# Copper(II) adsorption by Opuntia ficus-indica biochar fiber–MnO<sub>2</sub> composites

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Received 18 October 2018; Accepted 28 December 2018

### ABSTRACT

The adsorption efficiency of biochar fibers obtained from *Opuntia ficus-indica* cactus regarding the removal of Cu(II) ions from aqueous solutions has been investigated prior and after MnO<sub>2</sub> surface-deposition. The removal efficiency has been studied as a function of pH, Cu(II) concentration, ionic strength, temperature and contact time. In addition, the biochar fibers have been characterized by Fourier transform infrared spectroscopy studies to elucidate the adsorption mechanism. Evaluation of the experimental data indicated that the MnO<sub>2</sub>-composite product presents higher adsorption capacity, which is attributed to the formation of inner-sphere surface complexes, and that the dsorption reaction is a relatively fast and endothermic process. The results of the present study reveal that MnO<sub>2</sub>-biochar composites could be used as very effective adsorbents for the removal of bivalent metal-ions from contaminated waters.

Keywords: Adsorption; Biochar fibers; MnO2-biochar composites; Copper(II); Waters

### 1. Introduction

Removal of toxic metal ions from large volumes of industrial process solutions and (waste)waters is of particular interest with respect to environmental protection and sustainable development. However, successful water treatment methods require a cost-effective remediation technology. Conventional technologies relying on mineral adsorbents or chemical flocculating agents are relatively expensive and hence recently biosorption is presented as an alternative method for the removal of toxic metals from industrial process solutions and wastewaters [1]. Biosorption technologies based on living or dead biomass for accumulation and removal of heavy metals from aqueous solutions often replaces conventional processes in (waste) water treatment [1].

However, regarding living biomass (e.g., microbial systems) the major drawback is the price of growing a sufficient quantity of bacterial or algae biomass. Hence, the efforts have been focused on plant tissues and a variety of biomasses that represent by-products of various commercial activities (e.g., agriculture, food industry) have been investigated. The respective studies indicate that biomass by-products could represent good candidates for the development of inexpensive biosorption-based processes. Biomasses and particularly plant tissues are very abundant, non-hazardous materials, low cost, and can be easily disposed by incineration [1–6].

Adsorption of toxic metal ions on minerals [7,8] and biomasses [1–5] is extensively investigated. Moreover, there are several studies dealing with metal adsorption on non-treated and chemically modified biomass by-products [1–5,9–15]. Those modifications include surface oxidation, derivatization as well as coating with metal oxides to enhance selectivity toward specific metal ions [5,16]. There are also studies dealing with metal ion adsorption by carbon-based composite materials of interesting adsorbent properties such as selectivity, increased adsorption capacity, magnetism, etc. [16,17].

The present study deals with the adsorption of Cu(II) by biochar fibers derived from *Opuntia ficus-indica* prior and after MnO<sub>2</sub> deposition. MnO<sub>2</sub> deposition on the biochar

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Presented at the 6th International Conference on Sustainable Solid Waste Management (NAXOS 2018), 13-16 June 2018, Naxos Island, Greece. 1944-3994/1944-3986 © 2019 Desalination Publications. All rights reserved.

fibers is expected to enhance their separation efficiency and selectivity toward bivalent metal ions [6,18] and Cu(II) has been used as an analogue for bivalent toxic metal ions such as Cd(II) and Pb(II). The present study aims to investigate the effect of various parameters (e.g., pH, copper concentration, ionic strength, temperature and contact time) on the biosorption performance and determine thermodynamic parameters (e.g.,  $K_{ar} \Delta G^{\circ}$ ,  $\Delta H^{\circ}$  and  $\Delta S^{\circ}$ ), which are of fundamental importance for both the assessment of the chemical behavior of bivalent metal ions in aquatic environments and the development of water treatment technologies. Moreover, Fourier transform infrared spectroscopy (FTIR) studies have been performed to elucidate the mechanism upon which the adsorption is based.

### 2. Materials and methods

All experiments were performed at room temperature  $(23^{\circ}C \pm 2^{\circ}C)$  under ambient conditions in aqueous solutions and constant ionic strength. Generally, the experiments were performed in duplicate and the mean values have been used for data evaluation. The preparation of the metal ion solutions was carried out using a standard solution, and the Cu(II) concentration in solution was determined by potentiometry using a copper ion-selective electrode, which was calibrated before and after each experiment. pH measurements were performed by a commercial glass electrode, which was calibrated using a series of buffer solutions (pH 2, 4, 7 and 10). The adsorbents used in this study were biochar fibers prepared from *Opuntia ficus-indica* cactus fibres as described elsewhere [2,4,6].

### 2.1. Biochar fibers

Dried cactus cladodes were collected from local plants and the fibers of the cladodes were removed, washed thoroughly by distilled water and dehydrated at 110°C for 15 h. Following the fibers were carbonized under N<sub>2</sub> atmosphere and oxidized by means of 8 M HNO<sub>32</sub> as described elsewhere [9-15]. The product was sieved and the particle fraction between 200 and 500  $\mu m$  was selected for the adsorption experiments. The preparation of the MnO<sub>2</sub>-biochar fiber composites was carried out using KMnO<sub>4</sub> solutions as described elsewhere [2,4,6,19]. The determination of the morphology of the fibers was carried out by scanning electron microscopy (SEM, Vega TS5136LS-Tescan), specific surface area measurements were based on the Brunauer-Emmett-Teller theory (BET measurements) by means of N<sub>2</sub> adsorption (ASAP 2000, Micromeritics, USA) and FTIR spectroscopic measurements (FTIR spectrometer 8900, Shimadzu, Japan) were performed by means of prepared translucent KBr disks including finely ground biochar fibers encapsulated at a 10:1 mass ratio, prior and after Cu(II) adsorption.

#### 2.2. Adsorption measurements

In order to investigate the effect of various parameters (pH, copper concentration, ionic strength, temperature and contact time) on the Cu(II) adsorption on oxidized biochar fibers (OCF) and MnO<sub>2</sub>-biochar fiber composites (OCF-MnO<sub>2</sub>), different classes of experiments were conducted.

In these experiments, the parameter under investigation was varied, while other experimental parameters were kept constant. All test solutions were prepared using 0.1 M NaClO<sub>4</sub> (Sigma-Aldrich Co., Germany) as the background electrolyte. The effect of pH was studied in an adsorption system (0.1 g adsorbent and 15 mL of the test solution  $[Cu(II)] = 1 \times 10^{-4} M$ in which pH was varied between 1.5 and 8.5 by addition of HClO<sub>4</sub> or NaOH. The effect of other parameters was studied at pH 3 and 5 for OCF and OCF-MnO<sub>2</sub>, respectively, which were the self-adjusted pH values for the two systems. The effect of the ionic strength (salinity) was studied by addition of NaClO<sub>4</sub> salt to prepare solutions of various concentrations (0.001, 0.01, 0.1, 0.5, 0.7 and 1 M) at constant adsorbent amount (0.1 g), total copper concentration (1  $\times$  10<sup>-4</sup> M). For studying the effect of initial copper concentration, the latter was varied between  $1 \times 10^{-5}$  M and  $9 \times 10^{-3}$  M, at a prefixed amount of adsorbent (adsorbent dosage = 0.1 g and the optimum pH value for each adsorbent). The effect of temperature was studied between 10°C and 70°C at the same conditions that are described above. For kinetic studies, a certain amount of the biomasses (0.1 g) was mixed with 15 mL of copper solution (1  $\times$  10<sup>-4</sup> M). After an equilibration time of 24 h, the metal concentration was determined by potentiometry using a copper ion selective electrode. For each test solution, a corresponding reference solution was prepared and was similar to the test solution except that it did not contain the adsorbent material. Experiments and data evaluation have been performed as described elsewhere [6,8,14,15].

### 3. Results and discussion

### 3.1. Fiber characterization

In order to prove the formation of the biochar fiber– $MnO_2$  composites (OCF– $MnO_2$ ), the products were examined by X-ray powder diffraction (XRD) and a representative XRD diffractogram is shown in Fig. 1. The diffraction peaks at  $2\theta = 25.4$ , 36.5, 41.5 and 65.9 from the biochar fiber– $MnO_2$  composites match the standard XRD pattern of the manganese oxide (JCPDS 80-1098), indicating the successful deposition of  $MnO_2$  on the biochar surface [6,20]. The broad peak at  $2\theta = 24^\circ$  is characteristic for graphite-type materials [9,13].





In addition, Fig. 2 presents a SEM image of biochars fibers prior and after  $MnO_2$  deposition, which clearly shows distinct  $MnO_2$  crystallites formed on the biochar surface (white spots) and that chemical modification did not affect the texture of the fibers. Even after chemical modification the fibers keep their laminated texture, which is responsible for their increased external surface. However, BET measurements indicate that there is no internal mesoporous surface (BET surface <5 m<sup>2</sup> g<sup>-1</sup>), which is expected for plant fibers carbonized at temperatures below 650°C.

### 3.2. pH effect

The relative adsorption is related to the chemical affinity of the surface for the adsorbate, which depends on both the chemical behavior of Cu(II) ions in solution and the surface charge of the adsorbent. Hence, the solution pH is an important parameter affecting adsorption on surfaces, because pH may govern both the chemical behavior of a metal ion in solution and the surface charge of an adsorbent. To study the effect of pH on the Cu(II) adsorption, samples of the two different biochar products (OCF and OCF–MnO<sub>2</sub>) were conducted with Cu(II) solutions at different pH values (1.5 < pH < 8.5).

The effect of pH on the relative adsorption of Cu(II) by the two different biochar materials (OCF and OCF–MnO<sub>2</sub>) is shown in Fig. 3. In the case of the non-treated cactus fibers, the relative adsorption increases with increasing pH and reaches a maximum value (~100%) for pH > 5. For pH values below 5, the relative adsorption decreases with pH due to the protonation of the surface carboxylic and/or hydroxy groups which are present at the surface of the oxidized biochar fibers and their MnO<sub>2</sub> composites, respectively, and bind the Cu(II) ions as described schematically for instance by Eqs. (1) and (2).

$$=S-COOH + Cu^{2+} \leftrightarrow = S-COO-Cu^{+} + H^{+}$$
(1)

$$=S-OH + C11^{2+} \leftrightarrow =S-O-C11^{+} + H^{+}$$

#### 3.3. Effect of Cu(II) concentration

In order to evaluate the maximum adsorption capacity  $(q_{\max})$ , adsorption experiments with varying copper concentrations have been performed at the optimum pH for each adsorbent. The corresponding isotherms, which have been well fitted Langmuir isotherm model, are graphically shown in Fig. 4(a) and indicate that the biochar fiber–MnO<sub>2</sub> composites ( $q_{\text{max}} = 5.0 \pm 0.2 \text{ mol kg}^{-1}$ ,  $R^2 = 0.98$ ) present higher adsorption capacity for Cu(II) than the non-modified biochar fibers ( $q_{max}$  = 2.1 ± 0.1 mol kg<sup>-1</sup>,  $R^2$  = 0.99). The  $q_{max}$  value of the composite material shows that chemical modification of the material improves its adsorption attributes, and that biochar fiber-MnO2 composites can be used as an alternative for MnO<sub>2</sub> resins [19]. Moreover, the  $q_{\rm max}$  value determined for the biochar fiber-MnO, composite is significantly higher than corresponding values ( $q_{max} < 3 \mod kg^{-1}$ ) reported in literature for the adsorption of Cu(II) by MnO2-biochar composites [21-25]. This increased adsorption capacity of the material studied



Fig. 3. Effect of pH on the relative adsorption of Cu(II) by biochar fibers prior and after MnO, surface deposition.



Fig. 2. SEM image of Opuntia ficus-indica biochar fibres (a) prior and (b) after MnO<sub>2</sub> deposition.



Fig. 4. (a) Adsorption isotherms and (b) the effect of the adsorbent mass on the relative adsorption corresponding to Cu(II) adsorption by biochar fibers prior and after MnO<sub>2</sub> surface deposition.

here is attributed to both, the nano-sized MnO<sub>2</sub>-particles and high external surface of the biochar fibers [13,14]. In addition, investigations on the effect of the adsorbent mass on the relative adsorption (Fig. 4(b)) clearly show that at similar adsorbent amounts the modified material presents significantly higher adsorption efficiency.

### 3.4. Effect of ionic strength/salinity on the adsorption efficiency

Furthermore, the effect of ionic strength 0.001, 0.01, 0.1, 0.5, 0.7 and 1 M (NaClO<sub>4</sub>) on the relative adsorption of Cu(II) was investigated, in order to understand the interaction mechanisms on which Cu(II) binding on the two different types of biochar materials is based. The experimental data obtained from the corresponding experiments are graphically summarized in Fig. 5 and show clearly that the relative adsorption of the oxidized (OCF) and the biochar fiber–MnO<sub>2</sub> composites (OCF–MnO<sub>2</sub>) is not affected by increasing salinity indicating specific interactions between Cu(II) and the active surface groups (carboxylic and hydroxy groups) of the respective adsorbents. Hence, the adsorption can be attributed to the formation of innersphere complexes, which are only little affected by salinity changes [6,26].

#### 3.5. FTIR studies

Moreover, FTIR spectra were recorded to identify the functional groups and elucidate the adsorption mechanism corresponding to the Cu(II) adsorption by the biochar adsorbents prior and after modification. The FTIR spectra of OCF and the corresponding MnO<sub>2</sub> composites (OCF–MnO<sub>2</sub>) before and after Cu(II) adsorption are shown in Fig. 6. The IR spectra of the OCF show main absorption bands at 3,433; 1,717; 1,610; 1,342 and 1,250 cm<sup>-1</sup>. The broad band around 3,433 cm<sup>-1</sup> corresponds and the other bands to stretching and bending modes of the carboxylic moieties present at the biochar surface. After Cu(II) adsorption and depending on the Cu(II) concentration, the relative intensity of the bands attributed to the carboxylic moieties changes dramatically



Fig. 5. Effect of ionic strength on the relative adsorption of Cu(II) by biochar fibers prior and after MnO<sub>2</sub> surface deposition.



Fig. 6. FTIR spectra of the oxidized biochar fibers (OCF) prior and after successive Cu(II) adsorption.



Fig. 7. FTIR spectra of the oxidized biochar fiber– $MnO_2$  composites (OCF– $MnO_2$ ) prior and after successive Cu(II) adsorption.

indicating the formation of inner-sphere complexes between Cu(II) and the surface active species [14] and are in agreement with the results obtained from the experiments related to the effect of salinity.

Regarding the FTIR spectra of the biochar-MnO<sub>2</sub> composites (Fig. 7), the broad bands at ~3,369 cm<sup>-1</sup> are attributed to -OH groups, and the peaks at 2,927 and 2,856 cm<sup>-1</sup> correspond to CH, deformation vibrations. The bands in the region between 1,730 and 1,450 cm<sup>-1</sup> are ascribed to aliphatic or aromatic groups (e.g., C=C), and carbonyl C=O stretching vibrations. The peak at 1,360 cm<sup>-1</sup> is associated with -OH vibrations and the broad band observed after MnO<sub>2</sub> deposition in the low-frequency region around 510 cm<sup>-1</sup> corresponds to Mn-O vibrations. After Cu(II) adsorption, the relative intensity of the characteristic bands at 1,580; 1,360 and 511 cm<sup>-1</sup> changes significantly indicating that mainly the hydroxyl groups of the MnO<sub>2</sub> phase bind the Cu(II) ions to form stable inner-sphere complexes (e.g., Mn-O-Cu) [27] and are in agreement with the results obtained from the experiments related to the effect of salinity. The broad peak at 510 cm<sup>-1</sup> is a characteristic absorption band of birnessite, corresponding to Mn-O stretching modes of the octahedral layers in the birnessite structure which is consistent with the previous XRD result.

#### 3.6. Effect of temperature on the adsorption efficiency

The effect of temperature on Cu(II) adsorption biochar fibers prior and after MnO<sub>2</sub> surface deposition was investigated to estimate the corresponding thermodynamic data based on the van't Hoff equation. Evaluation of the data (Fig. 8) shows that adsorption of Cu(II) on the biochar composites (OCF–MnO<sub>2</sub>) is an endothermic and entropy-driven spontaneous process ( $\Delta H^{\circ}$  = 15.6 kJ mol<sup>-1</sup>,  $\Delta S^{\circ}$  = 156.7 J mol<sup>-1</sup> K<sup>-1</sup>, *R* = 0.90). On the other hand, adsorption of Cu(II) on the OCF is an exothermic process ( $\Delta H^{\circ}$  = -22.3 kJ mol<sup>-1</sup>,  $\Delta S^{\circ}$  = 2.0 J mol<sup>-1</sup> K<sup>-1</sup>, *R* = 0.95). Generally, the values of the thermodynamic parameters evaluated are close to corresponding values reported in the literature for similar systems [6,21–29].



Fig. 8. Effect of temperature on the relative adsorption of Cu(II) by biochar fibers prior and after MnO<sub>2</sub> surface deposition.



Fig. 9. Time-depended relative adsorption of Cu(II) by biochar fibers prior and after MnO<sub>2</sub> surface deposition.

#### 3.7. Kinetic measurements

According to Fig. 9, which shows the relative amount of adsorbed Cu(II) as a function of time, the adsorption of Cu(II) by the two different biochar materials is a two-step process with a relatively first fast step which could not be followed by the potentiometric method and a second slow process related to the diffusion of the Cu(II) ions within the fiber channels/ tubules [21–25].

### 4. Conclusions

In this study, oxidized and MnO<sub>2</sub>-modified biochar fibers prepared from *O. ficus-indica* cactus fibers were used to adsorb Cu(II) ions from aqueous solutions. SEM and XRD measurements indicate the successful preparation of the OCF–MnO<sub>2</sub> composites and the experimental data indicate increased affinity of both materials for Cu(II), which is further improved by chemical modification and the formation of  $MnO_2$ -composites. The equilibrium data were described well by a Langmuir isotherm, indicating that the adsorption of copper on biomass is based on specific chemical interactions, which is corroborated by experiments of varying ionic strength/salinity and FTIR studies prior and after Cu(II) adsorption. The thermodynamics change dramatically upon  $MnO_2$  deposition indicating different adsorption mechanism and adsorption kinetics are relatively fast for both materials.

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