



## Laboratory-scale study of the advanced Fenton process for silica removal from brackish underground water in arid areas in Saudi Arabia

Fathi Djouider\*, Mohammed S. Aljohani

Nuclear Engineering Department, Radiation Chemistry Division, King Abdulaziz University, P.O. Box 80204, Jeddah 21589, Saudi Arabia, Tel. +966 558822318; email: fathid@yahoo.com (F. Djouider), Tel. +966 545345000; email: mjohani@kau.edu.sa (M.S. Aljohani)

Received 12 April 2016; Accepted 27 November 2016

### ABSTRACT

Silica scaling of reverse osmosis (RO) membranes in water desalination poses a worldwide problematic issue in the cost-effective operations in seawater desalination plants. Underground water is one of the main water resources in the Kingdom of Saudi Arabia. Most of the inland cities and remote villages rely entirely on this supply mode for their domestic, industrial and agricultural purposes. However, this water resource is characterized by its relatively high silica contents. Therefore, pretreatment for silica removal from the RO feed is crucial in a system running on high silica waters. In this preliminary work, we investigated the effectiveness of advanced Fenton process (using hydrogen peroxide and zero valent metal iron  $Fe^0$  as source of  $Fe^{3+}$ ) for the silica pretreatment of water from the Buwaib deep aquifer by its co-precipitation with ferric hydroxide. The influences of several reaction parameters, i.e., iron powder dosage, hydrogen peroxide concentration, pH effect, were investigated. In the optimum conditions, up to 70% of silica was removed. This preliminary study showed that this environmentally friendly process is an effective silica mitigation strategy prior to the RO desalination.

*Keywords:* Silica removal; Advanced Fenton process; Hydrogen peroxide effect; Iron dosage effect; pH effect

### 1. Introduction

#### 1.1. Major water resources in Saudi Arabia

Saudi Arabia, with an area of around 2.1 million  $km^2$  and an estimated population of 26 million in 2011, is an extremely arid country. The annual average precipitation during the period 1979–2009 was 93.5 mm [1]. There are no permanent watercourses or lakes. For its water consumption, the country relies mainly on seawater desalination and deep/shallow underground brackish aquifers [2], which are invaluable resources of drinking water in remote arid areas (Fig. 1). It is usually organic compounds and biological species-free.

#### 1.2. Groundwater from deep and shallow aquifers

Fig. 2 shows the estimated site of these huge and very deep sedimentary underground aquifers, formed from

the Paleozoic era (600 million years ago) to the Cenozoic era (present age) [3]. With a reserve estimated between 260 and 760 billion  $m^3$  and with an average renewal rate of 900 million  $m^3$ , they are rapidly depleting [4]. The surface water comes from the limited seasonal rainfall. Part of the runoffs recharges the shallow aquifers (Tuwail, Skaka, Khuff, Jauf, Jileh and Jubayla) while the remaining evaporates due to the extreme hot weather. Most of these deep and shallow aquifers are categorized as brackish, where they contain over 1,000 ppm of total dissolved solids, mainly silica [5].

#### 1.3. Seawater desalination

The state-owned Saline Water Conversion Corporation (SWCC) runs 28 desalination plants located on the western and eastern coasts of the country, with an estimated production of more than 4.6 million  $m^3/d$  of potable water [6]. These plants use the reverse osmosis (RO) and the multi-stage flash distillation methods.

\* Corresponding author.

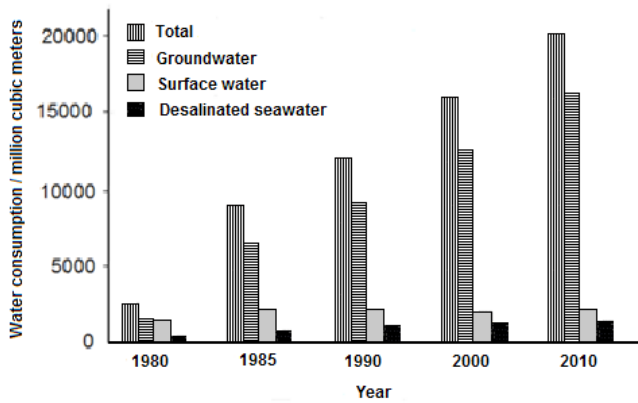


Fig. 1. Water consumption by source in Saudi Arabia. The surface water sources indicated here are captured from infrequent extreme rainfall events through a vast network of dams.

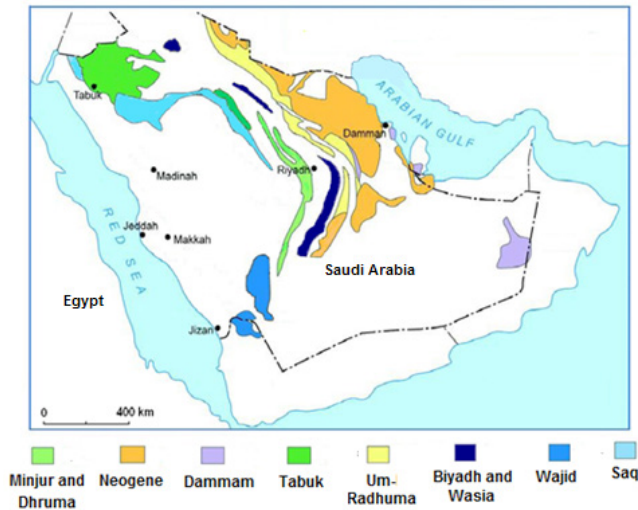


Fig. 2. Major sedimentary deep underground aquifers in Saudi Arabia.

#### 1.4. Silica scaling

Seawater and brackish groundwater must be desalinated prior to their use. RO is the main adopted process in treating these water resources. However, this method is suffering a technical setback from scaling and fouling problems due to the presence of high concentrations of silica in the water resources. This will reduce the lifetime of RO membrane and cause a decline in recovery rate. In order to avert silica scaling, a minimum silica concentration of 10 mg/L in RO feed water is necessary for 95% water recovery rate [7].

Removing silica is very costly, up to 80% of the operating cost in a water treatment plant [8]. Various treatment processes for silica removal exist: coagulation with metal hydroxides [9,10], flocculation [11], nanofiltration [12] and using antiscaling chemicals [13]. In the past few decades, advanced oxidation process has proven its high efficiency in the water treatment processes [14–18]. Table 1 gives the performance in silica removal for various procedures.

Table 1  
Silica removal performances for various procedures

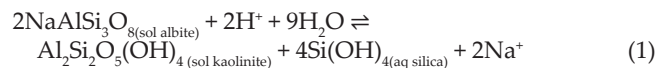
Method	Percentage silica removal	Ref.
Lime softening	47%	[19]
Electrocoagulation	60%–80%	[20]
Coagulation/ultrafiltration	65%	[21]
Magnesium compounds at high pH	Up to 85%	[22]
Coagulation with Al salts	76%	[23]

To the best of our knowledge, the advanced Fenton process (AFP), based on the formation of the powerful hydroxyl radicals when  $H_2O_2$  reacts with  $Fe^0$ , for the removal of silica from water was rarely investigated. This environmentally friendly wastewater treatment method is mostly attractive due to the fact that the hydrogen peroxide and the non-toxic iron are low cost and largely available.

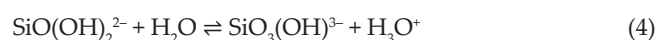
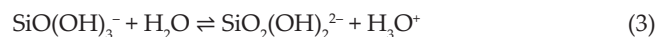
The aim of this preliminary study is to assess the efficiency of this alternative technology for the silica removal from the water feed RO system at Buwaib water treatment plant near the capital city Riyadh.

#### 1.5. Chemistry of silica: silica scaling

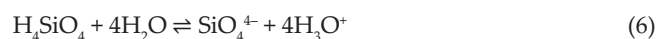
Although the kinetics of silica scaling is very slow, it is a major limiting factor in high-recovery RO technology for the desalination of brackish and seawater. Chemistry of silica is quiet complex. Silica found in surface water (1–20 ppm), well water (10–50 ppm) and brackish well water (120–150 ppm) can be present in amorphous form of solubility 6 ppm or crystalline form of solubility 100–150 ppm at 25°C [24]. It results from the chemical weathering of silica containing rocks such as albite [25].



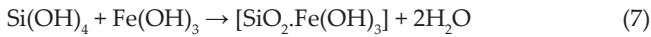
It is present as monosilicic, polysilicic or colloidal silica particles. In aqueous solutions, the monosilicic acid exists as orthosilicic acid  $Si(OH)_4$ . It is the soluble form found in low concentration in all natural waters. It is a weak acid ( $pK_a \sim 9.8$ ) where the four hydroxyl groups are bonded tetrahedrally to the silicon atom [26]. Silica solubility increases with an increase in temperature (up to 300 ppm at 70°C) and an increase in the pH (from 150 ppm at pH 4 to 380 ppm at pH 10) [27]. Successive hydrolysis of the orthosilicic acid produces different anionic species:



with a net dissociation reaction:



Silica is known to condense with a variety of metal ions possessing OH groups [28]. In particular, it aggregates on the ferric hydroxide to form a hydrocomplex, with Fe–O–Si bridge formation [29], which is then removed from the bulk by settling or filtration:



### 1.6. Advanced Fenton process

Under acidic conditions, the AFP uses zero-valent metallic iron with hydrogen peroxide,  $\text{H}_2\text{O}_2$  ( $\text{Fe}^0/\text{H}_2\text{O}_2$ ) [30], to generate ferrous ion. It is considered as a very attractive eco-friendly redox agent for sustainable water treatment [31].

The following reactions sequence occurs:

Oxidation via two electron oxidation of  $\text{Fe}^0$  with generation of ferrous ion:



Catalytic decomposition of ferrous ion by hydrogen peroxide, via Fenton reaction, which generates the hydroxyl free radical  $\text{OH}^\bullet$ :



This radical is one of the most reactive and non-selective oxidizing species when present in aqueous solution [32]. Its standard oxidation potential is 2.7 V in acidic solution and 1.8 V in neutral and alkaline solution [33]. This species can oxidize numerous organic and inorganic contaminants at rates close to diffusion controlled [34,35].

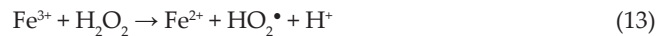
Oxidation by hydroxyl radical, via two electron transfer, of the metal:



Ferric ion hydrolysis as ferric hydroxide:



Catalytic regeneration of ferrous ion:



## 2. Materials and methods

All chemicals used in this work were Analar grade from BDH and were used without any further treatment. Iron metal powder ( $<10 \mu\text{m}$ , 99.99% pure) was from Sigma-Aldrich. Standard solutions of silica were prepared by dissolving sodium metasilicate nonahydrate  $\text{Na}_2\text{SiO}_3 \cdot 9\text{H}_2\text{O}$  (99% purity) in purified and deionized water obtained from Milli-Q water purification system. The pH solution was measured using the Jenway 3510 glass electrode pH-meter and was adjusted when necessary by adding sodium hydroxide or perchloric acid (99% purity).

In this laboratory-scale test, underground raw water samples from the Buwaib water desalination plant fed by the Minjur and Dhurma deep aquifer near the capital city Riyadh (Fig. 2) were selected to investigate the AFP for the silica removal. The conventional jar-test method using a Phipps & Bird Six-Paddle Stirrer (Fig. 3) with six 1-L beakers was performed at room temperature to simulate the mixing and settling conditions found in water treatment plants [36]. The coagulation testing cycle (Fig. 4) was as follows: 5 min continuous rapid stirring at 100 rpm of the water sample mixed with  $\text{Fe}^0$  and  $\text{H}_2\text{O}_2$ , at initial pH, to enhance the homogeneity throughout the 1-L beaker, followed by a slow stirring for 15 min at 20 rpm at equilibrium pH, to allow an efficient coagulation and up to 3-h settling time. For the sake of reproducibility, all the jar tests were performed six times.

After the settling, 5 ml aliquots were withdrawn from the reactor and analyzed for silica content by the silicomolybdate analytical standard protocol [37]. This analytical method is based on the reaction in acidic medium of soluble silica with ammonium heptamolybdate to form a yellow silicomolybdate complex according to the following reaction:

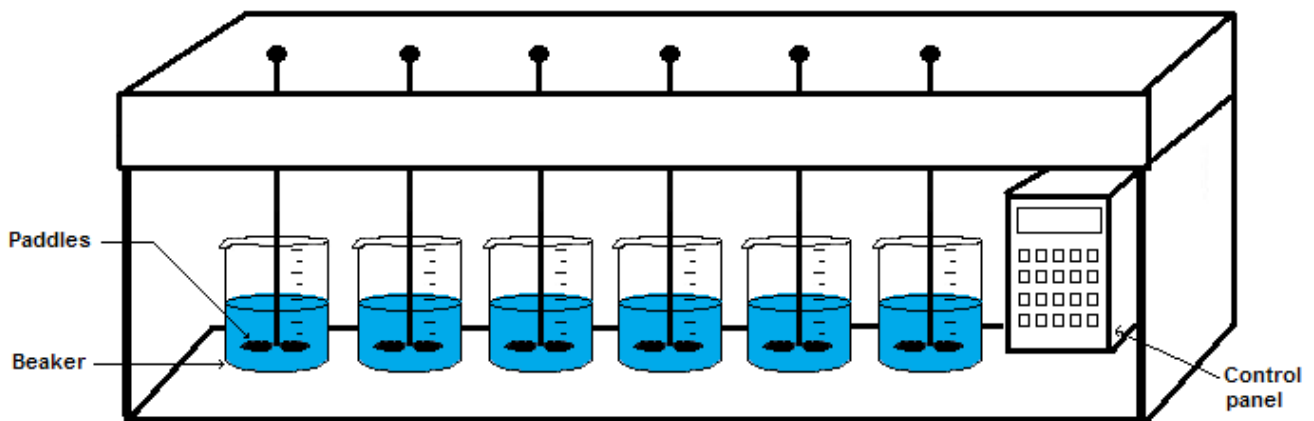
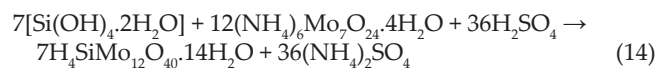


Fig. 3. Diagram of the jar test apparatus with the six 1-L beakers.

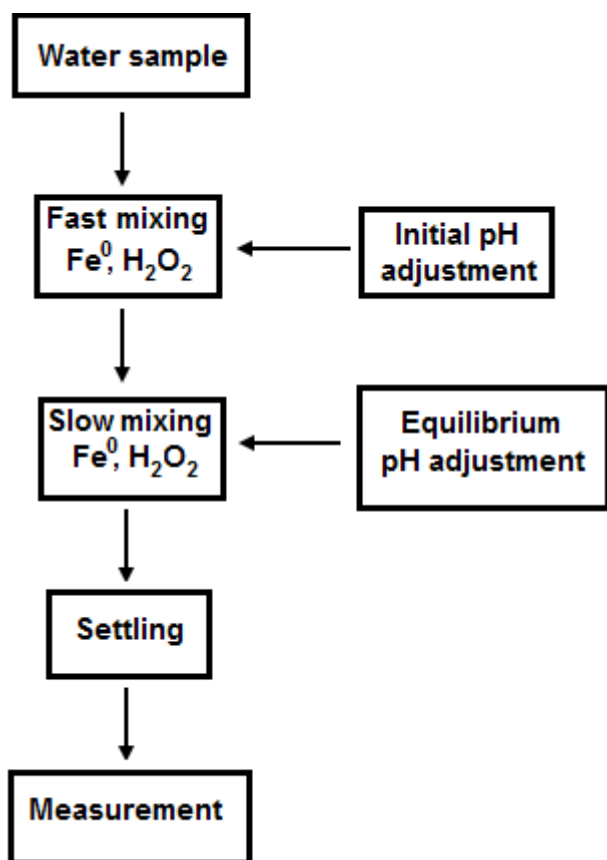


Fig. 4. Block diagram of the jar test method.

which is then reduced to the so-called silicomolybdenum complex blue by ascorbic acid. This complex exhibits a spectrophotometric absorption at 810 nm with a molar extinction coefficient of  $44,700 \text{ L mol}^{-1} \text{ cm}^{-1}$ , and its concentration was measured using its optical absorption at by the UV/visible double beam spectrophotometer Biochrom Libra S32 PC with 1.8 nm bandwidth. It should also be mentioned that this method is only valid to measure soluble silica [37].

### 3. Results and discussion

The silica concentration in the Buwaib water samples was found to be  $54.2 \pm 2.1 \text{ mg/L}$ . At this concentration, the polymerization of silica does not occur, and the dominant species is  $\text{Si(OH)}_4$  [26]. The percentage removal of silica was computed using the following equation:

$$\% R(t) = \frac{C_0 - C(t)}{C_0} \times 100\%$$

where  $C_0$  is the silica concentration before its removal, and  $C(t)$  is its concentration at time  $t$  after its removal.

#### 3.1. Effect of initial pH on silica removal

The initial pH is one of the key factors in the AFP reaction sequence due to its role in governing the stability of  $\text{H}_2\text{O}_2$ , the

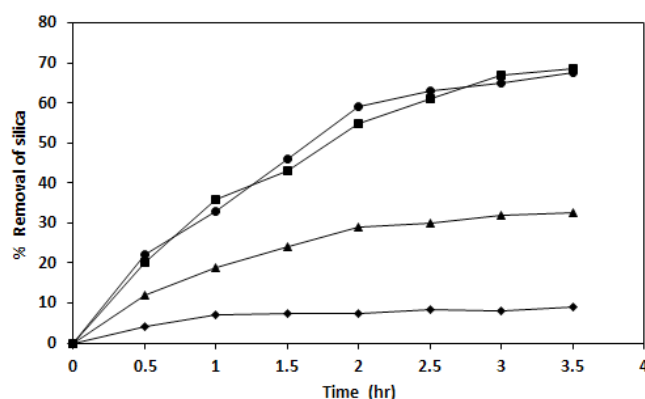
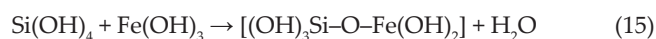


Fig. 5. Effect of  $\text{pH}_{\text{initial}}$  on the percentage removal of silica: (●)  $\text{pH}_{\text{initial}}$  2, (■)  $\text{pH}_{\text{initial}}$  3, (▲)  $\text{pH}_{\text{initial}}$  4, (◆)  $\text{pH}_{\text{initial}}$  5.  $[\text{H}_2\text{O}_2] = 15 \text{ mM}$ ,  $[\text{Fe}^0] = 8 \text{ g/L}$ , equilibrium pH 8.

catalytic action of  $\text{Fe}^{2+}$  and the iron speciation [38]. Removal of silica was determined under initial pH values of 2, 3, 4 and 5, while hydrogen peroxide and iron metal concentrations were fixed to 15 mM and 8 g/L, respectively. Fig. 5 illustrates the  $\% R(t)$  of silica dependence on the initial pH of the water sample. The removal reached a maximum value of  $\sim 70\%$  after a contact time of 3 h for a pH in the range 2. The decreased removal efficiency at higher pH can be attributed mainly to the deficiency of the  $\text{Fe}^{2+}$  catalyst (reaction (10)) and to a lesser extent to the disproportionation of hydrogen peroxide into oxygen and water [39]. To initiate the AFP, a low pH is crucial for the formation of ferrous ions (reaction (1)). During the reactions sequence of AFP, a slight increase in the pH is observed and is attributed to reactions (2)–(4) upon release of hydroxide ions.

#### 3.2. Effect of equilibrium pH on silica removal

Removal of silica was determined under various equilibrium pH in the range 3–11 with an initial pH of 2,  $[\text{H}_2\text{O}_2] = 15 \text{ mM}$  and  $[\text{Fe}^0] = 8 \text{ g/L}$  (Fig. 6). At low equilibrium pH, the adsorption by weak van der Waals attraction of silicic acid on ferric hydroxide results in ligand exchange in which the hydrogen atom the hydroxyl group of the ferric ion is replaced of the silanol group ( $\equiv\text{Si-OH}$ ) of dissolved silicic acid through  $\text{Fe-O-Si}$  bonds [29]. This leads to the formation of the silica-iron complex according to the stoichiometric reaction:



As the equilibrium pH is increased toward the  $\text{p}K_a$  of the silicic acid, this latter dissociates (reaction (6)) into negative silicate anions that adsorb on the ferric cations complex by electrostatic attraction leading to a removal increase reaching a maximum at a pH about 8, which is close to the  $\text{p}K_a$  of the silicic acid. However this removal decreases for  $\text{pH} > 8$  due to increasing competition of hydroxyl anions with silicate anions for positive surface sites of the ferric complex. This is in agreement with the works of Sanciolo et al. [40] and Salvador Cob et al. [9].

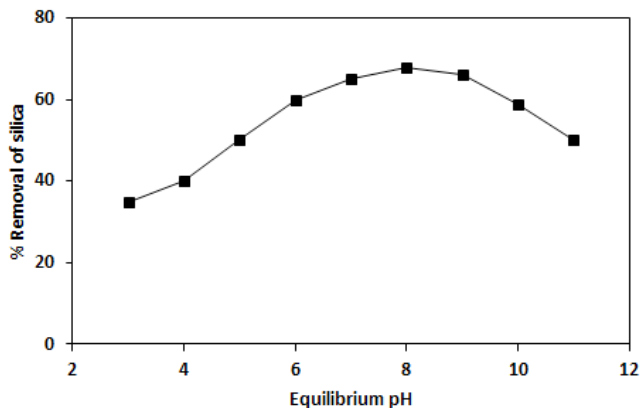


Fig. 6. Equilibrium pH effect on silicate removal by adsorption on ferric hydroxide: initial pH: 2,  $[H_2O_2] = 15$  mM,  $[Fe^0] = 8$  g/L.

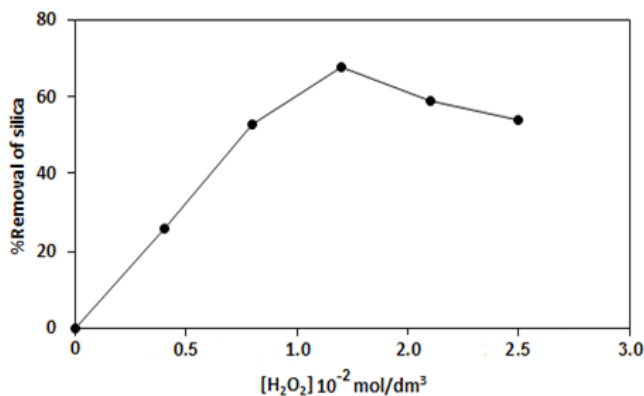


Fig. 7. Effect of  $[H_2O_2]$  on the silica removal:  $[Fe^0] = 8$  g/L, initial pH = 2, equilibrium pH = 8.

### 3.3. Effect of $H_2O_2$ dosage

The effect of hydrogen peroxide concentration on the removal of silica was investigated for different amounts of  $H_2O_2$  ranging from 0 to 25 mM, while keeping the iron dosage at 8 g/L, initial pH at 2 and equilibrium pH at 8. As it can be seen in Fig. 7, hydrogen peroxide affects greatly the percentage removal of silica. At lower concentrations, the removal increases with increasing  $H_2O_2$  concentration. However, at higher concentration, the hydrogen peroxide self-decays (into water and oxygen) and competes with  $Fe^{2+}$  for  $OH^\bullet$  radicals as reported by other workers [41–43]:



$OH^\bullet$  is then consumed, leading to a decrease in the ferric ion concentration and consequently to a reduction in the percentage removal of the silica. It should be noted that  $HO_2^\bullet$  peroxyhydroxyl radical has a less oxidizing effect on  $Fe^0$  and  $Fe^{2+}$  than  $OH^\bullet$  [32].

### 3.4. Effect of iron metal dosage

Removal of silica at different iron metal dosage was investigated in the range 0–10 g/L by keeping the concentration of hydrogen peroxide to 15 mM with an initial pH = 2 and an

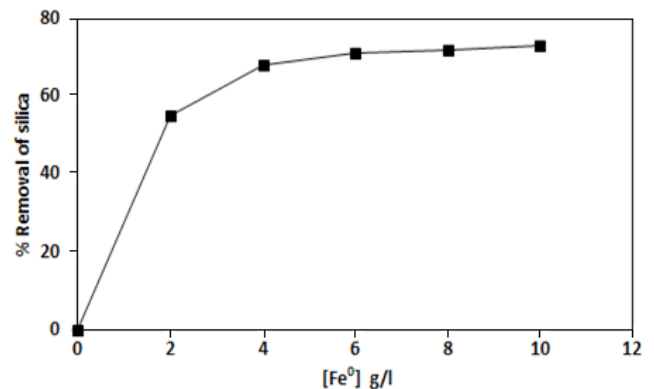


Fig. 8. Effect of metallic iron on the silica removal:  $[H_2O_2] = 15$  mM, initial pH = 2, equilibrium pH = 8.

equilibrium pH = 8. As shown in Fig. 8, higher amount of  $Fe^0$  increases the concentration of  $Fe^{2+}$  ions and consequently the concentration of  $Fe^{3+}$  (following reactions (10) and (11)), which provides a more surface area for silicate co-precipitation for its removal. However, as shown in Fig. 6, the increase in percentage removal is less pronounced at concentration of iron above 2 g/L. A plateau value of around 70% is reached at iron dosage  $\sim 6$  g/L suggesting a saturation in the silica coagulation taking place in the bulk solution as it was found by other workers [44–46]. This saturation in the removal may be attributed to the catalytic regeneration of ferrous ion as in reaction (13).

For a percentage removal as high as 70% obtained in this work, the optimum conditions of the main reaction parameters studied here were the following:  $Fe^0$  dosage: 8 g/L,  $[H_2O_2]$ :  $1.5 \times 10^{-2}$  mol/L, initial pH: 3 and equilibrium pH: 8. It should be noted that Al-Rehaili [47] reached 56% silica removal when using at neutral pH a dose combination of lime, soda ash and sodium aluminate as coagulant aid.

## 4. Conclusion

This laboratory-scale preliminary study investigated the viability of the AFP for the silica removal from the Buwai water desalination plant using hydrogen peroxide and zero-valent iron as the source of catalytic ferric iron.

Under the experimental conditions used in this study, the effect of four reaction parameters considered showed that the percentage removal:

- decreased with increasing initial pH the optimum initial pH is 2;
- increased with increasing equilibrium pH up to the  $pK_a$  of the silicic acid, then decreased at higher pH;
- increased with hydrogen peroxide dosage; however, an excess in hydrogen peroxide lowers the percentage removal; and
- reached a plateau at a concentration around 8 g/L of iron dosage.

This method of water treatment would be viable in arid regions of Saudi Arabia where there is a limited access to freshwater. However, implementation of this promising method on a larger scale will depend on many factors especially operational and economical.

## Acknowledgment

The authors would like to express their thanks to the Nuclear Engineering Department for partially funding this work.

## References

- [1] M. Almazroui, M. Nazrul Islam, P.D. Jones, H. Athar, M. Ashfaqur Rahman, Recent climate change in the Arabian Peninsula: seasonal rainfall and temperature climatology of Saudi Arabia for 1979–2009, *Atmos. Res.*, 111 (2012) 29–45.
- [2] H.K. Al-Zahrani, M.B. Baig, Water in the Kingdom of Saudi Arabia: sustainable management options, *J. Anim. Plant Sci.*, 21 (2011) 601–604.
- [3] J.W. Lloyd, R.H. Pim, The hydrogeology and groundwater resources development of the Cambro-Ordovician sandstone aquifer in Saudi Arabia and Jordan, *J. Hydrol.*, 121 (1990) 1–20.
- [4] S. Chowdhury, M. Al-Zahrani, Characterizing water resources and trends of sector wise water consumptions in Saudi Arabia, *J. King Saud Univ. Sci.*, 27 (2015) 68–82.
- [5] A.G. Maadhah, C.K. Wojcik, Performance study of water desalination methods in Saudi Arabia, *Desalination*, 39 (1981) 205–217.
- [6] SWCC, 2016. Available at: <http://www.swcc.gov.sa/english/AboutSWCC/Pages/About.aspx>
- [7] D.E. Potts, R.C. Ahlert, S.S. Wang, A critical review of fouling of reverse osmosis membranes, *Desalination*, 36 (1981) 235–264.
- [8] P. Pedenau, F. Dang, A New Water Treatment Scheme for Thermal Development: The SIBE Process, International Thermal Operations and Heavy Oil Symposium, Society of Petroleum Engineers, Calgary, Alberta, Canada, 2008.
- [9] S. Salvador Cob, B. Hof, C. Maffezzoni, J. Adamus, W.G. Siegers, E.R. Cornelissen, F.E. Genceli Güner, G.J. Witkamp, Silica removal to prevent silica scaling in reverse osmosis membranes, *Desalination*, 344 (2014) 137–143.
- [10] R. Semiat, I. Sutzkover, D. Hasson, Scaling of RO membranes from silica supersaturated solutions, *Desalination*, 57 (2003) 169–191.
- [11] R. Liu, T. Xue, J. Song, Y. Wang, T. Qi, J. Qu, A. Du, Removal of silicon in acid leaching and flocculation processes during zirconium oxychloride octahydrate production, *Ceram. Int.*, 40 (2014) 8801–8808.
- [12] R. Molinari, P. Argurio, L. Romeo, Studies on interactions between membranes (RO and NF) and pollutants ( $\text{SiO}_2$ ,  $\text{NO}_3^-$ ,  $\text{Mn}^{++}$  and humic acid) in water, *Desalination*, 138 (2001) 271–281.
- [13] R.Y. Ning, T.L. Troyer, R.S. Tominello, Chemical control of colloidal fouling of reverse osmosis systems, *Desalination*, 172 (2005) 1–6.
- [14] J. Bergendahl, S. Hubbard, D. Grasso, Pilot-scale Fenton's oxidation of organic contaminants in groundwater using autochthonous iron, *J. Hazard. Mater.*, 99 (2003) 43–56.
- [15] C.B. Chidambara Raj, H.L. Quen, Advanced oxidation processes for wastewater treatment: optimization of UV/ $\text{H}_2\text{O}_2$  process through a statistical technique, *Chem. Eng. Sci.*, 60 (2005) 5305–5311.
- [16] M. Xu, Q. Wang, Y.I. Hao, Removal of organic carbon from wastepaper pulp effluent by lab-scale solar photo-Fenton process, *J. Hazard. Mater.*, 148 (2007) 103–109.
- [17] R. Matta, K. Hanna, S. Chiron, Fenton-like oxidation of 2,4,6-trinitrotoluene using different iron minerals, *Sci. Total Environ.*, 385 (2007) 242–251.
- [18] F. Lucking, H. Koser, M. Jank, A. Ritter, Iron powder and graphite and activated carbon as a catalyst for the oxidation of 4-chlorophenol with hydrogen peroxide in aqueous solution, *Water Res.*, 32 (1998) 2607–2614.
- [19] S. Al-Mutaz, I.A. Al-Anezi, Silica Removal during Lime Softening in Water Treatment Plant, International Conference on Water Resources and Arid Environment, King Saud University, Riyadh, Saudi Arabia, 2004.
- [20] W. Den, C.J. Wang, Removal of silica from brackish water by electrocoagulation pretreatment to prevent fouling of reverse osmosis membranes, *Sep. Purif. Technol.*, 59 (2008) 318–325.
- [21] H.H. Cheng, S.S. Chen, S.R. Yang, In-line coagulation/ultrafiltration for silica removal from brackish water as RO membrane pretreatment, *Sep. Purif. Technol.*, 70 (2009) 112–117.
- [22] I. Latour, R. Miranda, A. Blanco, Silica removal with sparingly soluble magnesium compounds. Part I, *Sep. Purif. Technol.*, 138 (2014) 210–218.
- [23] I. Latour, R. Miranda, A. Blanco, Silica removal from newsprint mill effluents with aluminum salts, *Chem. Eng. J.*, 230 (2013) 522–531.
- [24] R. Sheikholeslami, S. Tan, Effects of water quality on silica fouling of desalination plants, *Desalination*, 126 (1999) 276–280.
- [25] R. Sheikholeslami, I.S. Al-Mutaz, T. Koo, A. Young, Pretreatment and the effect of cations and anions on prevention of silica fouling, *Desalination*, 139 (2001) 83–95.
- [26] R.K. Iler, *The Chemistry of Silica*, Wiley, New York, 1979.
- [27] P.V. Brady, S.J. Altman, L.K. McGrath, J.L. Krumhansl, H.L. Anderson, pH modification for silica control, *Desal. Wat. Treat.*, 51 (2013) 5901–5908.
- [28] H.E. Bergna, W.O. Roberts, *Colloidal Silica: Fundamentals and Applications*, CRC Press, New York, 2006.
- [29] I. Bremere, M. Kennedy, S. Mhyio, A. Jaljuli, G.-J. Witkamp, J. Schippers, Prevention of silica scale in membrane systems: removal of monomer and polymer silica, *Desalination*, 132 (2000) 89–100.
- [30] J. Bergendahl, T. Thies, Fenton's oxidation of MTBE with zero-valent iron, *Water Res.*, 38 (2004) 327–334.
- [31] B.S. Guimarães, N. Kleemann, S.S. Caldas, F.P. Costa, M.A.K. Silveira, F.A. Duarte, E.G. Primel, Environmentally friendly system for the degradation of multipesticide residues in aqueous media by the Fenton's reaction, *Environ. Sci. Pollut. Res.*, 21 (2014) 584–592.
- [32] G.V. Buxton, C.L. Greenstock, W.P. Helman, A.B. Ross, Critical review of rate constants for reactions of hydrated electrons, hydrogen atoms and hydroxyl radicals in aqueous solution, *J. Phys. Chem. Ref. Data*, 17 (1988) 712–719.
- [33] W.H. Koppenol, J.F. Liebman, The oxidizing nature of the hydroxyl radical. A comparison with the ferryl ion ( $\text{FeO}^{2+}$ ), *J. Phys. Chem.*, 88 (1984) 99–101.
- [34] J.I.Y. Lee, D.A. Lane, Unique products from the reaction of naphthalene with the hydroxyl radical, *Atmos. Environ.*, 43 (2009) 4886–4893.
- [35] J. Sommar, K. Gärdfeldt, D. Strömberg, X. Feng, A kinetic study of the gas-phase reaction between the hydroxyl radical and atomic mercury, *Atmos. Environ.*, 35 (2001) 3049–3054.
- [36] R. Sheikholeslami, J. Bright, Silica and metals removal by pretreatment to prevent fouling of reverse osmosis membranes, *Desalination*, 143 (2002) 255–267.
- [37] T. Coradin, D. Eglin, J. Livage, The silicomolybdic acid spectrophotometric method and its application to silicate/biopolymer interaction studies, *Spectroscopy*, 18 (2004) 567–576.
- [38] I.A. Katsoyiannis, T. Ruettimann, S.J. Hug, pH dependence of Fenton reagent generation and As(III) oxidation and removal by corrosion of zero valent iron in aerated water, *Environ. Sci. Technol.*, 42 (2008) 7424–7430.
- [39] W.G. Kuo, Decolorizing dye wastewater with Fenton's reagent, *Water Res.*, 26 (1992) 881–886.
- [40] P. Sancio, N. Milne, K. Taylor, M. Mullet, S. Gray, Silica scale mitigation for high recovery reverse osmosis of groundwater for a mining process, *Desalination*, 340 (2014) 49–58.
- [41] C.H. Weng, Y.T. Lin, N. Liu, H.Y. Yang, Enhancement of the advanced Fenton process by ultrasound for decolorization of real textile wastewater, *Color. Technol.*, 130 (2013) 133–139.
- [42] J. Anotai, C.-C. Sub, Y.C. Tsai, M.C. Lub, Effect of hydrogen peroxide on aniline oxidation by electro-Fenton and fluidized-bed Fenton processes, *J. Hazard. Mater.*, 183 (2010) 888–893.
- [43] M. Malakootiana, N. Jaafarzadeh, A. Dehdarirad, Efficiency investigation of photo-Fenton process in removal of sodium dodecyl sulphate from aqueous solutions, *Desal. Wat. Treat.*, 57 (2016) 24444–24449.

- [44] A. Al-As'ad, M. Husein, Treatment of steam-assisted gravity drainage water low coagulant dose and Fenton oxidation, *Environ. Technol.*, 35 (2014) 1630–1638.
- [45] M.A. Behnajady, N. Modirshahla, F. Ghanbary, A kinetic model for the decolorization of C.I. Acid Yellow 23 by Fenton process, *J. Hazard. Mater.*, 148 (2007) 98–102.
- [46] N. Daud, U. Akpan, B. Hameed, Decolorization of Sunzol Black DN conc. in aqueous solution by Fenton oxidation process: effect of system parameters and kinetic study, *Desal. Wat. Treat.*, 37 (2012) 1–7.
- [47] A. Al-Rehaili, Comparative chemical clarification for silica removal from RO groundwater feed, *Desalination*, 159 (2003) 21–31.