$

where *c* is the effective concentration of the foulant. By combining the Eqs. (1)–(3), ΔP is given by [33]:

$$\Delta P = \mu R_{\rm m} J + \mu \alpha c J^2 t = \mu J R_{\rm m} + \theta t \tag{4}$$

where θ is defined as the fouling rate. In a constant flux operation, θ can be calculated from the slope of *t* and ΔP :

$$\theta = \frac{\mathrm{d}\Delta P}{\mathrm{d}t} = \alpha \,\mu J^2 \,c \tag{5}$$



Fig. 5. Effect of microbubble floatation on particle counts for kaolin solution: (a) without PACl (b) with PACl.

Accordingly, θ is proportional to the product of the specific cake resistance and effective concentration of foulants, which represents the potential of fouling for the feedwater. Moreover, θ is also proportional to the J^2 , implying that fouling is accelerated at high flux rate.

2.4. Analytical methods

Turbidity and pH of the water sample were measured using a turbidity meter (TURIB 430 IR, Germany) and a pH meter (Orion Star, Thermo Scientific), respectively. The distribution of particles in the water samples was analyzed using a particle counter (CHEMTRAC laser trac particle counter PC 3400, USA). After the microfiltration tests, the surface of the membrane and foulant layer were examined by Field emission scanning electron microscopy (Model: S-4700, Hitachi, Japan).



Fig. 6. Effect of microbubble floatation on particle counts for alginate solution: (a) without PACl (b) with PACl.

3. Results and discussion

3.1. Effect of microbubble floatation on turbidity removal

As listed in Table 1, the turbidity values for the kaolin and alginate solutions were in the range between 110 and 125 NTU and between 50 and 70 NTU, respectively. After microbubble floatation, the turbidity of the feed solutions was changed. Fig. 2(a) shows the turbidity removal efficiency for the kaolin solution by microbubble floatation. At the floatation time of 2.5 min, the turbidity removal was less than 20%. However, the turbidity removal increased with increasing the floatation time up to 5 min. With the addition of PACl, the turbidity removal was higher (~72%) than that without using PACl (~57%).

However, the turbidity removal for the alginate solution was different from that for the kaolin solution. Without using PACl, the turbidity removal was less than 5%, implying that the removal efficiency is very low. The turbidity removal increased up to 29%



Fig. 7. SEM images of membrane surfaces after microfiltration experiments using the kaolin solution: (a) clean membrane, (b) control, (c) microbubble floatation (2.5 min), (d) microbubble floatation (5 min), (e) control with PACI, (f) microbubble floatation with PACI (2.5 min), and (g) microbubble floatation with PACI (5 min).

by adding PACl but it is still a low value. Since particulate forms of alginate result in turbidity in this feed solution, it is difficult to reduce turbidity by removing alginate by coagulation and air floatation. Accordingly, the turbidity removal in the alginate solution should be lower than that in the kaolin solution. These results suggest that microbubble floatation is effective to remove particulate matters but is not efficient to remove organic particles without added coagulant.

3.2. Effect of microbubble floatation on MF fouling rate

Fig. 3(a) compares the fouling rates (θ) at different floation time and flux values for the kaolin solution. In this case, microbubble floatation was applied without PACl addition. Before applying microbubble floatation, the fouling rate was high. As the flux increased, the fouling rate increased, which is evident from Eq. (5). After applying microbubble floatation, the fouling rate became negligible, implying that the

fouling by kaolin was efficiently reduced even without using coagulant.

The fouling rates of the kaolin solution treated by microbubble floatation with PACl were shown in Fig. 3(b). At the flotation time of 0 min, PACl was added to the feed solution and no microbubble floatation was applied. Nevertheless, the fouling rates were reduced by the effect of coagulation by the added PACl. After applying the microbubble, the fouling rates were further reduced. Again, the fouling by kaolin could be alleviated by microbubble floatation and PACl.

Fig. 4 shows the effect of microbubble floatation on the fouling rate of the alginate solution. Without PACl addition (Fig. 4(a)), the microbubble floatation was less efficient to decrease fouling rate for the alginate solution than for the kaolin solution. This could be attributed to low removal of turbidity by the microbubble floatation without PACl as shown in Fig. 2(b). Nevertheless, the microbubble floatation



Fig. 8. SEM images of membrane surfaces after microfiltration experiments using the alginate solution: (a) clean membrane, (b) control, (c) microbubble floatation (2.5 min), (d) microbubble floatation (5 min), (e) control with PACl, (f) microbubble floatation with PACl (2.5 min), and (g) microbubble floatation with PACl (5 min).

could reduce the fouling rate by more than 50%. When PACl was used in the microbubble floatation, the fouling rate was significantly reduced as shown in Fig. 4(b). This suggests that the fouling by alginate could be efficiently reduced by the combination of microbubble floatation and PACl addition.

3.3. Changes in particle counts by microbubble floatation

The results on the particle size analysis are shown in Figs. 5 and 6. The information on the particle size distribution in the feed solution to MF/UF is important for better understanding of the fouling potential. Although similar amounts of particles exist, the smaller particles cause more severe membrane fouling than the larger ones. As shown in Fig. 5(a), the total number of particles was decreased by applying microbubble floatation. Especially, the particles in the range between 10 and 20 μ m were removed after microbubble floatation of 2.5 min. The particles in the range between 0 and 10 μ m were also removed after microbubble floatation of 5 min. With the addition of PACl, the particles were further removed by microbubble floatation. The total number of particles in the range between 0 and 50 μ m was reduced to 2,500 per mL after microbubble floatation of 5 min with PACl.

Fig. 6 shows the particle size distribution for the alginate solution with microbubble floatation. Without PACl, the number of particles was reduced by the microbubble floatation but the efficiency was relatively low. With the addition of PAC, the number of particles significantly decreased by the microbubble floatation. Again, it was confirmed that microbubble floatation together with PACl addition is required to reduce fouling potential for the alginate solution.

3.4. SEM analysis of foulant layers on the membrane

After the microfiltration tests, the membrane was stored and prepared for the SEM analysis to visualize the changes in foulant layer by microbubble floatation. Fig. 7 shows the SEM images of the membrane surface after microfiltration of the kaolin solution. It appears that the foulant layer on the membrane surface is affected by the feedwater pretreatment using microbubble floatation. For example, a thick layer of the cake was observed in Fig. 7(e) (control with PACl), while a thin and porous cake layer was found in Fig. 7(f) and (g). This suggests that the microbubble floatation changed the morphology of the cake layer. Similar results were found with the alginate solution as shown in Fig. 8(b)–(d).

4. Conclusions

In this study, the effect of microbubble floatation on the fouling of microfiltration membranes was investigated. The following conclusions were withdrawn:

- (1) Microbubble floatation without PACl was found to be effective to remove turbidity from the kaolin feed solution. However, it was not efficient to remove turbidity from the alginate feed solution. In this case, the removal efficiency of the turbidity was improved by the addition of PACl (2 mg/L) to the microbubble floatation tank.
- (2) The fouling rate of the kaolin feed solution was reduced by microbubble floatation without PACl addition. A further decrease in the fouling rate was also observed with PACl addition.
- (3) The fouling rate of the alginate feed solution was not effectively reduced by microbubble floatation alone. In this case, PACl addition was required to reduce the fouling rate.
- (4) Microbubble floatation affected the particle counts of the feed solutions. It took more time to remove smaller particles without PACl addition.
- (5) The analysis of SEM images suggests that the morphology of the fouling layer changes by microbubble floatation. More porous cake layer could be formed after the pretreatment of feedwater using the microbubble floatation.

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List of abbreviations

$A_{\rm m}$	—	membrane area
С	—	effective foulant concentration
DAF	_	dissolved air flotation
J	—	flux
LMH		liters per square meter per hour $(L/m^2/h)$
$m_{\rm c}$		mass of the cake
MF	—	microfiltration
R _m	_	clean membrane resistance
$R_{\rm c}$	—	cake resistance
TMP	—	transmembrane pressure
μ	—	absolute viscosity
ΔP	_	transmembrane pressure
θ		fouling rate

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