



Efficiency of *Pseudomonas stutzeri* strain M15-10-3 in the treatment of leather tanning industrial wastewater using gravel-biofilm system

Ebtesam El-Bestawy^{a,*}, Reham Aburokba^b

^aDepartment of Environmental Studies, Institute of Graduate Studies and Research, Alexandria University, 163 Horria Ave. El-Shatby, P.O. Box 832, Alexandria, Egypt, Tel. +203 4295007; Fax: +203 4285793, email: ebtesamelbestawy@yahoo.com

^bDepartment of Biological Sciences, Faculty of Science, King Abdul Aziz University, P.O. Box 42805, Jeddah 21551, Saudi Arabia, email: raburokba@windowslive.com

Received 26 December 2016; Accepted 13 July 2017

ABSTRACT

The study aimed to investigate the ability of *Pseudomonas stutzeri* as biofilm (fixed mode) for decontaminating tannery wastewater based on the remarkable efficiency that planktonic *P. stutzeri* exhibited in the removal of all the contaminants during the batch treatment in a previous study. *P. stutzeri* were fixed on white stone as supporting material and tested as continuous treatment biofilm system for the tannery effluent at different flow rates (30–200 mL/h) for 5 h where samples were collected on hourly interval. Extremely high contaminants concentration of total suspended solids (TSS), total dissolved solids (TDS), biochemical oxygen demand (BOD), chemical oxygen demand (COD), fat, oil and grease (FOG), ammonia (NH₃), nitrates (NO₃), chromium (Cr) and hydrogen sulphide (H₂S) were determined in the raw tannery wastewater. The highest achieved removal efficiencies (RE(s)) by the proposed biofilm system after 5 h were 78.98%, 93.44% and 76.19% for TSS, BOD and FOG, respectively, at 30 mL/h while 82.60%, 25.09%, 97.67% and 38.30% were achieved for COD, NH₃, Cr and H₂S, respectively, at 200 mL/h. On the other hand, TDS, NO₃ and total viable count of *P. stutzeri* increased with time (49.95%, 41.4% and 26.4 fold at 50, 30 and 200 mL/h, respectively) due to bacterial metabolic activity reaching their highest levels after 5 h. Results also revealed huge variations in the RE of all the tested parameters achieved by *P. stutzeri* biofilm system compared with those obtained by the control (bacteria-free system) confirming the efficient role of *P. stutzeri* in removing effluent contaminants. Fixation of *P. stutzeri* enhanced contaminants removal from the highly polluted tannery effluent and protected biofilm bacteria from effluent toxicity, death and the wash out as shown by higher RE(s) for all the tested parameters coupled with shortening of the treatment time (5 h) instead of 7 d in the batch treatment. Therefore, it is evident that the proposed biofilm system with the highly active bacterium *P. stutzeri* represents a very promising, renewable and cheap biotechnology for the treatment of wide range of contaminated effluents not only in the industrial sector but also for domestic and agricultural wastewater.

Keywords: Biofilm; *Pseudomonas stutzeri*; Tannery wastewater; Treatment

1. Introduction

Tanning industry consumes large amounts of water and chemicals (at least 300 kg per ton of hides), and produces wastewater with very high pollution loads [1]. One ton of skin

generally leads to the production of 20–80 m³ of turbid and foul-smelling wastewater. This effluent includes high levels of biochemical oxygen demand (BOD: 1,750 mg/L), chemical oxygen demand (COD: 6,200 mg/L), suspended solids (SS: 5,300 mg/L), total dissolved solids (TDS: 37,000 mg/L), chromium (100–400 mg/L), sulfide (200–800 mg/L), total Kjeldahl nitrogen (273 mg/L), N-NH₃ (153 mg/L), PO₄⁻³ (21 mg/L) fat,

* Corresponding author.

solid wastes and notable pathogen contamination leading to considerable environmental pollution [2,3]. Great variability was observed in the effluent characteristics, depending on the type of hides and skins and the region from which they came, at the time of sampling [4]. Different ranges of tannery pollutants included total nitrogen (927–2,140 mg/L), COD (9,583–13,515 mg/L), ammonium N (149–178 mg/L), sulfide (466.3–794 mg/L) and total chromium (23.3–42.5 mg/L) during the feeding stages [5–7]. Several problems (inhibition of biodegradation) have been encountered during the biological treatment of tannery wastewater because of the high toxicity of chromium and sulphides.

The untreated release of tannery effluents with the high concentrations of pollutants leads to eutrophication and other adverse environmental effects in the receiving water bodies [2,8,9]. Being highly toxic, they affect flora and fauna of the ecosystem and increase the health risk of human beings [10–14].

Microbial cells growing in a biofilm are physiologically distinct from planktonic cells of the same organism, which, by contrast, are single cells that may float or swim in a liquid medium. A biofilm is a complex, organized and heterogeneous aggregation of living microorganisms (bacteria, fungi, algae and protozoa) in which cells are embedded into a self-produced matrix of extracellular polymeric substance (EPS) and adhere to each other and to a surface [15] and associated with wide biotic and abiotic surfaces [16]. Biofilm EPS (slime) is a polymeric substance composed of extracellular DNA, proteins and polysaccharides in various configurations [17]. EPS matrix is a dynamic system that fills and forms the space between cells and is responsible for organization and communication within the biofilm community [18]. It is primarily excreted by bacteria colonizing various surfaces [19]. Biofilms can accumulate on various types of substances (inorganic and organic solutes and particles) from the surrounding water through different processes such as sorption, adhesion, cohesion, uptake of ions and mechanical entrapment of particulate matters [20–22]. Biofilms may also be formed on living or non-living surfaces representing a prevalent mode of microbial life in natural, industrial and hospital settings [17].

Planktonic microbes form a biofilm in response to many factors, which may include cellular recognition of specific or non-specific attachment sites on a surface, nutritional cues or exposure to sub-inhibitory concentrations of antibiotics [23,24]. A biofilm community begins initially when a clean surface is exposed to an aqueous environment and becomes conditioned by the chemical constituents present. Inevitably, microorganisms become associated with the surface, adhere and then attach. Once cells are firmly bound, the activity of the community is dependent on the metabolism (substrate consumption, cellular growth and replication, and synthesis of exopolymers) of each member species under local surface conditions [25]. After formation, biofilms are more stable and extremely resistant against antibiotics [26], heavy metals [27] and temperature increase [28] due to the complexity of biofilm communities as well as immobilization of biofilm members onto the EPS matrix [25,29,30].

Bacteria fixed on biofilm systems are well known for their efficient role in treating tannery wastewater. For example, methanogenic biomass used in anaerobic fixed biofilm reactor to treat tannery wastewater achieved 60%–75% COD

removal and methane yield (0.36 m³ CH₄/kg COD removal) [31]. Similarly, Cr³⁺ tolerant *Acinetobacter* sp., *Aeromonas salmonicida* and *Pseudomonas maltophilia* isolated from tannery wastewater were fixed in a constructed sand–biofilm in repetitive units system and used to reduce high Cr³⁺, BOD and COD under the hazardous threshold concentration prior to the safe discharge into the environment as also found by other workers [32,33]. Solid substrates were used for the fixation of the biofilm directly affecting its efficiency as treatment system. Sundar et al. [34] used a consortium of *Bacillus subtilis* and *Bacillus cereus* biofilm on different substrates (glass beads, pebbles and coarse sand) as a continuous system for the bioremoval of trivalent chromium from tannery effluents at 20 mL/min flow rate, 30°C and pH 4. Biofilm biomass on the substrates was in the following sequence: coarse sand > pebbles > glass beads (4.8 × 10⁷, 4.5 × 10⁷ and 3.5 × 10⁵ CFU/cm²). Biofilms on coarse sand had more surface area and was able to remove 98% of Cr(III) of which 92.60% were adsorbed onto sand biofilms and considered better option for tannery industry, especially during post-chrome-tanning operation. *Pseudomonas aeruginosa*, another Gram-negative bacterium commonly isolated from soil and water, is known for its nutritional and ecological variety. *P. aeruginosa* is also able to escape many stress factors such as heavy metals [35], antibiotics [36,37] and ultraviolet (UV) light [38,39]. However, it is always preferable to use the indigenous microorganisms that are well adapted to all stresses in the polluted environment. In the present study, *Pseudomonas stutzeri* and *Providenciarettgeri* isolated among 17 other bacteria from heavily contaminated tannery effluent in Jeddah, Saudi Arabia, and considered the most efficient in removing not only Cr (93.66% RE) but all the tested parameters in the batch experiment which also included three exogenous strains (*Pseudomonas fluorescens/putida*) [40]. Therefore, it was highly recommended to use such effective, highly resistant and acclimatized indigenous species in the proposed biofilm system for the treatment of highly polluted wastewater.

The main aim of the present study was to investigate the ability of *P. stutzeri* M15-10-3 fixed biofilm to decontaminate tannery wastewater in a proposed continuous system.

2. Materials and methods

2.1. Bacterium

P. stutzeri strain M15-10-3 originally isolated and identified from heavily contaminated tannery effluent was selected based on the remarkable efficiency that its planktonic form exhibited in the removal of all contaminants during batch treatment [40]. The selected bacterium was maintained on nutrient agar (NA) medium and prior to each experiment, the culture was reactivated overnight.

2.2. Biofilm system construction

Two separating funnels (1 L each) were sealed at the bottom by a porous net ($d < 1$ mm) and supplied with a flow controller (tap) at the outlet (Figs. S1 and S2). They were sterilized by immersing in 75% ethyl alcohol overnight, rinsed twice with absolute ethanol, and rinsed five times with sterile distilled water and then dried in a sterile condition. White

stone (with average particle size 1.6 mm in diameter) was used as supporting material after thorough washing, rinsing and sterilizing four times at 121°C for 15 min. Each funnel was packed with sterile stone up to 80% of their height leaving the top 20% free. After packing, one column was used as a control where only wastewater was supplied during the treatment stage, while the other column was inoculated with 800 mL dense overnight reactivated *P. stutzeri* liquid culture (density 5.5×10^9 CFU/mL) and left 10 d to allow bacterial cells adhesion forming the biofilm. The two columns were connected with an up flow air supply, which was adjusted to operate alternately for 1 h and pause for 2 h.

2.3. Determination of population dynamics

The seeded column was left as a batch culture for 10 d at pH 7 and temperature ranged between 20°C and 24°C (room temperature). After 10 d, a sample from the biofilm column was collected every 24 h, serially diluted (up to 10^{-8}) and 1,000 μ L of the appropriate dilution was cultured under aseptic conditions on NA plates and incubated for 24 h at 37°C. Bacterial plate counts were recorded every day till constant count was obtained for three consecutive days.

2.4. Operation conditions

Raw samples were treated using the biofilm at different flow rates (30, 50, 100 and 200 mL/h). At each flow rate, samples were collected from both the biofilm and bacteria-free (control) columns every hour for five consecutive hours. After treatment, all samples were characterized for the same parameters as for the raw water and the efficiency of the treatment using the proposed biofilm for these contaminants was calculated.

2.5. Characterization of the raw and treated industrial effluent

Characterization of the wastewater before and after the proposed treatment included its pH, temperature, dissolved oxygen (DO) content, total suspended solids (TSS), total dissolved solids (TDS), biochemical oxygen demand (BOD), chemical oxygen demand (COD), fat, oil and grease (FOG), sulfides (H_2S), ammonia (NH_3), nitrate (NO_3), total viable count of bacteria (TVC) and total chromium (Cr) all of which were determined using the standard techniques described in Standard Methods for the Examination of Water and Wastewater [41]. Removal efficiency was calculated to determine the effectiveness of the remediation process according to the following equation:

$$\text{Removal efficiency (RE \%)} = C_0 - RC/C_0 \times 100 \quad (1)$$

where C_0 = initial concentration before treatment (zero time), RC = residual concentration after treatment at each exposure time.

2.6. Temperature, pH and total dissolved solids

Temperature, pH and TDS were determined by using digital thermometer and laboratory Bench Meter.

2.7. Total suspended solids

TSS were determined using simple, direct spectrophotometer (Hach Dr 5000) method (630 Suspended Solids) other than gravimetric one that requires filtration or ignition and weighing steps which is often used for checking in-plant processes. TSS were measured at 810 nm.

2.8. Fat, oil and grease

Determination of total content of grease and oily substances was carried out using the partition gravimetric method described by Clesceri et al. [41]. Fatty acid composition of the oily content was extracted using *n*-hexane followed by methylation and then determined using gas chromatograph model 8400GC, fitting with flame ionization detector and fused silica capillary column. Oil and grease content were calculated according to the standard procedure [41].

2.9. Biochemical oxygen demand

Method 5210 B was used for BOD₅ determination as described in the Standard Methods for Examination of Water and Wastewater [41]. BOD₅ can be calculated as follows:

$$\text{BOD}_5, \text{ mg/l} = \frac{D1 - D2}{P} \quad (2)$$

where D1 = DO of diluted sample immediately after preparation in mg/L, D2 = DO of diluted sample after 5-d incubation at 20°C in mg/L, P = decimal volumetric fraction of sample (300 mL).

2.10. Chemical oxygen demand

Closed Reflux Colorimetric Method 5220 D was used for COD determination using potassium dichromate as chemical oxidant as described in the Standard Methods for Examination of Water and Wastewater [41]. Colour developed was measured at 620 nm using DR/5000 HACH spectrophotometer DR/2010 HACH spectrophotometer and the concentration was calculated from the slope of the standard curve.

2.11. Determination of chromium

Chromium in tannery wastewater was digested using concentrated HNO_3 and determined following colorimetric method described by Clesceri et al. [41] using spectrophotometer (HACH DR 5000) at 357.9 nm wavelength.

2.12. Sulfides

Sulfide was determined using methylene blue method (4500-S2- D) described in the Standard Methods for the Examination of Water and Wastewater [41]. Hydrogen sulfide and acid-soluble metal sulfides react with *N,N*-dimethyl-*p*-phenylenediamine sulfate to produce methylene blue. The intensity of the blue colour is proportional to the sulfide concentration. Ammonium phosphate is added after

colour development to remove ferric chloride colour. Sample S^{2-} was measured at 665 nm and results obtained in mg/L S^{2-} .

2.13. Ammonia

Ammonia was determined using phenate method (4500-NH₃ F) described in the Standard Methods for the Examination of Water and Wastewater [41]. After immediate fixation upon collection of the samples and determined spectrophotometrically using the indophenol blue technique. Ammonia compounds combine with sodium hypochlorite and alkaline solution of phenol and disodium nitroprusside dihydrate to form monochloramine, which reacts with salicylate forming 5-aminosalicylate that is oxidized in the presence of a sodium nitroprusside catalyst to form indophenol blue. The blue colour developed after 2 h was measured at 665 nm. The results were expressed as mg/L.

2.14. Nitrate

Nitrates were determined using cadmium reduction method (4500-NO₃) described in the Standard Methods for the Examination of Water and Wastewater [41]. NO₃ was reduced almost quantitatively to nitrite (NO₂) in the presence of cadmium (Cd). This method uses commercially available Cd granules treated with copper sulfate (CuSO₄) and packed in a glass column. The NO₂ produced was converted into a reddish purple azo dye formed by adding 1 mL of sulfanilamide reagent to 50 mL of water sample followed by 1 mL of N-(1-naphthyl)-ethylenediamine dihydrochloride solution (NED) to form a highly coloured azo dye after 30 min. Colour density was measured spectrophotometrically at wavelength 540 nm. The nitrite content was expressed in mg/L. The concentration of nitrate in each sample was calculated after correction of nitrite concentration of the same sample.

2.15. Total viable count of bacteria

Raw and treated tannery samples were sequentially diluted, cultured (three replicas each) using the pour plate technique of the standard heterotrophic plate count method [41] in NA medium and incubated at 35°C for 24 h. Colony forming units (CFU) of the TVC were recorded (Stuart colony counter protected by BioCote) and averages were calculated.

2.16. Statistical analysis

Correlation coefficients (Friedman test) were used to determine the relations among the different contaminants present in the raw and treated tannery effluents.

3. Results

P. stutzeri population dynamics was determined to define the maturity of the biofilm. Biofilm considered mature after 16 d when bacterial total viable count recorded was constant for three consecutive readings as shown in Table 1. After maturation, raw samples were treated using the biofilm at different flow rates (30, 50, 100 and 200 mL/h) for 5 h with 1 h interval sampling.

Table 1

Population dynamics of *P. stutzeri* in the medium used for the biofilm formation

Time (d)	Total viable count (CFU/mL)
0	5.5×10^9
9	1.1×10^{10}
10	1.3×10^{10}
12	1.4×10^{10}
13	1.4×10^{10}
14	1.5×10^{10}
15	1.5×10^{10}
16	1.5×10^{10}

3.1. DO and pH levels

Very low DO concentrations (0.32–0.35 mg/L) were recorded in the raw effluents. These values were gradually reduced with increasing exposure time in biofilm and control systems reaching their lowest levels (0.29 and 0.32 mg/L) after the 5th h especially in the biofilm system at the flow rate 200 and 30 mL/h respectively (data are not shown). This may be attributed to consumption of DO by biofilm bacterium during biodegradation of the included organic contaminants. On the other hand, no significant variations were noticed in the pH values among the tested flow rates, exposure times in the biofilm (7.50–7.76) or control (7.51–7.77) systems before or after the remediation process. (data are not shown). It is well known that pH is a very important factor controlling microbial activity especially under heavy metals stress as in the present study where the highly toxic Cr exists at very high levels. Moreover, metals biosorption processes on microbial dead cells are particularly sensitive to changes in physical conditions such as pH or ionic strength [42].

3.2. Total dissolved solids

Significant TDS increases were recorded during treatment using batch mode (free-living bacteria) where biodegradation resulted in breaking down of complex contaminants into simple dissolved salts. However, using biofilm system decreased TDS increases except at the flow rates 50 and 100 mL/h (Fig. 1(A)). Raw tannery effluent had TDS range of 10,400–10,430 mg/L. TDS increased with time recording the highest TDS additions (49.95% and 49.28%) after 5 working h at 50 and 100 mL/h, respectively, resulting in 15,610 and 15,560 mg/L RC respectively. Much lower TDS additions (0.63% and 0.26%) were recorded using the control system at 30 and 200 mL/h flow rate after 5 working h yielding 10,465 and 10,457 mg/L RC, respectively.

Levels of TDS before and after the treatment are far exceeding the maximum permissible limit (MPL) of 2,000 mg/L.

Increasing TDS concentration in the tannery effluent after biofilm remediation is an expected result due to oxidation of the available ammonia (NH₃) to dissolved nitrate (NO₃) salt which increases TDS concentration in treated wastewater. It is well known that organic matter removal is more affected by changes in salinity than the changes in hydraulic retention time or organic loading rate [5]. So, it is essential to keep TDS changes to a minimum through a technique which converts nitrates into nitrogen gas.

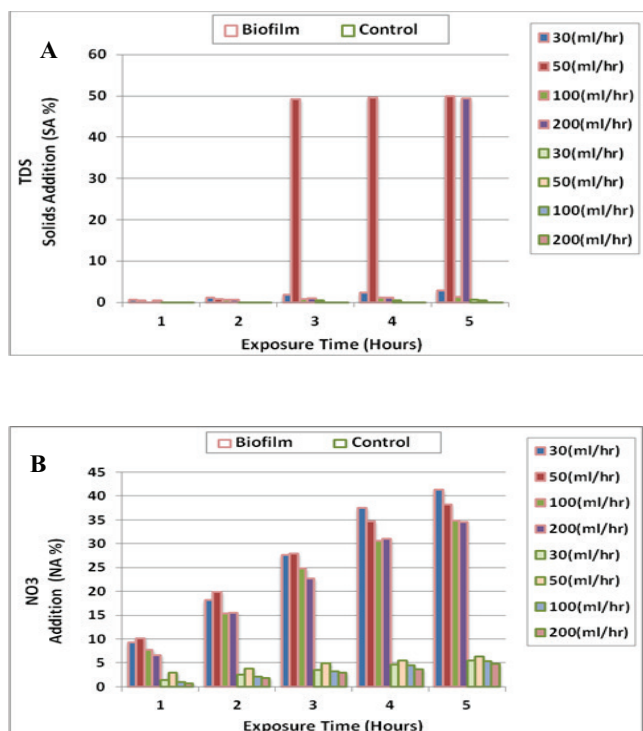


Fig. 1. TDS increases (%) (A) and NO₃ (B) in the treated tannery wastewater using biofilm and control systems at different flow rates and exposure times.

3.3. Nitrate

Raw tannery effluent had NO₃ range of 34.0–37.3 mg/L. As shown with the TDS, RC of nitrate showed general increases with time due to nitrification process where ammonia available in the tannery effluent oxidized into NO₃ reached the highest levels (41.4%) after 5 h at 30 mL/h. Much lower NO₃ addition (6.39%) was obtained using the control system at flow rate of 50 mL/h after 5 working h (Fig. 1(B)). But fortunately, levels of NO₃ before and after the treatment are below the MPL (50 mg/L) which is good for safe discharge.

3.4. Total suspended solids

Raw tannery wastewater recorded very high TSS level (15,610–15,700 mg/L). Treatment using the proposed biofilm system revealed high TSS removals regardless of the flow rate (Fig. 2(A)). The highest RE recorded 78.98% compared with 0.57% achieved by the control both at 30 mL/h after 5 h. The lowest obtained RC of the TSS was 3,300 mg/L which is 55 folds increase than MPL (60 mg/L) of the TSS. Accordingly, TSS could not reach safe limit for discharging during the experiment duration (5 h).

3.5. Biochemical oxygen demand

High BOD levels (3,170–3,200 mg/L) were recorded in the raw tannery wastewater. Both systems (biofilm and control) showed positive relation between RE of the BOD and exposure times. Biofilm system achieved the highest BOD removal (93.44%) compared with the extremely low (2.19%) removal

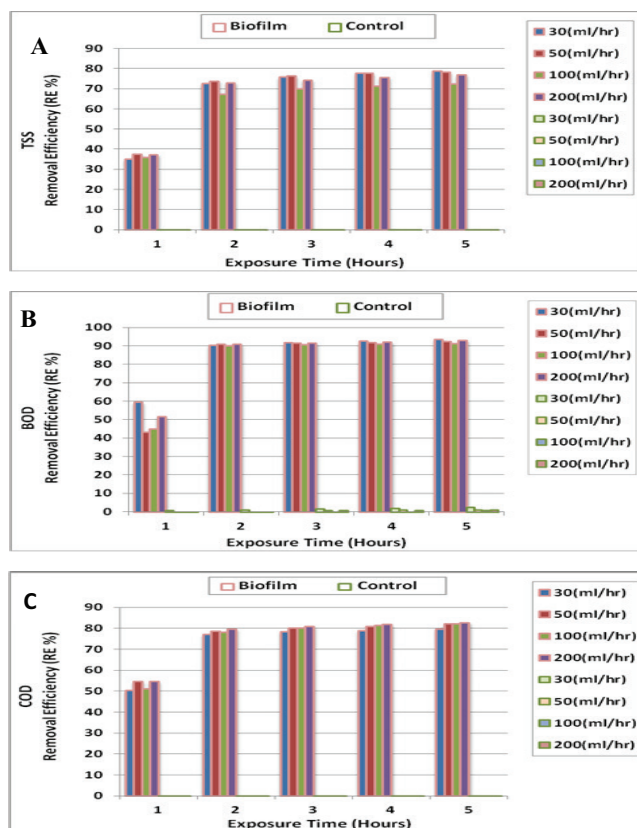


Fig. 2. Removal efficiency (RE %) of TSS (A), BOD (B) and COD (C) from the tannery wastewater using biofilm and control systems at different flow rates and exposure times.

obtained by the control system both at 30 mL/h flow rate after 5 running h (Fig. 2(B)). This is mainly attributed to the high capability of *P. stutzeri* for organic matter degradation. Treatment using biofilm resulted in TSS residues of 210 mg/L which is 3.5 folds higher than the MPL of BOD (60 mg/L).

3.6. Chemical oxygen demand

Extremely high COD levels (22,410–22,500 mg/L) were recorded in the raw tannery wastewater. COD removal followed a general increasing trend with increasing exposure time in the biofilm and control systems. The biofilm recorded highest COD removal of 82.60% at the highest flow rate (200 mL/L) after 5 exposure h compared with almost the negligible activity (0.33% RE) recorded by the control after the same time at 30 mL/L due to the lack of specialized bacterial activity (Fig. 2(C)). However, no significant variation in the RE of COD (79.6%, 82.17% and 82.1%) was noticed at the other tested flow rates (30, 50 and 100 mL/L respectively) after 5 h. Results of the biofilm system showed huge reduction in the COD (from 22,410 to 3,900 mg/L) but levels still 39 fold higher than MPL of COD (100 mg/L) for the safe discharge.

3.7. Fat, oil and grease

Extremely high FOG levels (4,165–4,200 mg/L) were recorded in the raw tannery wastewater. FOG removal in both biofilm and control systems positively related with increasing

exposure time. The biofilm achieved the highest (76.1%) FOG removal at 30 mL/L after 5 h whereas nearly no activity (0.71% RE) was recorded by the control system (Fig. 3(A)). Results of the biofilm system showed significant reduction in the FOG (from 4,200 to 1,000 mg/L) but levels still 10 fold higher than MPL (10 mg/L) of FOG for the safe discharge.

3.8. Ammonia

Raw tannery effluent had NH_3 range of 29.1–30.0 mg/L that showed general trend of increasing RE of NH_3 with increasing exposure time. Biofilm achieved the highest RE of NH_3 (25.09%) after 5 working h at 200 mL/h with lower efficiencies at other tested flow rates compared with much lower NH_3 removal (6.33%) obtained using the control system at 30 mL/h after 5 h (Fig. 3(B)) which indicated the important role of biofilm bacteria in the bioremediation process. Levels of NH_3 before and after the treatment are above the Egyptian MPL (3 mg/L).

3.9. Chromium

Raw tannery wastewater recorded contained high and dangerous Cr level (2,185–2,200 mg/L). The bulk RE% of Cr was achieved after the 1st h at all the tested flow rates in both biofilm and control systems. The highest RE of Cr (97.67%) was recorded by biofilm bacteria at the highest flow rate (200 mL/h) which is equivalent to 51 mg/L (Fig. 3(C)). On the other hand, extremely minor RE of Cr (1.91%) was achieved by the control at the lowest flow rate (30 mL/h). According to the Cr MPL (2 mg/L) the treated effluent still has 25.5 fold increases in the Cr content which required longer exposure to reach the safe limits.

3.10. Hydrogen sulfide

H_2S concentration in raw tannery wastewater recorded a range of 18.5–19.2 mg/L. The highest RE of H_2S (38.30%) was recorded by biofilm bacteria (equivalent to 11.6 mg/L) while minor H_2S removal (5.21%) was achieved by the control both after 5 working h at 100 and 30 mL/h flow rate, respectively (Fig. 3(D)). According to the H_2S MPL (1 mg/L), the treated effluent still has 11.6 fold increases in the H_2S content which required longer exposure and more aeration to help oxidize H_2S and to reach the safe limits.

3.11. Total viable count of bacteria

Fixation of bacteria on solid medium as a biofilm has many advantages. It enhances the bacterial growth, reduces wastewater toxicity and increase bacterial resistance towards the involved contaminants. Bacteria either in the biofilm or control systems had general trend of growth stimulation with time (5 h) with no inhibition at all which is opposite to their growth in a batch mode. Growth of biofilm bacteria stimulated reaching 26.4 fold increases at the highest flow rate (200 mL/h) after 5 working h (Table 2). The growth of the bacteria in the control system was much lower than that of the biofilm but behaves in a similar way where it is gradually stimulated reaching the highest growth density (1.25 f) after 5 h but at the lowest flow rate (30 mL/h).

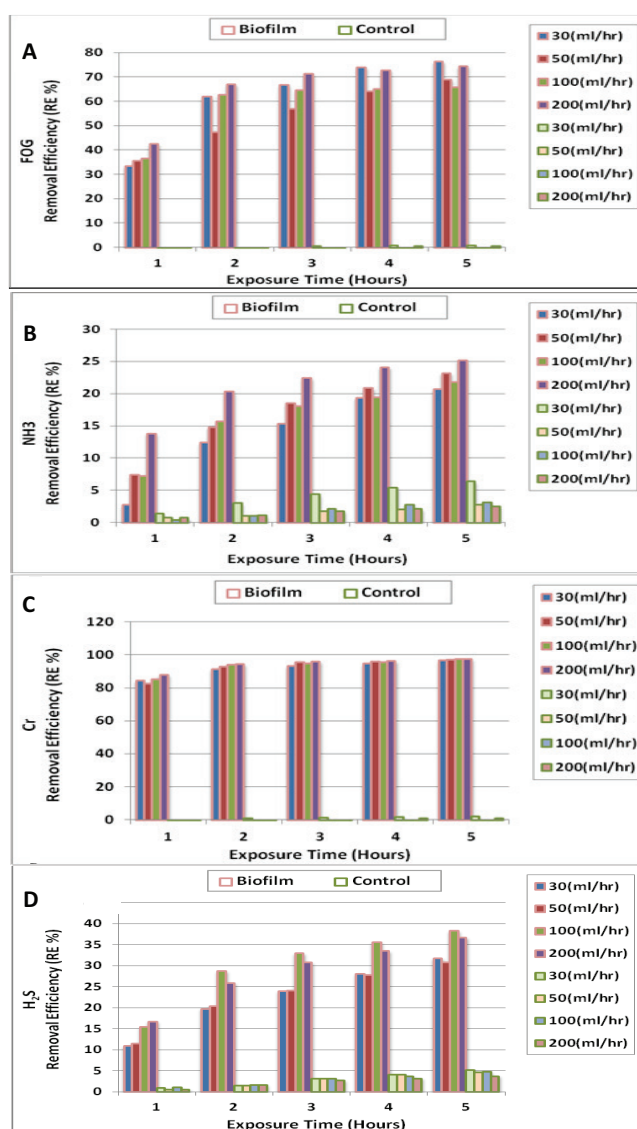


Fig. 3. Removal efficiency (RE %) of FOG (A), NH_3 (B), Cr (C) and H_2S (D) from tannery wastewater using biofilm and control systems at different flow rates and exposure times.

3.12. Statistical analysis

Friedman test was used to correlate between the different tested flow rates and the efficiency of contaminants removal during the continuous treatment using the proposed fixed biofilm system (Fig. 4). Analysis showed no significant differences between the four different flow rates since p value = 0.596 which is greater than $\alpha = 0.05$, we accept the null hypothesis.

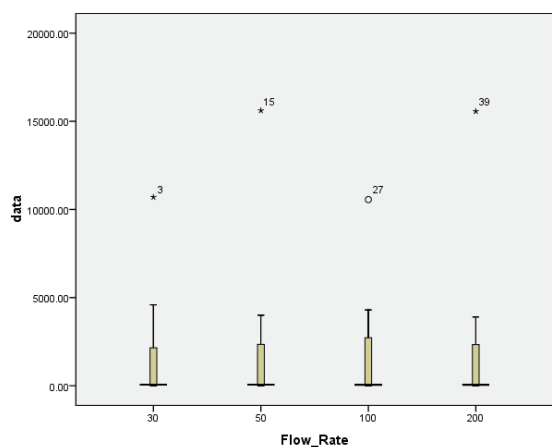
4. Discussion

Biological treatment using microbial biofilms considered highly efficient, eco-friendly and an economic alternative to the conventional physicochemical processes [40,42–44]. Biofilms are excellent pollution indicators due to their (a) ubiquity on almost any surface in the water, (b) sessile mode

Table 2
Stimulatory and/or inhibitory effects of tannery effluent on biofilm and control bacteria at different flow rates

Exposure time (h)	Biofilm system							
	30 (mL/h)		50 (mL/h)		100 (mL/h)		200 (mL/h)	
	RC	S&I%	RC	S&I%	RC	S&I%	RC	S&I%
Zero	0.5×10^9		0.5×10^9		0.5×10^9		0.5×10^9	
1	7.5×10^9	15 f	7.5×10^9	15 f	8.2×10^9	16.4 f	8.4×10^9	16.8 f
2	8.5×10^9	17 f	8.1×10^9	16.2 f	9.5×10^9	19 f	10.5×10^9	21 f
3	9.2×10^{10}	18.4 f	8.9×10^9	17.8 f	10.4×10^9	20.8 f	11.7×10^9	23.4 f
4	10.4×10^9	20.8 f	9.5×10^9	19 f	11.5×10^9	23 f	12.5×10^9	25 f
5	11.3×10^9	22.6 f	10.4×10^9	20.8 f	12.1×10^9	24.2 f	13.2×10^9	26.4 f
Exposure time (h)	Control							
	30 (mL/h)		50 (mL/h)		100 (mL/h)		200 (mL/h)	
	RC	S&I%	RC	S&I%	RC	S&I%	RC	S&I%
Zero time	4.97×10^8		5.13×10^8		5.24×10^8		5.31×10^8	
1	5.10×10^8	1.03 f	5.34×10^8	1.04 f	5.42×10^8	1.03 f	5.48×10^8	1.03 f
2	5.36×10^8	1.08 f	5.48×10^8	1.07 f	5.58×10^8	1.06 f	5.64×10^8	1.06 f
3	5.50×10^8	1.11 f	5.67×10^8	1.11 f	5.71×10^8	1.09 f	5.82×10^8	1.10 f
4	5.86×10^8	1.18 f	5.82×10^8	1.13 f	5.88×10^8	1.12 f	5.99×10^8	1.13 f
5	6.21×10^8	1.25 f	6.02×10^8	1.17 f	6.03×10^8	1.15 f	6.24×10^8	1.18 f

Note: The lowest growth stimulation GS% and the highest growth stimulation GS (%).



*Friedman Test			
Ranks		Test Statistics*	
	Mean Rank		
f200	2.08	N	12
f100	2.75	Chi-Square	1.888
f50	2.62	df	3
f30	2.54	p- value	0.596

Fig. 4. Friedman test.

of growth that reflects the actual habitat conditions, (c) short life cycle that enables a more rapid response to environmental changes than in higher level organisms, (d) species diversity in the community with various environmental tolerances and (e) the relative ease of collecting biofilm samples [45,46]. Biofilm provides biological treatment with many advantages over their free-living counterparts. These advantages include enhancement of contaminants removal, reduction of treatment time and protection of biofilm bacteria from effluent toxicity, death and the wash out of bacterial cells. In the present study, these advantages were clearly shown where higher RE(s) for all the tested parameters coupled with shortening of the treatment time (5 h) instead of 7 d.

P. stutzeri were selected to be fixed in a biofilm system based on its great performance in the removal of all the

contaminants during the batch treatment through either biodegradation of all the organic contaminants or biosorption of inorganic contaminants such as Cr. This finding is supported by many workers who documented the superior resistance and ability of the genus *Pseudomonas* in degradation and accumulation of environmental pollutants. In a recent study, *P. aeruginosa* was able to escape many stress factors such as heavy metals [36]. Two general trends were noticed while treating tannery effluent using the proposed biofilm. The first trend was increasing the RE of all the tested parameters with time and the second was the huge variations in the RE of all the tested parameters achieved by *P. stutzeri* biofilm system compared with those obtained by the control (bacteria-free system) confirming the efficient role of *P. stutzeri* in removing effluent contaminants.

Also, clear variations but no specific trends were noticed in the RE of all the tested parameters at the different flow rates where the highest REs were achieved at the lower flow rates for some parameters and at the higher flow rates for other parameters.

Extremely high contaminants levels were determined in the raw tannery wastewater. Surprisingly, as high as 78.98%, 93.44%, 82.60%, 76.19%, 25.09%, 97.67% and 38.30% were achieved as highest RE% for TSS, BOD, COD, FOG, NH₃, Cr and H₂S, respectively, by the proposed biofilm system after only 5 h which considered superior achievements. Three parameters were increased with time due to bacterial metabolic activity reaching their highest levels after 5 h. These included TDS (49.95% at 50 mL/h), NO₃ (41.4% at 30 mL/h) and the TVC of *P. stutzeri* (26.4 f at 200 mL/h). Huge amounts of the tested contaminants were removed in such short time leaving residues (mg/L) of 15,570–15,610 (TDS); 3,300 (TSS); 210 (BOD); 3,900 (COD); 1,000 (FOG); 21.8 (NH₃); 48.1 (NO₃); 51 (Cr); 11.6 (H₂S) and 13.2 × 10⁹ CFU of *P. stutzeri* /mL. However, extremely lower RE(s) of the tested contaminants were removed by the control system as follows: 0.57%, 2.19%, 0.33%, 0.71%, 6.33%, 1.91% and 5.21% for TSS, BOD, COD, FOG, NH₃, Cr and H₂S, respectively, at the same exposure time. Moreover much lower increases were achieved for TDS (0.63%), NO₃ (6.39%) and TVC (1.25 f) in the tannery wastewater treated with the control system confirming the remarkable ability of *P. stutzeri* biofilm. This is mainly attributed to the great characteristics that the genus *Pseudomonas* possesses including resistance against heavy metals [36], antibiotics [37,38] and UV light [39].

Similarly and in consistent with the present study *P. maltophilia* isolated from tannery wastewater were fixed in a constructed sand–biofilm system and used to reduce high Cr³⁺, BOD and COD contamination load in that effluent. Results indicated the advantage of using the biofilm system in repetitive units to reduce chromium contamination in tanning wastewater to a level under the hazardous threshold concentration prior to the safe discharge into the environment. In addition, this type of bioremediation has already proven itself to be a cost-effective and more beneficial compared with chemical and physical methods for managing wastes and environmental pollutants [33,34,47]. It was reported also that through the metabolic activities of the biofilm system, degradable organic matter present in the surrounding water is gradually broken down [48] as shown by high biodegradation for the BOD, organic nitrogen compounds and FOG in the present study.

Biofilm systems have proved to play an important role in the removal of contaminated industrial wastewater [49–51]. Metabolic activities of the biofilm system gradually degrade organic matter and accumulate/absorb inorganic contaminants contained in the surrounding wastewater [48]. Biofilm can efficiently be used to monitor the impact of pollution on the biofilm community (i.e., biomass, diversity, presence or absence of species) and monitoring the self-accumulation of toxic elements in biofilm dry mass [52–54], therefore they represent an important part in the aquatic ecosystem and wastewater treatment systems. Pollutants are processed and eliminated by means of the complex food chain established within the biofilm [55]. Different types of biofilm reactors have been used for biological treatment of water and wastewater [56].

5. Conclusion

Tannery wastewater showed extremely high levels of all the tested parameters that make it one of the strongest industrial effluents that has high pollution potential and dangerous effects on the receiving environments and also creates many difficulties in the treatment. Fixation of bacteria on solid medium as a biofilm showed many advantages over their planktonic free-living counterparts. It enhances the bacterial growth, reduces wastewater toxicity and increases bacterial resistance towards the involved contaminants. Considering the very short time that biofilm runs for (5 h), it seems that the proposed biofilm system is very efficient for treating the tannery effluents. This system could reach higher removal for all the tested parameters reaching acceptable limits for safe discharge by applying longer exposure time, using two or three biofilm units in sequence and using a preliminary oxidation step before biological treatment that could reduce levels of contaminants. The proposed biofilm system with the highly active bacterium *P. stutzeri* represents a very promising, renewable and cheap biotechnology for the treatment of wide range contaminated effluents not only in the industrial sector but also for domestic and agricultural wastewater.

References

- [1] L.A.H.M. Verheijen, D. Weirsem, L.W. Hwshoffpol, J. Dewit, Live Stock and the Environment: Finding a Balance Management of Waste from Animal Product Processing, International Agriculture Centre, Wageningen, the Netherlands, 1996.
- [2] S. Leta, F. Assefa, L. Gumaelius, G. Dalhammar, Biological nitrogen and organic matter removal from tannery wastewater in pilot plant operations in Ethiopia, Appl. Microbiol. Biotechnol., 66 (2004) 333–339.
- [3] S. Kongjao, S. Damronglerd, M. Hunsom, Simultaneous removal of organic and inorganic pollutants in tannery wastewater using electro coagulation technique, Korean J. Chem. Eng., 25 (2008) 703–709.
- [4] O. Lefebvre, N. Vasudevan, M. Torrijos, K. Thanasekaran, R. Moletta, Halophilic biological treatments of tannery soak liquor in a sequencing batch reactor, Water Res., 39 (2005) 1471–1480.
- [5] W.M. Wiegant, T.J.J. Kalker, V.N. Sontakke, R.R. Zwaag, Full scale experience with tannery water management: an integrated approach, Water Sci. Technol., 39 (1999) 169–176.
- [6] M. Wiemann, H. Schenk, W. Hegemann, Anaerobic treatment of tannery wastewater with simultaneous sulphide elimination, Water Res., 32 (1998) 774–780.
- [7] M.L.M. Stoop, Water management of production systems optimized by environmentally oriented integral chain management: case study of leather manufacturing in developing countries, Technovation, 23 (2003) 265–278.
- [8] H.H. Someda, E.A. El-Shazly, R.R. Sheha, The role of some compounds on extraction of chromium (VI) by a mine extractants, J. Hazard. Mater., B117 (2005) 213–219.
- [9] G. Durai, M. Rajasimman, Biological treatment of tannery wastewater – a review, J. Environ. Sci. Technol., 4 (2011) 1–17.
- [10] B. Chattopadhyay, S. Datta, A. Chatterjee, S.K. Mukhopadhyay, On the environmental impact of waste chromium of tannery agglomerates in the East Calcutta wetland ecosystem, J. Soc. Leather Technol. Chem., 84 (2000) 94–100.
- [11] J.C. Igwe, A.A. Abia, Mazie cob and husk as adsorbents for the removal of Cd, Pb and Zn ions from waste water. Phys. Sci., 2 (2003) 83–94.
- [12] N.K. Uberoi, Environmental Management, Vol. 269, Excel Books Publisher, New Delhi, 2003.
- [13] M. Horsfall, A.I. Spiff, Studies on effects of temperature on the sorption of Pb²⁺ and Cd²⁺ from aqueous solution by *Caladium bicolor* (wild cocoyam) biomass, Electron. J. Biotechnol., 8 (2005) 1–8.

- [14] K. Kolomaznik, M. Adamek, I. Anđel, M. Uhlířova, Leather waste potential threat to human health, and a new technology of its treatment, *J. Hazard. Mater.*, 160 (2008) 514–520.
- [15] W.G. Characklis, K.C. Marshall, *Biofilms*, Wiley, New York, 1990.
- [16] J.W.T. Wimpenn, An Overview of Biofilms as Functional Communities, A.D.G., Gilbert, H.M. Lappin-Scott, M. Wilson, Eds., *Community Structure and Cooperation in Biofilms*, Cambridge University Press, Cambridge, 2000, pp. 1–24.
- [17] L. Hall-Stoodley, J.W. Costerton, P. Stoodley, Bacterial biofilms: from the natural environment to infectious diseases, *Nature Rev. Microbiol.*, 2 (2004) 95–108.
- [18] Z. Lewandowski, P. Stoodley, S. Altobelli, E. Fukushima, Hydrodynamics and kinetics in biofilm systems—recent advances and new problems, *Water Sci. Technol.*, 29 (1994) 223–229.
- [19] H.C. Flemming, *Biofilms and environmental protection*, *Water Sci. Technol.*, 27 (1993) 1–10.
- [20] T.R. Neu, K.C. Marshall, Bacterial polymers physicochemical aspects of their interactions at interfaces, *J. Biomater. Appl.*, 5 (1990) 107–133.
- [21] J.W. Costerton, Z. Lewandowski, D. De Beer, D. Caldwell, D. Korber, G. James, *Biofilms, the customized microniche*, *J. Bacteriol.*, 176 (1994), 2137–2142.
- [22] M. Schorer, M. Eisele, Accumulation of inorganic and organic pollutants by biofilms in the aquatic environment, *Water Air Soil Pollut.*, 99 (1997) 651–659.
- [23] L.R. Hoffman, D.A. D'Argenio, M.J. MacCoss, Z. Zhang, R.A. Jones, S.I. Miller, Aminoglycoside antibiotics induce bacterial biofilm formation, *Nature*, 436 (2005) 1171–1175.
- [24] E. Karatan, P. Watnick, Signals, regulatory networks, and materials that build and break bacterial biofilms, *Microbiol. Mol. Biol. Rev.*, 73 (2009) 310–347.
- [25] P. Watnick, R. Kolter, Biofilm: city of microbes, *J. Bacteriol.*, 182 (2000) 2675–2679.
- [26] A. França, V. Carvalhais, M. Vilanova, G.B. Pier, N. Cerca, Characterization of an in vitro fed-batch model to obtain cells released from *S. epidermidis* biofilms, *AMB Express*, 6 (2016) 23. doi:10.1186/s13568-016-0197-9.
- [27] S. Golby, H. Ceri, L.L.R. Marques, R.J. Turner, Mixed-species biofilms cultured from an oil sand tailings pond can biomineralize metals, *Microb. Ecol.*, 68 (2014) 70–80.
- [28] I. Ylla, C. Canhoto, A.M. Roman, Effects of warming on stream biofilm organic matter use capabilities, *Microb. Ecol.*, 68 (2014) 132–145.
- [29] D.M. Gonzales, E. Moreno, Q.J. Samiento, R.A. Cormenzana, Studies activity of wastewater from olive oil mills (Alepechin) inhibitory activity of phenolic and fatty acids, *Chemosphere*, 20 (1990) 423–432.
- [30] A.G. Stams, E.S.J. Oude, Understanding and advancing wastewater treatment, *Curr. Opin. Biotechnol.*, 8 (1997) 328–334.
- [31] Z. Song, C.J. Williams, R.G.J. Edyvea, Tannery wastewater treatment using an upflow anaerobic fixed biofilm reactor (UAFBR), *Environ. Eng. Sci.*, 20 (2003) 587–599.
- [32] E. El-Bestawy, M.H. El-Masry, A.H. El-Koweidy, Removing Chromium from Tannery Wastewater Effluent Using Bench Sand Biofilm Filter, International Conference on Environmental Management, Health and Sustainable Development, 22–25 March, Alexandria, 1999.
- [33] M.H. El-Masry, E. El-Bestawy, N.I. El-Adl, Bioremediation of vegetable oil and grease from polluted wastewater using a sand biofilm system, *J. Microbiol. Biotechnol.*, 20 (2004) 551–557.
- [34] K. Sundar, I. Mohammed Sadiq, A. Mukherjee, N. Chandrasekaran, Bioremoval of Trivalent Chromium Using *Bacillus* Biofilms Through Continuous Flow Reactor, Centre for Nano biotechnology, Nano Bio-Medicine Laboratory School of Bio Sciences and Technology, VIT University, Vellore, India, 2011.
- [35] V.N. Kavamura, E. Esposito, Biotechnological strategies applied to the decontamination of soils polluted with heavy metals, *Biotechnol. Adv.*, 28 (2010) 61–69.
- [36] H. Ceri, M.E. Olson, R.J. Turner, Needed, new paradigms in antibiotic development, *Expert Opin. Pharmacother.*, 11 (2010) 1233–1237.
- [37] O. Ciofu, T. Tolker-Nielsen, Antibiotic Tolerance and Resistance in Biofilms, T. Bjarnsholt, C. Moser, P. Jensen, N. Høiby, Eds., *Biofilm Infections*, Springer Publishing Company, New York, 2010, pp. 215–229.
- [38] J. Harrison, H. Ceri, J. Yerly, C.A. Stremick, Y. Hu, R. Martinuzzi, R.J. Turner, The use of microscopy and three-dimensional visualization to evaluate the structure of microbial biofilms cultivated in the Calgary biofilm device, *Biomed. Life Sci.*, 8 (2006) 194–215.
- [39] J.J. Harrison, H. Ceri, R.J. Turner, Multimetal resistance and tolerance in microbial biofilms, *Nature Rev. Microbiol.*, 5 (2007) 928–938.
- [40] E. El Bestawy, F. Al-Fassi, R. Aburokba, Biological treatment of leather-tanning industrial wastewater using free living bacteria, *Adv. Life Sci. Technol.*, 12 (2013) 46–65.
- [41] L.S. Clesceri, C.G. Greenberg, A.D. Eaton, Standard Method for the Examination of Water and Wastewater, 20th edn, American Public Health Association (APHA), USA, 1999.
- [42] A. Malik, Metal bioremediation through growing cells, *Environ. Int.*, 30 (2004) 261–278.
- [43] H. Eccles, Treatment of metal-contaminated wastes: why select a biological process, *Trends Biotechnol.*, 17 (1999) 462–465.
- [44] C.N. Mulligan, R.N. Yong, B.F. Gibbs, Remediation technologies for metal-contaminated soils and groundwater: an evaluation. *Eng. Geol.*, 60 (2001) 193–207.
- [45] P.V. McCormick, J. Cairns, Algae as indicators of environmental change, *J. Appl. Phycol.*, 6 (1994) 509–526.
- [46] S. Fuchs, T. Haritopoulou, M. Willhelmi, Biofilms in freshwater ecosystems and their use as a pollutant monitor, *Water Sci. Technol.*, 34 (1996) 137–140.
- [47] C. Bonaventura, F.M. Johnson, Health environments for health people: bioremediation today and tomorrow, *Environ. Health Perspect.*, 105 (1997) 5–20.
- [48] M.C. Cammarota, G.L.S. Annajr, Metabolic blocking of exopolysaccharides synthesis effects on microbial adhesion and biofilm accumulation, *Biotechnol. Lett.*, 20 (1998) 1–4.
- [49] C.J. Gantzer, A.B. Cunningham, W. Gujer, B. Gutekunst, J.J. Heijnen, E.N. Lightfoot, G. Odham, E. Rosenberg, S.A.J.B. Zehander, Exchange Process at the Fluid Biofilm Interface, W.G. Characklis, W.P.A. Berlin, Eds., Report of Dahlem Workshop on Structure and Function of Biofilms, November 27–December 2, New York, John Wiley and Sons, 1989.
- [50] W.M. Fogarty, C.T. Kelly, In: W.M. Fogarty, C.T. Kelly, Eds., *Microbial Enzymes and Biotechnology*, Vol. 2, Elsevier Applied Science, London, 1990, pp. 71–133.
- [51] F. Gebara, Activated sludge biofilm wastewater treatment system, *Water Res.*, 33 (1999) 230–238.
- [52] C. Gold, A. Feurtet-Mazel, M. Coste, A. Boudou, Effects of cadmium stress on periphytic diatom communities in indoor artificial streams, *Freshwater Biol.*, 48 (2003) 316–328.
- [53] M. Mages, M. Óvári, W. Tümpling, K. Kröpfel, Biofilms as bio-indicator for polluted waters? Total reflection X-ray fluorescence analysis of biofilms of the Tisza river (Hungary), *Anal. Bioanal. Chem.*, 378 (2004) 1095–1101.
- [54] K. Kröpfel, P. Vladár, K. Szabó, E. Ács, A.K. Borsódi, S. Szikora, Chemical and biological characterization of biofilms formed on different substrata in Tisza river (Hungary), *Environ. Pollut.*, 144 (2006) 626–631.
- [55] Y. Liu, J.H. Tay, The essential role of hydrodynamic shear force in the formation of biofilm and granular sludge, *Water Res.*, 36 (2002) 1653–1665.
- [56] R. Vilchez, C. Pozo, M.A. Gomez, B. Rodelas, J. Gonzalez-Lopez, Dominance of sphingomonads in a copper-exposed biofilm community for groundwater treatment, *Microbiology*, 153 (2007) 325–337.

Supplementary material



Fig. S1. White stone, the supporting material and the biofilm system formation.



Fig. S2. The biofilm and control systems during operation.