



Effect of forward osmosis membrane process on biogas slurry concentration

Ping Xiao, Liang Liu*

Green Intelligence Environmental School, Yangtze Normal University, No. 16 Juxian Road, Fuling District, Chongqing 408100, China, Tel.: +86 18875436711; email: liuliang198104@163.com (L. Liu), Tel.: +86 18983242003; email: xiaoping19831020@126.com (P. Xiao)

Received 5 September 2023; Accepted 20 November 2023

ABSTRACT

The use of forward osmosis (FO) for treating biogas slurry was systematically investigated in this study. Membrane performance was investigated with draw solution (DS) concentrations, membrane orientation, solution flow rate and temperature to identify the optimal conditions. All of the parameters including DS concentration, membrane orientation, solution flow rate and temperature had significant influence on membrane performance. Results showed that the optimal operation conditions were DS concentration 2 M, flow rate 1.5 L/min, solution temperature 20°C and FO mode. The recovery rates of total dissolved solids and other substances can reach over 95.6%, no matter how many times concentrated the biogas slurry was. The membrane flux decreased with the increased concentration multiple of biogas slurry. The highest membrane concentration efficiency was about 3 L²/h·m²·g, when biogas slurry was concentrated between 1.5 and 2 times. FO technology which can improve the application value of biogas slurry as fertilizer effectively was feasible to biogas slurry concentration.

Keywords: Forward osmosis; Biogas slurry; Recovery rate; Concentration

1. Introduction

Biogas slurry is the residue of organic substances such as crop straw, the feces of human or livestock after anaerobic fermentation. It is a organic compound fertilizer, which not only contains the tremendously high amount of organic matters but also contains nutrient such as nitrogen (N), phosphorus (P) and potassium (K). Previous research has shown that return the biogas slurry into farmland plays an important role in increasing crop yield on, improving agricultural product and soil quality [1]. Annually, China's biogas plants produce about 400 million tons of biogas slurry. If biogas slurry cannot be returned to the field or be consumed, it may cause resources wasting. In addition, discharge of untreated biogas slurry poses serious threats on the environment. There are serious doubts such as low nutrient content and high transportation costs with direct application of biogas slurry as liquid fertilizer [2,3]. Therefore, it is significant to concentrate biogas slurry and obtain high-quality liquid

fertilizer with easy transport. The main technologies for biogas slurry concentration include humidification–dehumidification, electro dialysis and membrane treatment [4–6].

Membrane filtration technology has been applied in wastewater treatment as a promising new technology with great development prospects [7–9]. It can be divided by membrane pore size and mechanism into microfiltration, ultrafiltration, nanofiltration, reverse osmosis (RO) and forward osmosis (FO). It is more economical and occupies a smaller area than traditional technology [10]. More importantly, solution can be concentrated, and substances can be rejected by membrane. A pilot dual stage RO membrane process was designed to concentrate, and to recover the nutrient in the biogas slurry [11]. The recovery rates of nutrient and water were 98% and 92.52%, respectively. And the rejection rates of chemical oxygen demand (COD) and ammonia nitrogen (NH₃-N) were over 99%. A hybrid membrane technology was used for biogas slurry concentration. The RO membrane can concentrate the biogas

* Corresponding author.

slurry with the concentration factor of 5, showed over 97% removal for COD and $\text{NH}_3\text{-N}$, which proved the feasibility of the integrated membrane technology in application [12]. Additionally, membrane separation combined with other technologies such as catalytic ozonation can also improve the nutrition of biogas slurry and reduce its ecological risk [13]. However, due to the properties of biogas slurry, serious membrane fouling often appear during membrane process. Thus, the selection of membrane treatment technology with a low fouling tendency has important meanings in biogas slurry.

For the past few years, FO has generated the public's interest because of the need for more sustainable processes. It depends on a highly concentrated draw solution (DS) as a driving force to extract pure water from the feed solution (FS) based on the difference in osmotic pressure between the DS and FS [14]. Hence, FO has the advantages of low energy consumption, low fouling tendency, high fouling reversibility and high recovery rates compared to the pressure-driven membranes [15–17]. The rejections for most of metal ions under FO process are high [18–20]. FO using NaCl as DS combined with a membrane reactor could reject over 96% total phosphorus (TP), 98% COD and 76% ammonium [21]. Furthermore, FO which can effectively reduce solution volume and realize the reuse of waste is an alternative method for treating high concentration organic wastewater such as landfill leachate [22,23]. It has been used to enrich nutrients from sludge concentrate, and membrane performance has also investigated [24–26]. Most of phosphate in sludge concentrate was recovered as calcium phosphate precipitates in FO process [24].

In general, membrane filtration efficiency in FO process is impacted by various factors, including operating conditions, water conditions and membrane structure. The concentration effect of biogas slurry in FO membrane can be effectively improved by optimizing operating conditions and solution properties. The parameters including type of DS and its concentration, flow rate and solution temperature can affect the concentration of FO membrane, especially the solution temperature. Biogas slurry is a product of anaerobic fermentation. The temperature of anaerobic fermentation is usually divided into three categories: low temperature (less than 20°C), medium temperature (20°C–45°C), and high temperature (45°C–60°C) [27]. The temperature of biogas slurry is related to the type of anaerobic fermentation. Furthermore, solution temperature can influence the key characteristics in membrane separation processes such as solute mass transfer, water viscosity, water transportation, concentration polarization (CP) and membrane fouling [28]. To date, there has been relatively little research conducted on optimizing operating conditions during the FO membrane treatment of biogas slurry [29–31]. Therefore, it is very important to obtain optimized operating conditions for the concentration of biogas slurry by FO membrane treatment. Further research is needed.

The objective of this study is to evaluate the feasibility of FO for biogas slurry concentration. The effects of operating conditions including DS concentration, membrane orientation, water velocity, solution temperature and concentration multiple were investigated. Furthermore, membrane concentration efficiency of biogas slurry was also

studied. The recovery rates of organic matters and nutrients were analyzed. This study has potential implications on FO membrane in biogas slurry treatment.

2. Materials and methods

2.1. Membrane and solution

Cellulose triacetate (CTA) membranes (CTA-ES) for FO used in this study was obtained from the Hydration Technology Innovations (HTI, Albany, USA). The membrane at pH ranges (3–8) had a filtration area of 40 cm². The limit of membrane temperature were 71°C.

The raw biogas slurry came from the effluent of hog-gery wastewater after biogas fermentation. The characteristics are displayed in Table 1. The biogas slurry contained large quantities of N, P, K and organic matters. The DS was sodium chloride solution (NaCl, Beijing Chemical Works, China).

2.2. FO membrane set-up

A schematic diagram of the FO performance are shown in Fig. 1. It consisted of a membrane cell, a FS tank and a DS tank, two peristaltic pumps, a temperature control device, an electronic balance and a computer. Membrane cell had symmetric channels on both sides, allowed for both the FS and DS to flow tangential to the membrane. Two peristaltic pumps were used to recirculate the feed and draw liquids under different flow rate in a closed loop. The temperature control device was used to control the solution temperature. An electronic balance was placed under the DS tank, which connected to the computer to record the weight of the DS. Prior to each experiment, a virgin membrane was soaked in deionized water for 24 h in order to remove the protective solution. To obtain a satisfied filtrate flux, the FO membrane was initially stabilized for 24 h with deionized water as the FS and NaCl as the DS (baseline test). After that membrane experiments were performed. The experiment conditions are shown in Table 2, and the baseline test conditions were consistent with the membrane experiment. A virgin membrane was used in each experiment in order to compare the results under the same condition. All the tests had been repeated three times.

2.3. Analytical methods

Total organic carbon (TOC) was measured with a TOC analyzer (TOC-VCPH, Shimadzu, Japan). Total potassium was detected by the inductively coupled plasma-atomic emission spectrometry (ICP-AES, OPTIMA-2000, PerkinElmer, USA). The total dissolved solids (TDS), nitrogen and phosphorus were analyzed according to Chinese National Standards.

Membrane flux can be expressed using Eq. (1):

$$Q = \frac{V_1 - V_2}{T \times A} \quad (1)$$

where Q is the flux (L/h·m², LMH); V_1 and V_2 are the volume of biogas slurry before and after treatment (L); T is the filtration time (h); A is the membrane area (m²).

Table 1
Characteristics of raw biogas slurry

Parameters	pH	Conductivity ($\mu\text{S}/\text{cm}$)	TDS (g/L)	TK (mg/L)	TP (mg/L)	$\text{NH}_4\text{-N}$ (mg/L)	TN (mg/L)	TOC (mg/L)
Biogas slurry	7.63	$7,735 \pm 81$	4.98 ± 0.51	712 ± 31	253 ± 23	776 ± 18	$2,118 \pm 40$	$1,501 \pm 65$

TK - Total potassium; TN - Total nitrogen; TP - Total phosphorus

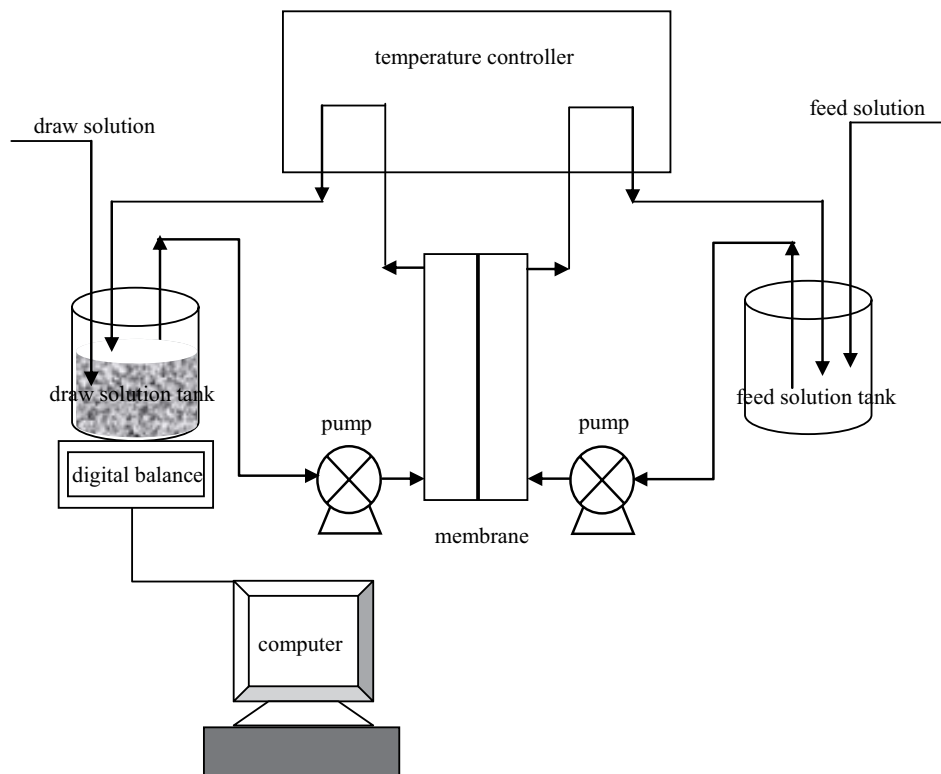


Fig. 1. Schematic diagram of the forward osmosis system.

Table 2
List of the experimental conditions

	Feed solution (FS)	Draw solution (DS)	Solution temperatures ($^{\circ}\text{C}$)	Flow rate (L/min)	Membrane orientation
1	Biogas slurry	1 M NaCl	20	1.5	FO
		2 M NaCl			
		3 M NaCl			
		4 M NaCl			
2	Biogas slurry	2 M NaCl	20	1.5	FO Pressure retarded osmosis (PRO)
3	Biogas slurry	2 M NaCl	20	0.5	FO
				1.0	
				1.5	
				2.0	
4	Biogas slurry	2 M NaCl	10	1.5	FO
			20		
			30		
			40		
			50		

Baseline test conditions were consistent with the membrane experiment.

Recovery rate of substance after membrane treatment is given by Eq. (2):

$$R = \frac{M_2 - V_2}{M_1 \times V_1} \times 100 \quad (2)$$

where R is the recovery rate (%); M_2 is the concentration of substance after treatment (mg/L); V_2 is the volume of biogas slurry after treatment (L); M_1 is the concentration of substance before treatment (mg/L); V_1 is the volume of biogas slurry before treatment (L).

The membrane concentration efficiency depended on the membrane flux when it concentrated per unit TDS. The concentration efficiency is calculated:

$$Z = \frac{Q'}{\Delta TDS} \quad (3)$$

where Z is membrane concentration efficiency ($L^2/h \cdot m^2 \cdot g$); Q' is flux at different stages ($L/h \cdot m^2$, LMH); ΔTDS is the variation of TDS during different stages (g/L).

3. Results and discussion

3.1. Influence of DS concentration

Previous research have been reported that NaCl with low-cost had a high osmotic pressure [32]. Therefore, it was an ideal DS. Membrane behavior under different NaCl concentrations was studied as shown in Fig. 2. High concentration DS possessed high osmotic pressure. Thus, the highest pure water flux was observed with 4 M NaCl in the FO mode. The DS was diluted, resulting in osmotic pressure reduced. The pure water flux gradually decreased with filtration time. In order to keep a constant osmotic pressure, the DS was replaced every hour. So, the pure water flux remained basically unchanged. More membrane flux loss was observed in the filtration of biogas slurry. At the end of the filtration, membrane fluxes with different concentrations of NaCl were 1.65, 2.78, 2.06 and 2.36 LMH, respectively. The reductions of flux were 73.39%, 65.98%, 81.30% and 80.93%, respectively. Biogas slurry were continuously concentrated, causing the osmotic pressure of FS increased. In addition, membrane fouling was another reason for flux decrease. Foulants can blocked membrane pores

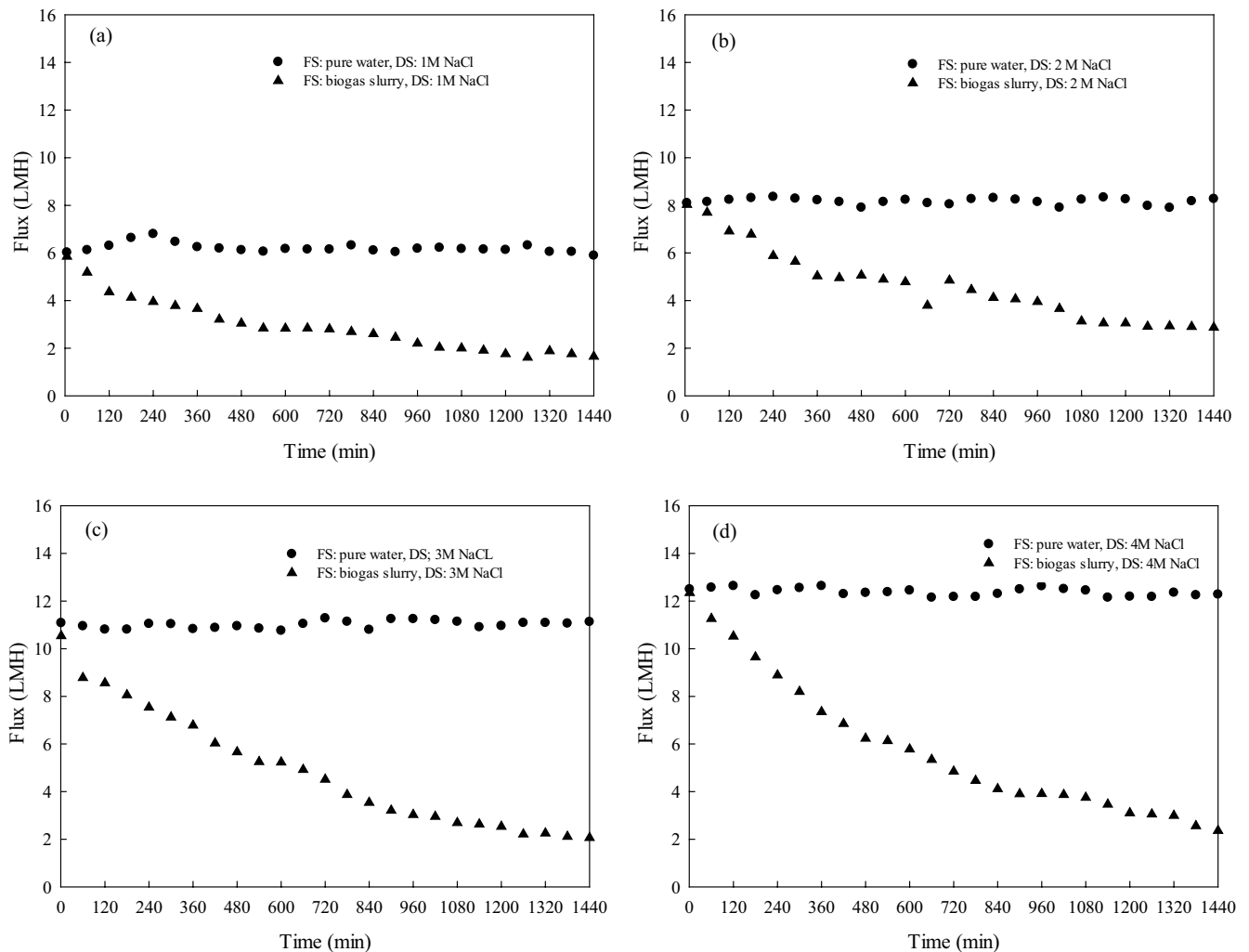


Fig. 2. Membrane performance at different concentrations of draw solution (FS: biogas slurry flow rate: 1.5 L/min, FO mode, 20°C).

and deposited on membrane surface, formed fouling layer, which increased the membrane resistance, leading the flux decrease. It had reported that high membrane flux can cause membrane flux decrease sharply, which meant a fast-fouling rate [33]. According to the experimental results, 2 M NaCl was selected as the appropriate concentration.

3.2. Characteristics of membrane orientation in flux decline

Membrane orientation had influence on membrane performance and fouling. Therefore, the filtration of biogas slurry in FO mode and pressure retarded osmosis (PRO) mode were evaluated, as displayed in Fig. 3. The pure water flux in PRO mode was higher than that of the FO mode. It was due to more severe dilutive internal concentration polarization (ICP) in the FO mode, compared to the concentrative ICP in the PRO mode [34,35]. Sharp decline of flux was observed in the PRO mode compared with that of the FO mode, while the biogas slurry was selected as FS. The flux was decreased by 75.4% in the PRO mode, higher than that of the FO mode with flux decreased by 64.89%. In addition, the slopes of the curves can be obtained through linear fitting. The slopes of the filtration of the biogas slurry were -3.28×10^{-3} and -4.93×10^{-3} in FO and PRO modes, respectively, showed severe fouling in PRO mode. According to previous research, three factors, the porous layer morphology (pore size, porosity and roughness), enhanced ICP and cake enhanced osmotic pressure (CEOP) due to pore blocking, contributed to rapid flux decline in the PRO mode [36].

3.3. Influence of flow rate

The influence of different flow rate on membrane flux has been investigated. The results are shown in Fig. 4. The flow rate had a little effect on membrane pure water flux. As the solution flow rate increased from 0.5 to 2.0 L/min, average flux of pure water was gradually increased from 6.23 to 9.65 LMH. Nevertheless, a noticeable influence in flux was observed in the filtration of biogas slurry, as Fig. 5 shows. membrane fluxes decreased from 6.17, 7.53, 8.15 and

9.35 LMH to 1.78, 1.93, 2.58 and 2.64 LMH, respectively. The decline rates were respectively 71.15%, 74.36%, 68.34% and 71.76%. The influence of external concentration polarization

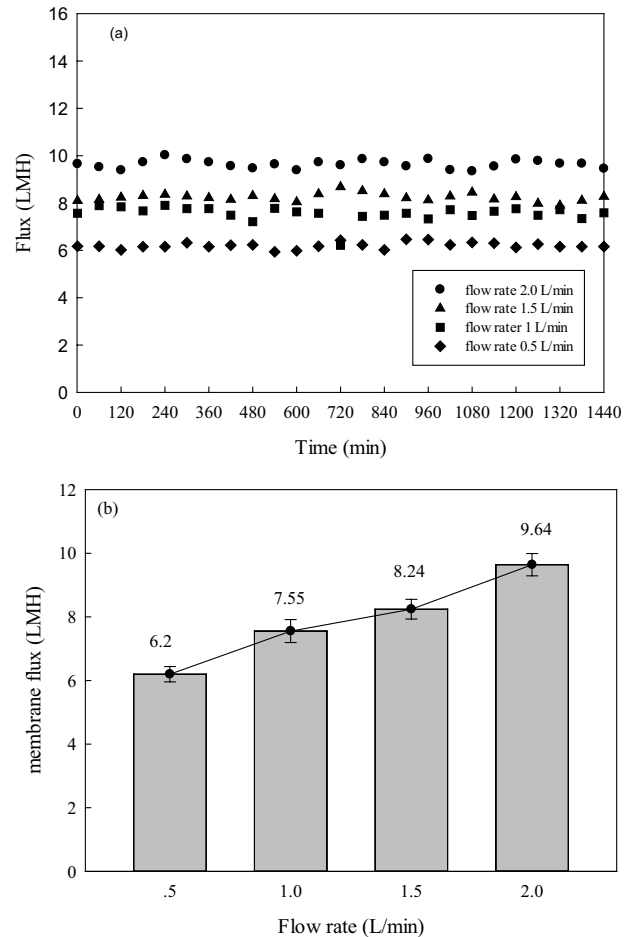


Fig. 4. Variation of pure water flux at different flow rate (FS: pure water, DS: 2 M NaCl, FO mode, 20°C).

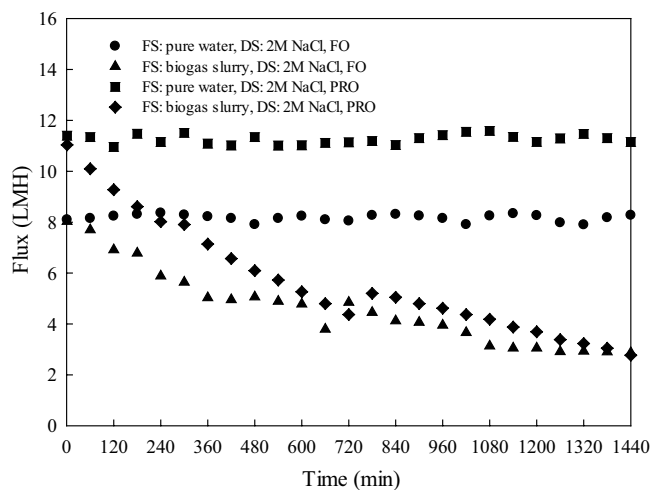


Fig. 3. Influence of membrane orientation (flow rate: 1.5 L/min, DS: 2 M NaCl, 20°C).

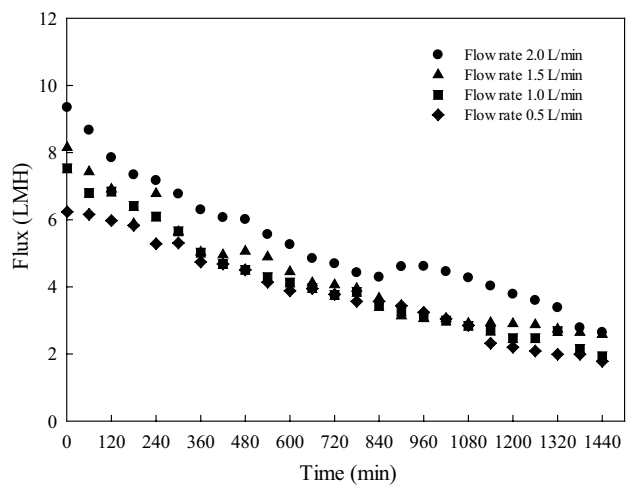


Fig. 5. Variation of flux at different flow rate (FS: biogas slurry, DS: 2 M NaCl, FO mode, 20°C).

(ECP) can be mitigated with the increasing of flow rate, causing the flux increased. Moreover, a high liquid flow rate can increase the shear force of the membrane surface, alleviate the foulants deposition effectively. However, an excessive flow rate will damage the membrane surface, shorten membrane life. And as mentioned earlier, high membrane flux may cause a high flux reduction. Therefore, 1.5 L/min was selected as the optimum flow velocity with considering various factors.

3.4. Membrane performance under different solution temperatures

Temperature influenced the thermodynamic characteristics of solution and the membrane properties, which directly influenced the water permeability, salt permeability and reverse solute flux selectivity [28,35]. The pure water flux was increased by 145.66%, when the solution temperature (both FS and DS) was increased from 10°C to 50°C, as shown in Fig. 6a. The net bulk osmotic pressure increased with the temperature. As both the FS and DS temperatures increased, pure water permeability coefficients

of the FO membrane increased due to an increase in solute diffusivity and a decrease in water viscosity [28,37]. However, there was contradiction on this topic. Different results had been reported. It suggested that heated only to FS or only to DS can enhance FO membrane process instead of the case when both solutions were heated. It had been reported that temperature of the FS was more important than DS temperature. Because a hot FS produced a warmer membrane, encouraging higher permeability and mitigated polarization [38]. And some researches had found that unilateral strengthening the DS temperature was more beneficial to improve FO performance [39,40]. Therefore, there was a positive influence on water flux with the temperature increasing. And further research would be conducted to understand the effect of temperature on biogas slurry concentration during FO membrane.

Membrane behavior influenced by temperature was also investigated during the biogas slurry filtration process, as displayed in Fig. 6b. The flux decreased with the filtration time increased. Furthermore, there was significant difference on the flux decline as the temperature increased from 10°C to 50°C. At the filtration time between 0 to 600 min, membrane fluxes decreased to 3.15, 5.06, 5.38, 6.74 and 9.19 LMH with the solution temperature increased from 10°C to 50°C, respectively (Fig. 6b). It showed a declined tendency. The decline rates of membrane flux were 49.35%, 39.87%, 51.31%, 48.07% and 40.17%. Subsequently, flux was dropped further. At the end of the filtration, fluxes were down by 62.08%, 69.34%, 76.11%, 77.04% and 79.56%, respectively, when the solution temperatures were 10°C, 20°C, 30°C, 40°C and 50°C. The fastest decline rate was 79.56% with the solution temperatures 50°C. The results was due to the joint influence of both organic convection and temperature polarization [41]. It suggested temperature had a significant effect on FO membrane performance.

3.5. Variation of flux under different concentration multiple of biogas slurry

There were obvious changes in membrane performance under different conditions. Under the condition of solution flow rate of 1.5 L/min, solution temperature of 20°C, FO mode, the variation of membrane flux in different concentration multiple of biogas slurry were investigated, as shown in Fig. 7. As the concentration multiple of biogas slurry increased, membrane flux decreased gradually. When the FS was raw biogas slurry, the maximum membrane flux was 8.4 LMH. When the biogas slurry was concentrated from 2.5 to 3 times, membrane flux decreased sharply from 6.76 to 3.58 LMH. When biogas slurry was concentrated 4 times, membrane flux was decreased to 2.15 LMH. membrane flux decrease rate was 74.4%, comparison to that of raw biogas slurry. As the increase of concentration multiple, the concentration of the substance in biogas slurry was increased. Due to a decrease of driving force on both sides of the membrane and the serious fouling, membrane flux decreases evidently. Previous research had shown a similar result [42].

The concentrations of the main substances were detected respectively and the recovery rate was calculated under different concentration multiple as shown in Table 3. The recovery rates of phosphorus, nitrogen and potassium

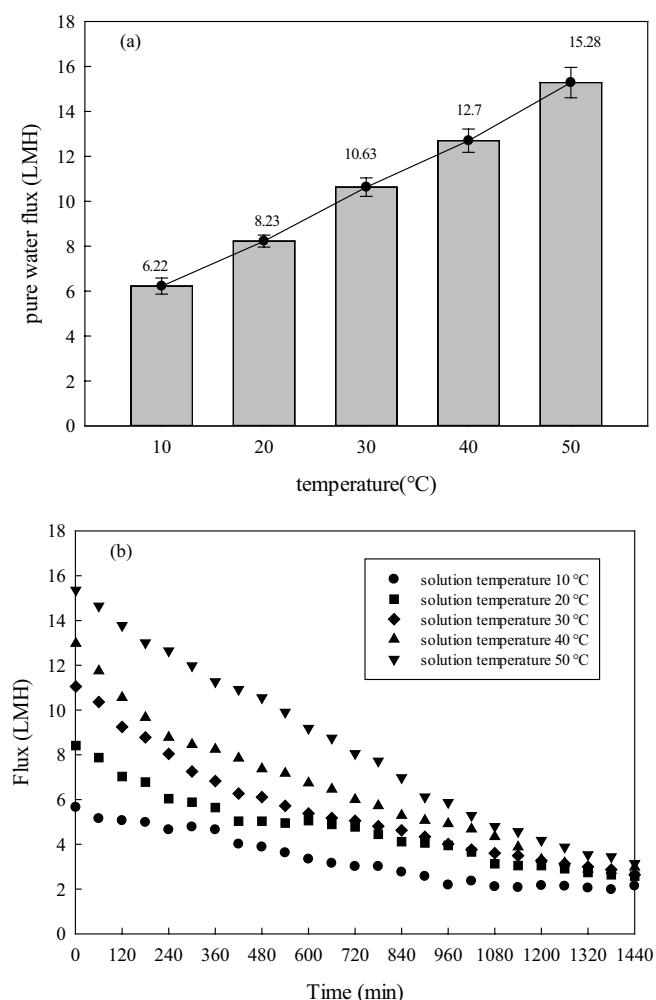


Fig. 6. Influence of solutions temperature on membrane performance, (a) average flux of pure water and (b) variation of biogas slurry (DS: 2 M NaCl, FO mode, flow rate 1.5 L/min).

were above 98%, no matter how many times concentrated the biogas slurry was. The recovery rate of TDS gradually decreased from 99.3% to 95.6% with the concentration multiple increased. However, it still can keep a high recovery rate over 95%. In addition, the recovery rates of TOC were also to keep above 98%. The results indicated that FO membrane can reject most of ions and organic matters. FO membrane with tiny pores had the advantage of high retention rate for solute. Solution–diffusion was the main mechanism for ions transport across the FO membrane. The Donnan equilibrium effect may hinder ionic permeation degrees of the feed ions across the active layer due to the presence of highly concentrated DS [43]. Furthermore, Metal ions with larger hydrated radius can be rejected for diffusivity decreases with increasing hydrated radius. Hence, it can be used as an effective treatment for biogas slurry concentration.

Due to the driving force of FO membrane originated from the concentration gradient of both sides of the membrane, the relationship between TDS concentration in biogas slurry and membrane flux was also analyzed. As Fig. 8a shown, membrane flux reduced gradually with the

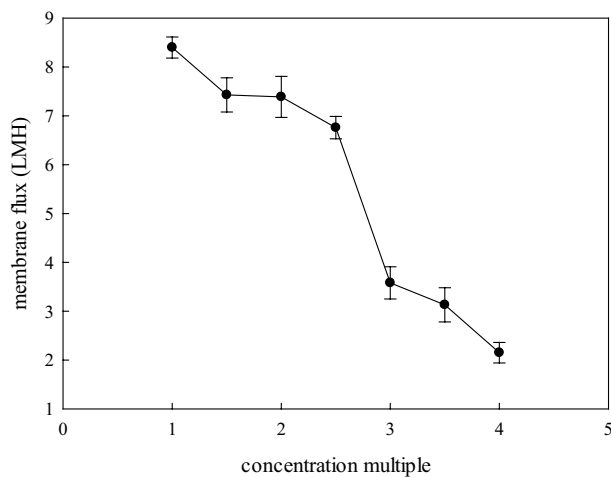


Fig. 7. Membrane flux with different concentration multiple of biogas slurry (DS: 2 M NaCl, FO mode, solution temperature: 20°C, flow rate 1.5 L/min).

TDS concentration of biogas slurry increased. While the TDS concentration was above 12 g/L, membrane flux rapidly decreases. There were two reasons for membrane flux reduction. Firstly, concentration gradient across the membrane gradually reduced with the TDS concentration of

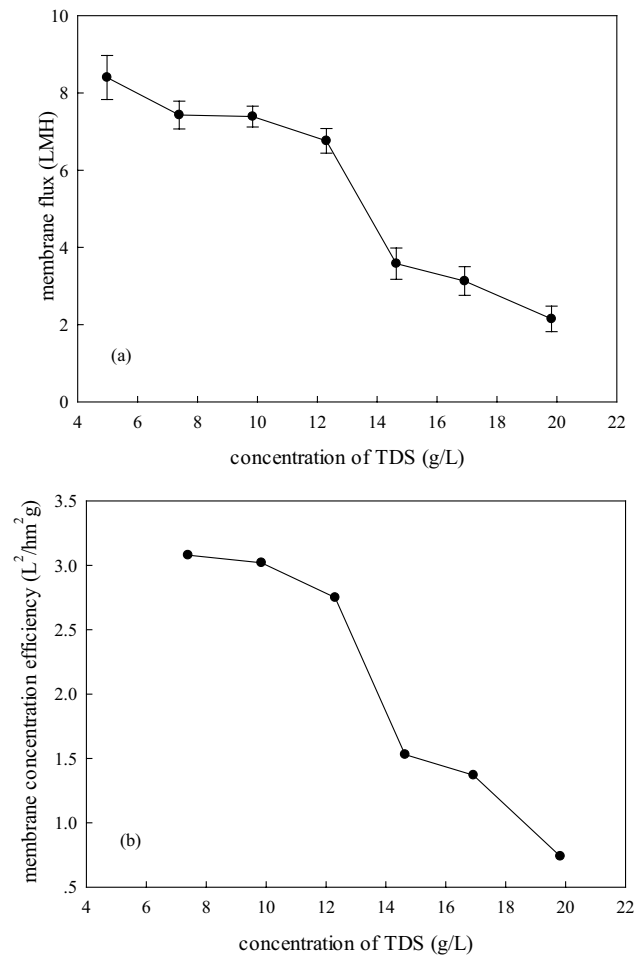


Fig. 8. Membrane concentration efficiency (DS: 2 M NaCl, FO mode, solution temperature: 20°C, flow rate 1.5 L/min).

Table 3
Concentration of main substances in biogas slurry and the recovery

Concentration multiple		1 (raw biogas slurry)	1.5	2	2.5	3	3.5	4
TDS	Concentration (g/L)	4.98 ± 0.51	7.39	9.84	12.3	14.46	19.92	19.82
	Recovery (%)	–	99.3	98.1	98.1	97.9	97.3	95.6
TK	Concentration (mg/L)	712 ± 31	1,053	1,413	1,766	2,129	2,434	2,823
	Recovery (%)	–	98.9	98.9	98.7	99.1	99.1	98.6
TP	Concentration (mg/L)	253 ± 23	373	502	630	745	870	993
	Recovery (%)	–	98.7	98.8	98.4	98.7	98.8	98.1
TN	Concentration (mg/L)	2,118 ± 40	3,157	4,221	5,275	6,301	7,360	8,419
	Recovery (%)	–	99.1	99.2	99.2	98.5	98.4	98.7
TOC	Concentration (mg/L)	1,501 ± 65	2,216	2,995	3,741	4,473	5,209	5,951
	Recovery (%)	–	99.1	99.4	99.1	98.5	98.7	98.4

TK - Total potassium; TN - Total nitrogen; TP - Total phosphorus

biogas slurry increased. As a result, the driving force of the membrane showed a little decline, resulting in a continuous decrease in membrane flux. The variation of membrane flux at different concentrations multiple of biogas slurry also exhibited the same trend. Secondly, membrane fouling gradually became serious, resulting in a sharp decrease in flux. In fact, the latter was more important.

By calculating the ratio of membrane flux and the variation of TDS at different stages, membrane concentration efficiency can be obtained as shown in Fig. 8b. With the increase of TDS concentration, membrane concentration efficiency displayed a decreasing trend. When the TDS concentration was less than 9.84 g/L, membrane concentration efficiency had little change. The change rate was just only 6%. The highest concentration efficiency 3.08 L²/h·m²·g was observed with the TDS concentration 7.39 g/L. As the TDS concentration increased, membrane concentration efficiency declined sharply. When the TDS concentration was 19.82 g/L, the minimum concentration efficiency was only 0.74 L²/h·m²·g. According to the results, it can be concluded that a high concentration efficiency was obtained while biogas slurry was concentrated between 1.5 and 2 times.

4. Conclusions

In this study, some influence factors were investigated in FO process for biogas slurry recovery. According to the results, conclusions can be drawn:

- Properties of DS concentration, membrane orientation, solution flow rate and temperature had significant influence on membrane performance. Water flux was higher in PRO mode than that of the FO mode. But more rapid flux decline was also observed in the PRO mode. Increased liquid flow rate was favor to the membrane process because of the increasing shear force, which alleviated the foulants deposition effectively. However, an excessive flow rate could damage membrane surface and shorten membrane life. Higher temperature results more rapid flux decline rate. Although operation at higher temperature may yield higher pure water flux. Through the investigation, the optimum operation conditions with DS concentration 2 M, flow rate 1.5 L/min, temperature 20°C and FO mode can be obtained. The optimum operating conditions should be changed based on the use case required, when the bench scales up.
- Under the optimum conditions, the maximum membrane flux was about 8.4 LMH, when the FS was raw biogas slurry. The recovery rates of TDS and other substances can reach over 95.6%, no matter how many times concentrated the biogas slurry was. The membrane flux decreased with the increased concentration multiple of biogas slurry. The maximum membrane concentration efficiency was about 3 L²/h·m²·g, when the concentration time of biogas slurry was between 1.5 and 2. FO technology was feasible to concentrate biogas slurry. It can improve the application value of biogas slurry as fertilizer effectively.

Acknowledgement

This research was supported by grants from the National Natural Science Foundation of China (Nos. 52200003) and Chongqing Municipal Education Commission (Nos. KJQN202201430).

References

- [1] T.Y. Gao, H.M. Zhang, X.T. Xu, J.H. Teng, Integrating microbial electrolysis cell based on electrochemical carbon dioxide reduction into anaerobic osmosis membrane reactor for biogas upgrading, *Water Res.*, 190 (2021) 116679, doi: 10.1016/j.watres.2020.116679.
- [2] C. Sun, Q.L. Yu, Z.Q. Zhao, Y.B. Zhang, Enhancing photo-synthetic CO₂ fixation in microbial electrolysis cell (MEC)-based anaerobic digestion for the *in-situ* biogas upgrading, *Chem. Eng. J.*, 462 (2023) 142341, doi: 10.1016/j.cej.2023.142341.
- [3] M. Sobhi, J.B. Guo, M.S. Gaballah, B.W. Li, J.B. Zheng, X. Cui, H. Sun, R.J. Dong, Selecting the optimal nutrients recovery application for a biogas slurry based on its characteristics and the local environmental conditions: a critical review, *Sci. Total Environ.*, 814 (2022) 152700, doi: 10.1016/j.scitotenv.2021.152700.
- [4] P.P. Wang, X. Zhang, S.G. Gouda, Q.X. Yuan, Humidification–dehumidification process used for the concentration and nutrient recovery of biogas slurry, *J. Cleaner Prod.*, 247 (2020) 119142, doi: 10.1016/j.jclepro.2019.119142.
- [5] M. Mondor, L. Masse, D. Ippersiel, F. Lamarche, D. Masse, Use of electro dialysis and reverse osmosis for the recovery and concentration of ammonia from swine manure, *Bioresour. Technol.*, 99 (2008) 7363–7368.
- [6] M. Mondor, D. Ippersiel, F. Lamarche, L. Masse, Fouling characterization of electro dialysis membranes used for the recovery and concentration of ammonia from swine manure, *Bioresour. Technol.*, 100 (2009) 566–571.
- [7] H. Luo, T. Lyu, A. Muhmood, Y. Xue, H. Wu, F. Meers, R. Dong, S. Wu, Effect of flocculation pre-treatment on membrane nutrient recovery of digested chicken slurry: mitigating suspended solids and retaining nutrients, *Chem. Eng. J.*, 352 (2018) 855–862.
- [8] M.R. Bilal, N.I. Mat Nawi, D.D. Subramaniam, N. Shamsuddin, A.L. Khan, J. Jaafar, A.B.D. Nandiyanto, Low-pressure submerged membrane filtration for potential reuse of detergent and water from laundry wastewater, *J. Water Process Eng.*, 369 (2020) 101264, doi: 10.1016/j.jwpe.2020.101264.
- [9] S. Hube, M. Eskafi, K.F. Hrafnkelsdottir, B. Bjarnadottir, M.A. Bjarnadottir, S. Axelsdottir, B. Wu, Direct membrane filtration for wastewater treatment and resource recovery: a review, *Sci. Total Environ.*, 710 (2020) 136375, doi: 10.1016/j.scitotenv.2019.136375.
- [10] J. Thuvander, A.-S. Jönsson, Techno-economic impact of air sparging prior to purification of alkaline extracted wheat bran hemicelluloses by membrane filtration, *Sep. Purif. Technol.*, 253 (2020) 117498, doi: 10.1016/j.seppur.2020.117498.
- [11] Z.Z. Zhou, L.H. Chen, Q.G. Wu, T. Zheng, H.R. Yuan, N. Peng, M.Y. He, The valorization of biogas slurry with a pilot dual stage reverse osmosis membrane process, *Chem. Eng. Res. Des.*, 142 (2019) 133–142.
- [12] H.N. Ruan, Z.R. Yang, J.Y. Lin, J.N. Shen, J.B. Ji, C.J. Gao, B.V. Bruggen, Biogas slurry concentration hybrid membrane process: pilot-testing and RO membrane cleaning, *Desalination*, 368 (2015) 171–180.
- [13] L.P. Gu, X. Tang, Y. Sun, H.J. Kou, Bioavailability of dissolved organic matter in biogas slurry enhanced by catalytic ozonation combined with membrane separation, *Ecotoxicol. Environ. Saf.*, 196 (2020) 110547, doi: 10.1016/j.ecoenv.2020.110547.
- [14] J.T. Martin, G. Kollopoulos, V.G. Papangelakis, An improved model for membrane characterization in forward osmosis, *J. Membr. Sci.*, 598 (2020) 117–126.

- [15] S. Lee, Y. Kim, J. Park, H.K. Shon, S. Hong, Treatment of medical radioactive liquid waste using forward osmosis (FO) membrane process, *J. Membr. Sci.*, 556 (2018) 238–247.
- [16] X. Wang, V.W.C. Chang, C.Y. Tang, Osmotic membrane bioreactor (OMBR) technology for wastewater treatment and reclamation: advances, challenges, and prospects for the future, *J. Membr. Sci.*, 504 (2016) 113–132.
- [17] C. Boo, M. Elimelech, S. Hong, Fouling control in a forward osmosis process integrating seawater desalination and wastewater reclamation, *J. Membr. Sci.*, 444 (2013) 148–156.
- [18] M. Qiu, C.J. He, Efficient removal of heavy metal ions by forward osmosis membrane with a polydopamine modified zeolitic imidazolate framework incorporated selective layer, *J. Hazard. Mater.*, 367 (2019) 339–347.
- [19] P. Mondal, A.T.K. Tran, B. Van der Bruggen, Removal of As(V) from simulated groundwater using forward osmosis: effect of competing and coexisting solutes, *Desalination*, 348 (2014) 33–38.
- [20] Y. Cui, Q. Ge, X.-Y. Liu, T.-S. Chung, Novel forward osmosis process to effectively remove heavy metal ions, *J. Membr. Sci.*, 467 (2014) 188–194.
- [21] Y. Dong, Z.W. Wang, C.W. Zhu, Q.Y. Wang, J.X. Tang, Z.C. Wu, A forward osmosis membrane system for the post-treatment of MBR-treated landfill leachate, *J. Membr. Sci.*, 471 (2014) 192–200.
- [22] S. Iskander, S.Q. Zou, B. Brazil, J.T. Novak, Z. He, Energy consumption by forward osmosis treatment of landfill leachate for water recovery, *Waste Manage.*, 63 (2017) 284–291.
- [23] M. Qin, H. Molitor, B. Brazil, J.T. Novak, Z. He, Recovery of nitrogen and water from landfill leachate by a microbial electrolysis cell-forward osmosis system, *Bioresour. Technol.*, 200 (2016) 485–492.
- [24] A.J. Ansari, F.I. Hai, W.E. Price, L.D. Nghiem, Phosphorus recovery from digested sludge centrate using seawater-driven forward osmosis, *Sep. Purif. Technol.*, 163 (2016) 1–7.
- [25] M.T. Vu, W.E. Price, T. He, X.W. Zhang, L.D. Nghiem, Seawater-driven forward osmosis for pre-concentrating nutrients in digested sludge centrate, *J. Environ. Manage.*, 247 (2019) 135–139.
- [26] J.L. Soler-Cabezas, J.A. Mendoza-Roca, M.C. Vincent-Vela, M.J. Luján-Facundo, L. Pastor-Alcañiz, Simultaneous concentration of nutrients from anaerobically digested sludge centrate and pre-treatment of industrial effluents by forward osmosis, *Sep. Purif. Technol.*, 193 (2018) 289–296.
- [27] D.R. Kashyap, K. Dadhich, S.K. Sharma, Biomethanation under psychrophilic conditions: a review, *Bioresour. Technol.*, 87 (2003) 147–153.
- [28] M. Xie, M. Price, I. Nghiem, M. Elimelech, Effects of feed and draw solution temperature and transmembrane temperature difference on the rejection of trace organic contaminants by forward osmosis, *J. Membr. Sci.*, 438 (2013) 57–64.
- [29] M.L. Xu, Q. Ye, Y.G. Li, Y.Q. Song, F. Xiao, Optimization of forward osmosis process for concentration of biogas slurry, *Trans. Chin. Soc. Agric. Eng.*, 32 (2016) 193–198.
- [30] H.N. Li, Z.W. Shi, C.X. Zhu, Concentration of biogas slurry with forward osmosis technology, *Trans. Chin. Soc. Agric. Eng.*, 24 (2014) 240–245.
- [31] Y. Li, X.M. Xie, R.X. Yin, Q.Z. Dong, Q.Q. Wei, B.X. Zhang, Effects of different draw solutions on biogas slurry concentration in forward osmosis membrane: performance and membrane fouling, *Membrane*, 12 (2022) 476–483.
- [32] X.H. Zhang, Q.G. Li, J. Wang, J. Li, C.W. Zhao, D.Y. Hou, Effects of feed solution pH and draw solution concentration on the performance of phenolic compounds removal in forward osmosis process, *J. Environ. Chem. Eng.*, 5 (2017) 2508–2514.
- [33] G. Blandin, H. Vervoort, P. Le-Clech, A.R.D. Verliefde, Fouling and cleaning of high permeability forward osmosis membranes, *J. Water Process Eng.*, 9 (2016) 161–169.
- [34] P. Zhao, Q.Y. Yue, B.Y. Gao, J.J. Kong, H.Y. Rong, P. Liu, H.K. Shon, Q. Li, Influence of different ion types and membrane orientations on the forward osmosis performance, *Desalination*, 344 (2014) 123–128.
- [35] J. Heo, K.H. Chu, N. Her, J. Im, Y.G. Park, J. Cho, S. Sarp, A. Jang, M. Jang, Y. Yoon, Organic fouling and reverse solute selectivity in forward osmosis: role of working temperature and inorganic draw solutions, *Desalination*, 389 (2016) 162–170.
- [36] P. Xiao, J. Li, Y.W. Ren, X. Wang, A comprehensive study of factors affecting fouling behavior in forward osmosis, *Colloids Surf., A*, 499 (2016) 163–172.
- [37] M.R. Chowdhury, J.R. McCutcheon, Elucidating the impact of temperature gradients across membranes during forward osmosis: coupling heat and mass transfer models for better prediction of real osmotic systems, *J. Membr. Sci.*, 553 (2018) 189–199.
- [38] S. Chintalacheruvu, Y. Ren, J. Maisonneuve, Effectively using heat to thermally enhance pressure retarded osmosis, *Desalination*, 556 (2023) 116570, doi: 10.1016/j.desal.2023.116570.
- [39] L. Feng, L. Xie, G. Suo, X. Shao, T. Dong, Influence of temperature on the performance of forward osmosis using ammonium bicarbonate as draw solute, *Trans. Tianjin Univ.*, 24 (2018) 571–579.
- [40] A.H. Hawari, N. Kama, A. Altaee, Combined influence of temperature and flow rate of feeds on the performance of forward osmosis, *Desalination*, 398 (2016) 98–105.
- [41] Y. Kim, S. Lee, H.K. Shon, S. Hong, Organic fouling mechanisms in forward osmosis membrane process under elevated feed and draw solution temperatures, *Desalination*, 355 (2015) 169–177.
- [42] M.I. Dova, K.B. Petrotos, H.N. Lazarides, On the direct osmotic concentration of liquid foods. Part I: impact of process parameters on process performance, *J. Food Eng.*, 78 (2007) 422–430.
- [43] N.T. Hancock, W.A. Phillip, M. Elimelech, T.Y. Cath, Bidirectional permeation of electrolytes in osmotically driven membrane processes, *Environ. Sci. Technol.*, 45 (2011) 10642–10651.